

1 **Quantifying spatiotemporal variability in zooplankton dynamics in the Gulf of Mexico with**
2 **a physical-biogeochemical model**

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16 **Abstract**

17 Zooplankton play an important role in global biogeochemistry, and their secondary production
18 supports valuable fisheries of the world's oceans. Currently, zooplankton standing stocks cannot
19 be estimated using remote sensing techniques. Hence, coupled physical-biogeochemical models
20 (PBMs) provide an important tool for studying zooplankton on regional and global scales.
21 However, evaluating the accuracy of zooplankton biomass estimates from PBMs has been a major
22 challenge due to sparse observations. In this study, we configure a PBM for the Gulf of Mexico
23 (GoM) from 1993-2012 and validate the model against an extensive combination of biomass and
24 rate measurements. Spatial variability in a multi-decadal database of mesozooplankton biomass
25 for the northern GoM is well resolved by the model with a statistically significant ($p < 0.01$)
26 correlation of 0.90. Mesozooplankton secondary production for the region averaged $66 \pm 8 \times 10^6$
27 kg C yr^{-1} , equivalent to $\sim 10\%$ of net primary production (NPP), and ranged from 51 to 82×10^6 kg
28 C yr^{-1} , with higher secondary production inside cyclonic eddies and substantially reduced
29 secondary production in anticyclonic eddies. Model results from the shelf regions suggest that
30 herbivory is the dominant feeding mode for small mesozooplankton ($<1\text{-mm}$) whereas larger
31 mesozooplankton are primarily carnivorous. In open-ocean oligotrophic waters, however, both
32 mesozooplankton groups show proportionally greater reliance on heterotrophic protists as a food
33 source. This highlights an important role of microbial and protistan food webs in sustaining
34 mesozooplankton biomass in the GoM, which serves as the primary food source for early life
35 stages of many commercially important fish species, including tuna.

36 **1. Introduction**

37 Within marine pelagic ecosystems, zooplankton function as an important energy pathway between
38 the base of the food chain and higher trophic levels such as fish, birds, and mammals (Landry et
39 al., 2019; Mitra et al., 2014). Zooplankton also have a well-documented impact on chemical
40 cycling in the ocean (Buitenhuis et al., 2006; Steinberg and Landry, 2017; Turner, 2015). The
41 ecological roles of zooplankton, however, are varied and taxon dependent. Globally, protistan
42 grazing is the largest source of phytoplankton mortality, accounting for 67% of daily
43 phytoplankton growth (Landry and Calbet, 2004). Protistan zooplankton function primarily within
44 the microbial loop leading to efficient nutrient regeneration in the surface ocean (Sherr and Sherr,
45 2002; Strom et al., 1997). By contrast, mesozooplankton contribute significantly less to
46 phytoplankton grazing pressure, consuming an estimated 12% of primary production globally
47 (Calbet, 2001), but strongly impact the biological carbon pump. In addition to grazing pressure on
48 phytoplankton, mesozooplankton affect the biological carbon pump through top-down pressure on
49 protistan grazers, production of sinking fecal pellets, consumption of sinking particles, and active
50 carbon transport during diel vertical migration (Steinberg and Landry, 2017; Turner, 2015).
51 Herbivorous mesozooplankton are particularly important to study as they are often associated with
52 shorter food chains that enable efficient energy transfer from primary producers to higher trophic
53 levels of immediate societal interest such as economically valuable fish species and/or their
54 planktonic larvae.

55 Zooplankton populations have been identified as being vulnerable to impacts of a warming ocean
56 (Caron and Hutchins, 2013; Pörtner and Farrell, 2008; Straile, 1997), through direct temperature
57 effects on metabolic rates (Ikeda et al., 2001; Kjellerup et al., 2012) and thermal stratification-
58 driven alterations in food web structure (Landry et al., 2019; Richardson, 2008). Studies aimed at
59 monitoring and predicting zooplankton populations are therefore critical for understanding the
60 first-order effects of a warming ocean on marine ecosystems given the importance of secondary
61 production and the impact zooplankton have on biogeochemical cycling. Despite their importance,
62 zooplankton have been historically sampled with limited temporal and spatial resolution. Unlike
63 ocean hydrodynamics and phytoplankton variability, zooplankton abundance cannot currently be
64 estimated remotely from space. Thus, numerical models provide a useful tool for synoptic
65 assessments of zooplankton stocks on basin and global scales (Buitenhuis et al., 2006; Sailley et
66 al., 2013; Werner et al., 2007). Nonetheless, evaluating the accuracy of zooplankton abundance

67 estimates in numerical experiments, such as three-dimensional physical-biogeochemical ocean
68 models (PBMs), is a major challenge due to the sparse ship-based observations in most regions
69 (Everett et al., 2017). Consequently, PBMs are typically predominately validated against surface
70 chlorophyll (Chl) from remote sensing products (Doney et al., 2009; Gregg et al., 2003; Xue et al.,
71 2013).

72 In most marine environments, phytoplankton net growth rates and biomass are determined
73 primarily by the imbalance between phytoplankton growth and zooplankton grazing (Landry et
74 al., 2009). PBMs can accurately predict phytoplankton standing stock (i.e. compare well with
75 satellite Chl observations) despite being driven by the wrong underlying dynamics, leading to
76 major errors in model estimates of secondary production and nutrient cycling (Anderson, 2005;
77 Franks, 2009). For instance, parameter tuning using only surface Chl as a validation metric can
78 allow broad patterns in phytoplankton biomass to be reproduced even with gross over- or
79 underestimation of phytoplankton turnover times. Similarly, even a model that is validated against
80 satellite Chl and net primary production might completely misrepresent the proportion of
81 phytoplankton mortality mediated by zooplankton groups, leading to inaccurate estimates of
82 important ecological metrics like secondary production and carbon export. Hence, validating
83 PBMs against zooplankton dynamics is key to increasing confidence in model solutions. The
84 importance of validation is further evident when considering zooplankton impacts on the behaviors
85 of biogeochemical models (Everett et al., 2017). Differences in simulated zooplankton
86 communities expressed through the number of functional types, various mathematical grazing
87 functions, or the arrangement of transfer linkages have been shown to have substantial impacts on
88 the dynamics of simple and complex biogeochemical models (Gentleman et al., 2003b; Gentleman
89 and Neuheimer, 2008; Mitra et al., 2014; Murray and Parslow, 1999; Sailley et al., 2013).

90 The Gulf of Mexico (GoM) is a particularly suitable region for examining zooplankton dynamics
91 with PBMs. In the northern and central Gulf, zooplankton abundances have been extensively
92 measured for over three decades (1982-present) by the Southeast Area Monitoring and Assessment
93 Program (SEAMAP). Within the SEAMAP dataset, zooplankton biomass exhibits strong
94 spatiotemporal variability, reflecting complex physical circulation in the GoM. Circulation off the
95 shelf is characterized by substantial upper layer mesoscale activity driven primarily by the
96 energetic Loop Current (Forristall et al., 1992; Maul and Vukovich, 1993; Oey et al., 2005). In

97 contrast, coastal and shelf circulation patterns are predominantly wind-driven (Morey et al., 2003a,
98 2013). Freshwater discharged by the Mississippi River and other smaller rivers is frequently
99 entrained offshore by shelf break interaction with mesoscale features (e.g., anti-cyclonic loop
100 current eddies), leading to strong horizontal and vertical gradients in physical and biogeochemical
101 quantities (Morey et al., 2003b). Overlap of these gradients with the SEAMAP study region result
102 in zooplankton collections across biogeochemically heterogeneous environments, providing a
103 powerful model constraint. For instance, Chl ranges across three orders-of-magnitude ($\sim 0.01 - 10$
104 mg Chl m^{-3}) from oligotrophic to eutrophic waters.

105 Over the past decade several PBM studies have been conducted in the GoM, all primarily
106 examining nutrient and phytoplankton dynamics. Early work by Fennel et al. (2011) examined
107 phytoplankton dynamics on the Louisiana and Texas continental shelf, concluding that loss terms
108 (e.g., grazing) rather than growth rates dictated accumulation rates of phytoplankton biomass. With
109 the same biogeochemical model, Xue et al. (2013) conducted the first gulf-wide PBM study to
110 investigate broad seasonal biogeochemical variability and to constrain a shelf nitrogen budget.
111 More recently, Gomez et al. (2018) implemented a biogeochemical model with multiple
112 phytoplankton and zooplankton functional types to gain a more detailed understanding of nutrient
113 limitation and phytoplankton dynamics in the GoM. To examine phytoplankton seasonality and
114 biogeography in the oligotrophic Gulf, Damien et al. (2018) validated a PBM based on a unique
115 subsurface autonomous glider dataset. Together, these studies have demonstrated the utility of
116 PBMs for investigating GoM lower trophic levels and have also highlighted the key ecosystem
117 roles of zooplankton. Specifically, both Fennel et al. (2011) and Gomez et al. (2018) identified the
118 importance of zooplankton in modulating the simulated seasonal patterns of phytoplankton
119 biomass, emphasizing the importance of top-down control on the shelf. Although simulated
120 zooplankton community results were not presented, Damien et al. (2018) noted that biotic
121 processes such as grazing pressure, are “essential to fully understanding the functioning of the
122 GoM ecosystem.” However, in all of these studies, zooplankton validation was largely absent.

123 In this study, we configured a PBM for the GoM to estimate zooplankton abundance and analyze
124 zooplankton community dynamics. The PBM is forced by three-dimensional hydrodynamic fields
125 from a data assimilative Hybrid Coordinate Ocean Model (HYCOM) hindcast of the GoM
126 (<http://www.hycom.org>). The PBM is based on the biogeochemical model NEMURO (North

127 Pacific Ecosystem Model for Understanding Regional Oceanography; Kishi et al., 2007), which is
128 substantially modified here for application to the GoM. The model is integrated over 20-years
129 (1993-2012) and validated against an extensive combination of remote and in situ measurements
130 including total and size-fractioned mesozooplankton biomass and grazing rates, microzooplankton
131 grazing rates, phytoplankton growth rates and net primary production as well as validation of
132 surface chlorophyll and vertical profiles of chlorophyll and nitrate. Our goals were: 1) to develop
133 and validate a PBM to estimate mesozooplankton abundance, 2) to characterize the spatiotemporal
134 variability in mesozooplankton dietary composition, and 3) to quantify regional mesozooplankton
135 secondary production. We focus primarily on the oligotrophic open-ocean GoM, where prey (i.e.
136 zooplankton) availability may be limiting for fish, their larvae and other higher trophic levels.

137 **2 Methods and data**

138 **2.1 Biogeochemical model configuration**

139 **2.1.1 NEMURO model description**

140 The biogeochemical model for this study is based on NEMURO (Kishi et al., 2007) but has been
141 modified and parameterized to more accurately reflect the ecology of the GoM (herein called
142 NEMURO-GoM). NEMURO is a concentration-based, lower-trophic-level ecosystem model
143 originally developed and parameterized for the North Pacific. Like most marine functional-group
144 biogeochemical models, it is structured around simplified representations of the lower food web
145 originating from earlier nutrient-phytoplankton-zooplankton models (Fasham et al., 1990; Franks,
146 2002; Riley, 1946). Complexity is added through additional state variables and transfer functions
147 with the specific goal of resolving dynamics within the nutrient, phytoplankton and zooplankton
148 pools. In total, NEMURO has eleven state variables: six non-living state variables – nitrate (NO_3),
149 ammonium (NH_4), dissolved organic nitrogen (DON), particulate organic nitrogen (PON), silicic
150 acid ($\text{Si}(\text{OH})_4$) and particulate silica (Opal); two phytoplankton state variables – small (SP) and
151 large phytoplankton (LP); and three zooplankton state variables – small (SZ), large (LZ) and
152 predatory zooplankton (PZ).

153 Each biological state variable in NEMURO is an aggregated representation of taxonomically
154 diverse plankton groups that function similarly in the ecosystem. The phytoplankton community
155 is modeled as two functional types of obligate autotrophs: small phytoplankton (SP, predominantly
156 cyanobacteria and picoeukaryotes in the GoM) and large phytoplankton (LP, diatoms). Small

157 zooplankton (SZ) represent heterotrophic protists, and metazoan zooplankton are divided into
158 suspension-feeding mesozooplankton (LZ) and predatory zooplankton (PZ). Here, we assume that
159 LZ and PZ are non-migratory. Heterotrophic bacteria are implicitly represented by temperature-
160 dependent decomposition rates, which represent nitrification and remineralization processes.
161 NEMURO uses nitrogen as a model “currency” since it is the major limiting macronutrient in
162 much of the ocean. Silica is also included as a potentially co-limiting nutrient for diatoms (i.e. LP).

163 By default, sinking in NEMURO is restricted to PON and Opal, and benthic processes are not
164 included. Here, because of the large shelf area in the GoM, we implemented a simple diagenesis
165 of PON/Opal to NO_3/SiO_4 and removal of PON/Opal through sedimentation, where 1% of the flux
166 sinking out of bottom cell was removed and 10% converted back into NO_3/SiO_4 . However, we
167 found that this had no significant impact on the simulated surface Chl or mesozooplankton biomass
168 on the shelf. The inclusion of a more complex sediment diagenesis model (including
169 denitrification) would have added further realism (Fennel et al., 2011). However, our main focus
170 was to evaluate zooplankton dynamics in the oligotrophic region where higher trophic levels that
171 depend on mesozooplankton secondary production may experience food limitation and where
172 benthic processes are negligible.

173 NEMURO was chosen for the present study because it distinguishes SZ, LZ and PZ, permitting a
174 detailed analysis of dynamics for multiple functional types in the GoM zooplankton community.
175 During initial GoM simulations, default NEMURO parameterizations, configured for the North
176 Pacific (Kishi et al., 2007), substantially overestimated surface nitrate, surface Chl, and
177 mesozooplankton biomass relative to observations. We attribute these differences to: 1)
178 substantially higher temperatures in the GoM compared with the North Pacific, which significantly
179 increase decomposition and growth rates in the model resulting in higher nutrient recycling and
180 elevated near-surface stocks of phytoplankton and zooplankton, and 2) distinct differences in
181 taxonomic composition of phytoplankton and zooplankton communities in the GoM and North
182 Pacific, with significant differences in key parameter values associated with growth and grazing.

183 For more details on the specific processes represented and the interactions between state variables
184 in NEMURO, we direct readers to Kishi et al. (2007). All model equations are provided in the
185 supplement to this manuscript. Biogeochemical model forcing, initial and open boundary

186 conditions are also outlined in Supplement S1. Briefly, daily average shortwave radiation fields
187 obtained from Climate Forecast System Reanalysis (CFSR) were used to force light limitation of
188 phytoplankton. Once a final parameter set was determined (see section 2.1.3), initial and open
189 boundary conditions for all state variables were prescribed from a spun up idealized one-
190 dimensional version of NEMURO-GoM. After initializing, the three-dimensional model was spun
191 up over four years before conducting the full 20-year experiment. River nutrient input from the
192 Mississippi was prescribed using nitrate samples collected by United States Geological Survey
193 (USGS) and due to a lack of observations for other rivers was prescribed for all 37 rivers
194 represented in the model.

195 **2.1.2 Modifications to default NEMURO model**

196 To improve realism for application to the GoM, five structural changes were made to the original
197 NEMURO model. First, we removed the SP to LZ grazing pathway. The original SP state variable
198 for the North Pacific represents nanophytoplankton (e.g. coccolithophores), which can be
199 important prey of copepods and other mesozooplankton. In the GoM, however, cyanobacteria and
200 picoeukaryotes (too small for direct feeding by most mesozooplankton) comprise much of the
201 phytoplankton biomass and hence are represented as SP in our model. In addition to adding
202 ecological realism, this change in direct trophic connection between SP and LZ allowed the model
203 to produce a more realistic LP-dominated phytoplankton community on the shelf (see Discussion).

204 Next, quadratic mortality was replaced with linear mortality for all biological state variables with
205 the exception of predatory zooplankton (PZ). In biogeochemical models, quadratic mortality is
206 often used for numerical stability and/or to represent implicit loss terms to an un-modeled parasite
207 or predator that covaries in abundance with its prey (e.g. viral lysis of phytoplankton or predation
208 by un-modeled higher predators) (Anderson et al., 2015). However, grazing mortality is explicitly
209 modeled in NEMURO and viral mortality is generally not a substantial loss term for bulk
210 phytoplankton (Staniewski and Short, 2018). Quadratic mortality was retained for PZ, to account
211 for predation pressure of un-modeled predators (e.g. planktivorous fish). During the model tuning
212 process, we found that removal of quadratic mortality from the four other plankton functional
213 groups was an important parameterization change that allowed the model to simulate more realistic
214 mesozooplankton biomass in the oligotrophic GoM (see Discussion).

215 The default ammonium inhibition term and light limitation functional form in NEMURO were
216 replaced in NEMURO-GoM with more widely adopted parameterizations. The exponential
217 ammonium inhibition term in the nitrate limitation function was replaced with the term described
218 by Parker (1993), as has been done in previous PBM studies (Fennel et al., 2006) due to the non-
219 monotonic behavior of the default NEMURO ammonium inhibition term. At high NO_3
220 concentrations, the default term is known to generate unrealistic phytoplankton nutrient uptake
221 patterns in which total nutrient uptake (i.e. uptake of NO_3 + uptake of NH_4) can actually decrease
222 despite increases in NH_4 (and constant NO_3).

223 Light limitation in NEMURO is based on an optimal light parameterization that implicitly includes
224 photoinhibition. This formulation was replaced with the Platt et al. (1980) functional form that
225 allows one to explicitly control the amount of photoinhibition, which can be important in the GoM
226 where surface irradiances are high. Additionally, the Platt functional form is commonly used and
227 thus parameter values are easier to find for comparison (e.g. initial slope of the PI curve (α)). This
228 formulation is also implemented in newer versions of NEMURO, such as the code used in the
229 Regional Ocean Modeling System (ROMS) NEMURO biogeochemical package.

230 Finally, to account for photoacclimation and more accurately simulate Deep Chlorophyll
231 Maximum (DCM) dynamics, we replaced the constant C:Chl parameter with a variable C:Chl
232 model where ratios for SP and LP were allowed to vary based on the formulation described by Li
233 et al. (2010), which considers both light and nutrient limitation (see Supplemental). The Li et al.
234 (2010) equations build on a previously constructed dynamic regulatory model of phytoplankton
235 physiology which describes C:Chl variability under balanced growth and nutrient saturated
236 conditions at constant temperature (see Geider et al., 1998)). Herein, “default” NEMURO includes
237 the modified ammonium inhibition, light formulation, and variable C:Chl model.

238 **2.1.3 NEMURO-GoM model tuning procedure**

239 In total, NEMURO includes 71 parameters, 23 of which were modified in the present study. For
240 initial model tuning, we used an idealized one-dimensional model designed to mimic the
241 oligotrophic GoM. To guide our tuning procedure, we relied on a semi-quantitative approach
242 where the one-dimensional model solution was evaluated based on five ecosystem benchmarks.
243 Target values for benchmarks and other ecosystem attributes were determined from observations

244 or a theoretical basis. Ecosystem benchmarks included: surface Chl, mesozooplankton biomass,
245 DCM depth, DCM magnitude, and SP:LP ratio. Surface Chl and mesozooplankton biomass were
246 chosen as benchmarks to evaluate the realism of plankton biomass in the model. The DCM depth
247 and magnitude were chosen to evaluate the vertical structure of the simulated ecosystem, and
248 SP:LP ratio was used to gauge the realism of the plankton community composition (i.e. high SP:LP
249 is expected in the oligotrophic GoM). The model was also tuned by considering the relative
250 magnitudes of loss terms for phytoplankton (grazing, mortality, respiration, and excretion), total
251 protistan zooplankton grazing relative to mesozooplankton grazing, as well as surface and deep
252 nitrate concentrations. We outline each parameter change, justification and the resulting impact on
253 the ecosystem benchmarks simulated by the idealized one-dimensional model in Supplement S3.
254 Where possible, we modified parameters in groups so that relative changes were consistent
255 throughout the model (e.g. doubling all zooplankton mortality terms). After tuning in the one-
256 dimensional model, parameter sets were implemented into the full three-dimensional model where
257 additional tuning was performed. Once a final parameter set was determined we conducted a
258 parameter sensitivity analysis over 18 individual experiments to identify impacts of parameter
259 changes from default NEMURO values (S4).

260 **2.2 Physical model configuration**

261 **2.2.1 Description of the offline numerical environment**

262 To run large numbers of three-dimensional simulations efficiently for basin-scale tuning,
263 NEMURO-GoM was run offline using the MITgcm offline tracer advection package. MITgcm
264 was selected as it contains convenient packages for running offline simulations (McKinley et al.,
265 2004). That is, the dynamical equations of motion are not computed during the NEMURO-GoM
266 integration, but rather the physical prognostic variables (i.e., temperature, salinity and three-
267 dimensional velocity fields) are prescribed from daily-averaged flow fields saved from a previous
268 hydrodynamic model integration. This allows the recycled use of flow fields leaving only the tracer
269 equations to be computed. In the offline MITgcm package, the prognostic variables provide input
270 to an advection scheme and mixing routine that conservatively handles offline advection and
271 diffusion of the biogeochemical tracer fields. MITgcm has many options for linear and non-linear
272 advection schemes. Here we use a 3rd order direct space time flux limiting scheme. Sub grid-scale
273 mixing of the biogeochemical fields is handled offline through the nonlocal K-Profile
274 Parameterization (KPP) package based on mixing schemes developed by Large et al. (1994). For

275 more information about the MITgcm packages, we direct readers to the MITgcm manual
276 (<http://mitgcm.org/>).

277 There are two main advantages to running PBMs in an offline environment: 1) the momentum
278 equations are not integrated during the model run; and 2) the physical time step is no longer bound
279 by the dynamical Courant–Friedrichs–Lewy (CFL) numerical stability criterion, which together
280 significantly reduces the computational cost. Instead, the stability of the tracer advection scheme
281 and time scales needed to resolve biological/physical processes of interest set the limits on the time
282 steps and prescription frequencies of flow fields. When the physical time step is shorter than the
283 flow field prescription frequency, a simple linear interpolation of the flow fields is performed
284 between time steps. Offline simulations of tracer advection have been found to closely resemble
285 online runs (that is, computed together with the integration of the hydrodynamic model’s
286 prognostic equations) when the three-dimensional flow fields are prescribed at a frequency that is
287 at or below the inertial period ($T = 2\pi/f$, $T_{GoM} > 24 \text{ hr}$) for a region (Hill et al., 2005).

288 In the present study, the offline time step (30 minutes) is an order of magnitude greater than the
289 hydrodynamic model’s (HYCOM-GoM, described in Section 2.2.2) baroclinic time step (120
290 seconds). For reference, HYCOM-GoM required ~76 days to run to completion on 64 parallel
291 cores. These time requirements would increase considerably with the 11 additional
292 biogeochemical tracers used in NEMURO. In contrast, NEMURO-GoM ran significantly faster,
293 taking a total of ~50 hours on 80 parallel cores. Offline models offer a valuable tool for integrating
294 PBMs particularly as spatial resolution and complexity in these models continues to increase (e.g.,
295 DARWIN (Follows et al., 2007), GENOME (Coles et al., 2017)). While computationally
296 advantageous, however, offline simulations have inherently greater input and output (I/O)
297 demands that can become bottlenecks in some applications. Issues with conservation can also arise
298 as three-dimensional advection schemes are only approximately positive definite.

299 **2.2.2 Description of the offline dynamical fields**

300 The NEMURO-GoM model is “forced” by daily averaged three-dimensional velocity, temperature
301 and salinity fields from a pre-existing 20-year (1993-2012) HYCOM (HYbrid Coordinate Ocean
302 Model) (Chassignet et al., 2003) regional GoM hindcast (H-GoM). H-GoM is based on version
303 2.2.99B of the HYCOM code, originally provided by the Naval Oceanographic Office

304 (NAVOCEANO) Major Shared Resource Center. H-GoM was run at 1/25th degree (~4 km)
305 horizontal resolution with 36 vertical hybrid coordinate layers and assimilated historic, in situ and
306 satellite observations. The domain encompasses the entire GoM and extends south of the Mexican-
307 Cuba Yucatan channel to 18 °N and as far east as 77 °W (**Fig. 1**). Further details on H-GoM
308 (experiment ID: GOMu0.04/expt_50.1), including model forcing and the main model
309 configuration file (i.e. blkdat.input_501), can be found at <https://www.hycom.org>.

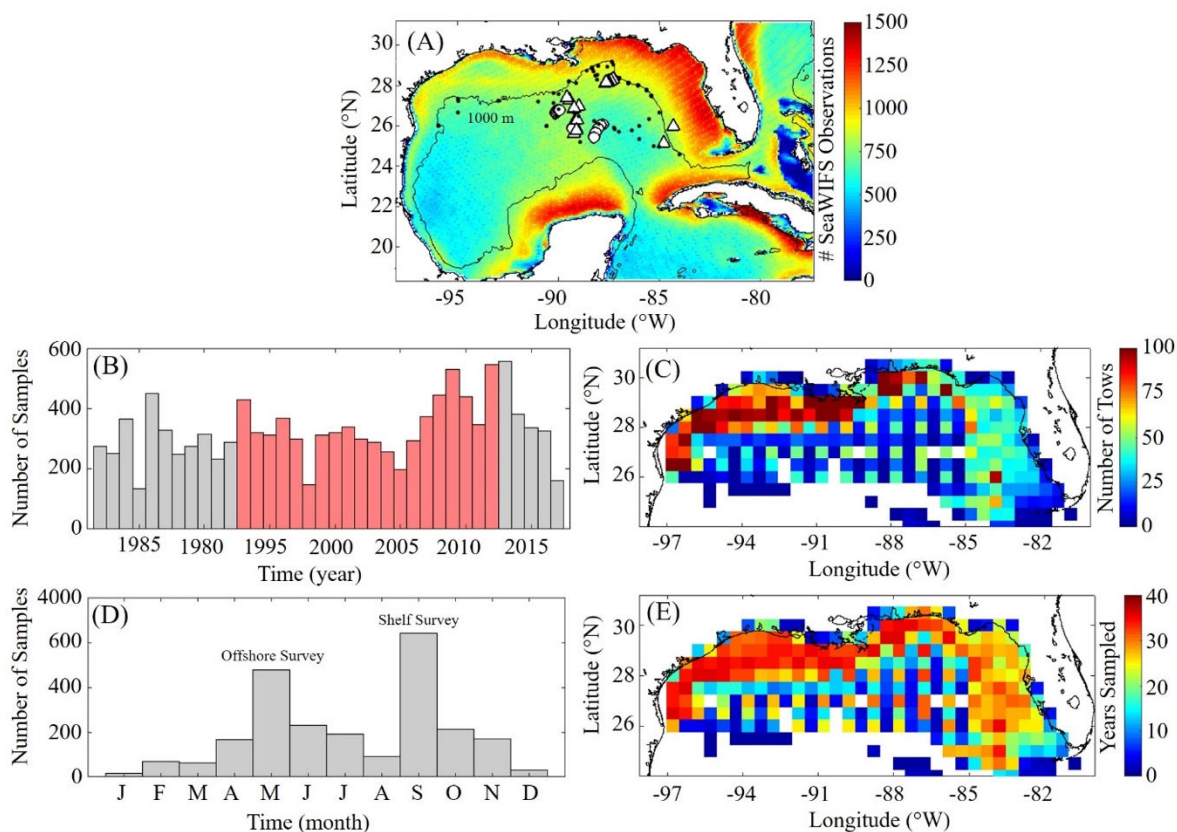
310 The H-GoM flow fields were mapped from the HYCOM native hybrid vertical coordinate to z-
311 levels used by the MITgcm. NEMURO-GoM was configured for 29 vertical z-levels (10-m
312 intervals from 0-150 m, 25-m intervals from 150-300 m, 50-m intervals from 300-500m, and 1000
313 m, 2000 m, ~4000 m). Mapping was performed by computing total zonal and meridional
314 transports across the lateral boundaries of each MITgcm grid cell (e.g., 0-10 m bin; which may
315 include multiple HYCOM layers) and then dividing by the area of the respective cell face. This
316 vertical mapping approach is consistent as both HYCOM and MITgcm use an Arakawa C-grid
317 orientation for model variables. The H-GoM bathymetry was adjusted such that no partial cells
318 existed in the domain to avoid thin cells. The continuity equation was subsequently used to
319 calculate vertical velocities. The use of transports in this approach ensures conservation and
320 approximately identical profiles of vertical velocity to those in H-GoM fields. For mapping of
321 temperature and salinity fields (used in the KPP mixing routine and for scaling biological
322 temperature dependent rates), a simple linear interpolation was performed.

323 **2.3 Model validation**

324 **2.3.1 Surface chlorophyll observations**

325 A benchmark for surface Chl was determined using the Sea-Viewing Wide Field-of-View Sensor
326 (SeaWIFS) product from the Ocean Biology Processing Group (OBPG) of the National
327 Aeronautics and Space Administration (NASA). The product used here is the mapped, level-3,
328 daily, 9-km resolution images from 4 September 1997 to 10 December 2010 processed according
329 to the algorithm of Hu et al. (2012). To compute model-data point-to-point comparisons, we take
330 the corresponding daily-averaged simulated surface Chl field and interpolate to the SeaWIFS grid
331 before applying the daily cloud coverage mask corresponding to the matching SeaWIFS image. In
332 total 4,291 daily images consisting of 22,244,513 non-zero cell values (herein referred to SeaWIFS

333 measurements) were used to validate NEMURO-GoM. Approximately 500-1200 daily model-data
 334 point-to-point comparisons were made for each SeaWIFS grid cell (**Fig 1**).



335
 336 **Figure 1 (A-E):** Spatial and temporal coverage of all observational data sets used for model
 337 validation. Total number of non-zero SeaWIFS gridded values from the level 3 product from 4
 338 September 1997 to 10 December, 2010 along with cruise sample locations collected during May,
 339 2017 (circles) and 2018 (triangles) and nitrate profiles from the World Ocean Database (dots) (A).
 340 Total annual sampling of the SEAMAP surveys from 1983-2017 (B) with samples overlapping
 341 with the PBM simulation period denoted in red. Total sample density within each $0.5^\circ \times 0.5^\circ$ box
 342 (C). Total seasonal sampling (D). Number of years with at least one sample (E). 1000 m isobaths
 343 and coastline are denoted by black continuous lines.

344 2.3.2 Mesozooplankton biomass observations

345 To evaluate model mesozooplankton biomass estimates, we used plankton tows collected during
 346 SEAMAP surveys in the northern and central GoM. In total, 11,781 plankton tows were collected
 347 from 1983-2017, with two main annual surveys in the spring (offshore) and fall (shelf) (**Fig. 1**).

348 On average, SEAMAP collected approximately 300 samples per year with a specific sampling
349 array offshore and more general sampling coverage on the shelf. In total, 6,835 samples were used
350 for direct point-to-point model-data comparisons. Samples were collected using standard gear
351 consisting of a 61-cm diameter bongo frame fitted with two 333- μm mesh nets. The nets were
352 fished in a double-oblique tow pattern from the surface down to 200 m or 5 m off the bottom and
353 back to the surface. Simultaneous samples were also collected using a 202- μm mesh net during 82
354 tows. Of these samples, roughly half were collected in the oligotrophic GoM. The average ratio
355 between biomass measured in the 333- and 202- μm bongo tows (0.5093 ± 0.12) was used to
356 convert 333- μm samples so that direct comparisons could be made with model mesozooplankton
357 (LZ+PZ) biomass fields.

358 In NEMURO-GoM, the small zooplankton (SZ) state variable represents early stages of
359 mesozooplankton and heterotrophic protists (e.g. ciliates), which are typically $< 200 \mu\text{m}$ in the
360 ocean. The large zooplankton (LZ) state variable represents small suspension-feeding
361 mesozooplankton (e.g. small to medium sized copepods), which were assumed to range in size
362 from 0.2 to 1.0 mm. Predatory zooplankton (PZ) are considered to be large mesozooplankton (e.g.
363 large copepods) ranging in size from 1.0 to 5.0 mm. Mesozooplankton size classes were defined
364 to allow comparisons to be made with field measurements (see section 2.3.4). Zooplankton
365 biomass in net tows was originally quantified as displacement volumes (DV). Carbon mass (CM)
366 equivalents were subsequently calculated as $\log_{10}(\text{CM}) = (\log_{10}(\text{DV}) + 1.434)/0.820$ (Wiebe, 1988;
367 Moriarty and O'Brien, 2013). For comparison to the SEAMAP climatology the model
368 mesozooplankton fields were similarly depth averaged to the bottom or 200 m and converted to
369 units of carbon assuming Redfield C:N ratio. For point-to-point model-data comparisons,
370 simulated mesozooplankton biomass fields were interpolated to SEAMAP sample locations/times
371 before being depth averaged to the corresponding sample tow depth.

372 **2.3.3 Observed vertical profiles of chlorophyll and nitrate**

373 Depth profiles of Chl were also collected during SEAMAP surveys using a SeaBird WETStar
374 fluorometer attached to a CTD. Calibration of the fluorimeter was infrequent, and thus profiles
375 were used to determine the depth of the fluorescence maxima for comparisons to DCM depths in
376 the model. In total, 2,435 profiles were collected from 2003 to 2012, with 1,052 profiles overlying

377 bottom depths >1000 m. Profiles were available for earlier SEAMAP surveys; however, no
378 standard QA/QC protocol for fluorometer data was in place prior to 2003.

379 To evaluate DCM magnitudes in the model, we used 145 fluorescence profiles collected during
380 May 2017 and 2018 process study cruises (see section 2.3.4). The fluorometer was attached to a
381 CTD and calibrated using 126 in situ Chl samples. Chl concentrations were determined from
382 filtered samples collected at depths ranging from 5 to 115 m using High Performance Liquid
383 Chromatography (HPLC). Since the cruise sampling does not overlap with our NEMURO-GoM
384 simulation period, model-data comparisons were made for all 20 years of the model run using
385 sample locations and time of the year. This was also done with other field measurements from the
386 process cruises (see section 2.3.4). For model-data comparisons of nitrate, we utilized profiles
387 from the World Ocean Database (WOD). In total, 96 profiles were available during our simulation
388 period and located in the oligotrophic GoM (>1000 m isobath). Profiles were collected during all
389 months except March and December with the majority of samples collected during May, July and
390 August (**Fig. 1A**).

391 **2.3.4 Biomass and rate measurements from process study cruises**

392 Although in situ rate measurements are made much less frequently than biological standing stock
393 measurements, they offer very powerful constraints for validating the internal dynamics of a
394 biogeochemical model (Franks, 2009). Consequently, we made phytoplankton and zooplankton
395 rate measurements on two cruises in the open ocean GoM in May 2017 and 2018 and used these
396 measurements to validate the model (**Fig. 1A**). On the process study cruises, we utilized a quasi-
397 Lagrangian sampling scheme to investigate plankton dynamics in the oligotrophic GoM. Two
398 drifting arrays (one sediment trap array and one in situ incubation array) were deployed to serve
399 as a moving frame of reference during ~4-day studies (“cycles”) characterizing the water parcel
400 (Landry et al., 2009; Stukel et al., 2015). During these cycles, we measured daily profiles of Chl,
401 photosynthetically active radiation, phytoplankton growth rates and productivity, protistan grazing
402 rates, and size-fractionated mesozooplankton biomass and grazing rates.

403 Size-fractionated mesozooplankton biomass and grazing rates were determined from daily day-
404 night paired oblique ring-net tows (1-m diameter, 202- μ m mesh). In total, 40 oblique bongo net
405 tows (16 in 2017 and 24 in 2018) sampled the oligotrophic GoM mesozooplankton community

406 from near surface to a depth ranging from 100-135 m. Upon recovery, the sample was anesthetized
407 using carbonated water, split using a Folsom splitter, filtered through a series of nested sieves (5,
408 2, 1, 0.5, and 0.2 mm), filtered onto pre-weighed 200- μm Nitex filters, rinsed with isotonic
409 ammonium formate to remove sea salt, and flash frozen in liquid nitrogen. In the lab, defrosted
410 samples were weighed for total wet weight, and subsampled in duplicate (wet weight removed)
411 for gut fluorescence analyses. The remaining wet sample was dried and subsequently reweighed
412 and combusted for CHN analyses to determine total dry weight and C and N biomasses. Gut
413 fluorescence subsamples were homogenized using a sonicating tip, extracted in acetone, and
414 measured for Chl and phaeopigments using the acidification method. The phaeopigment
415 concentrations in the zooplankton guts were the basis for calculated grazing rates using gut
416 turnover times based on temperature relationships for mixed zooplankton assemblages. For
417 additional details, see Décima et al. (2011) and Décima et al. (2016).

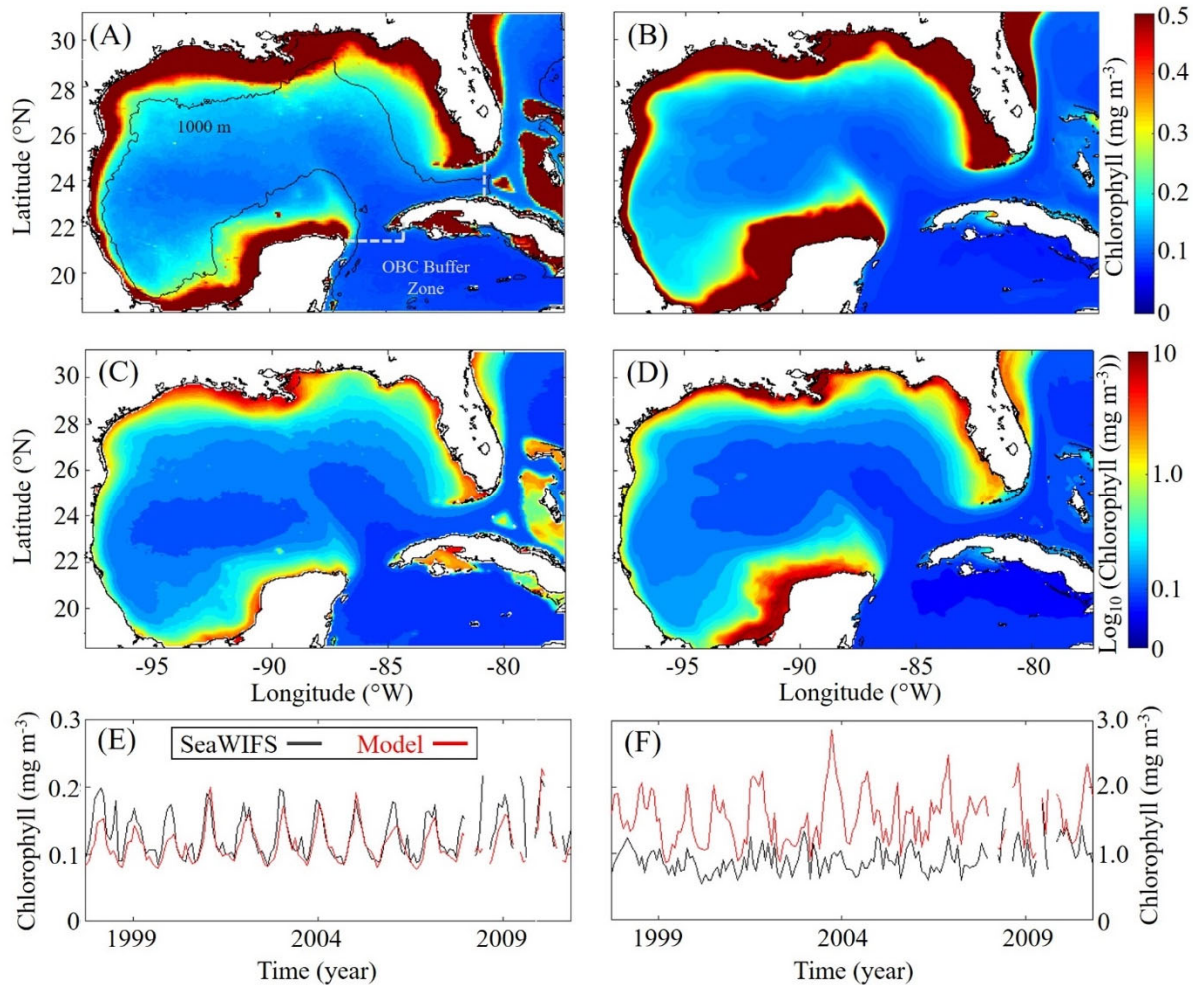
418 Protistan grazing rates were measured using the two-point, “mini-dilution” variant of the
419 microzooplankton grazing dilution method (Landry et al., 1984, 2008; Landry and Hassett, 1982).
420 Briefly, one 2.8-L polycarbonate bottle was gently filled with whole seawater taken from six
421 depths (from the surface to the depth of the mixed layer). A second 2.8-L bottle was then filled
422 with 33% whole seawater and 67% 0.2- μm filtered seawater. Both bottles were then placed in
423 mesh bags and incubated in situ at natural depths for 24 h. These experiments were conducted on
424 each day of the ~4-day cycle. After 24 h, the bottles were retrieved, filtered onto glass fiber filters,
425 and Chl concentrations were determined using the acidification method (Strickland and Parsons.,
426 1972). Net growth rates ($k = \ln(\text{Chl}_{\text{final}}/\text{Chl}_{\text{init}})$) in each bottle were determined relative to initial Chl
427 samples. Phytoplankton specific mortality rates resulting from the grazing pressure of protists were
428 calculated as $m = (k_d - k_0)/(1-0.33)$, where k_d is the growth rate in the dilute bottle and k_0 is the
429 growth rate in the control bottle. Phytoplankton specific growth rates were calculated as $\mu = k_0 +$
430 m . For additional details, see Landry et al. (2016) and Selph et al. (2016). Phytoplankton net
431 primary production was quantified at the same depths by $\text{H}^{13}\text{CO}_3^-$ uptake experiments. Triplicate
432 2.8-L polycarbonate bottles and a fourth “dark” bottle were spiked with $\text{H}^{13}\text{CO}_3^-$ and incubated in
433 situ for 24 h at the same sampling depths as for the dilution experiments. Samples were then filtered
434 and the $^{13}\text{C}:^{12}\text{C}$ ratios of particulate matter determined by isotope ratio mass spectrometry.

435 **3.0 Results**

436 3.1 Surface chlorophyll model-data comparisons

437 Model surface Chl estimates were found to agree closely with satellite observations reproducing
438 patterns in both the oligotrophic and shelf region (**Fig. 2**). Spatial covariance between SeaWIFS
439 climatology and model surface Chl climatology (calculated with daily cloud cover mask applied)
440 is statistically significant ($p < 0.01$) with a correlation (ρ) of 0.72. When model estimates are
441 compared to all 22,244,513 SeaWIFS measurements at corresponding times and locations (i.e.
442 daily grid cell pairs), we find a ρ value of 0.50 ($p < 0.01$). To facilitate more detailed model-data
443 comparisons, the GoM domain was divided into an oligotrophic region (≥ 1000 m bottom depth)
444 and a shelf region (< 1000 m bottom depth). In the oligotrophic region, the correlation between
445 model-data daily grid cell pairs is significant but weak ($\rho = 0.17$, $p < 0.01$) as a result of relatively
446 low large-scale spatial variability, and hence dominance at the mesoscale. However, bias is quite
447 low (-0.014 mg Chl m^{-3}), equivalent to 10% of the observed mean. In the shelf region, the
448 correlation is higher ($\rho = 0.47$, $p < 0.01$) yet the bias is greater ($+0.90$ mg Chl m^{-3}), equivalent to
449 92% of the mean. Previous GoM studies have determined ρ values for monthly averages, which
450 we calculate here for comparison. Based on 30-day averages, the ρ values are 0.70 ($p < 0.01$) for
451 the oligotrophic region and 0.26 ($p < 0.01$) for the shelf region.

452 In addition to resolving the dominant spatiotemporal variability, the model also captures the
453 amplitude of the seasonal surface Chl signal reasonably well. In the oligotrophic region, the model
454 accurately estimates the observed annual surface Chl minimum (Model: 0.065 ± 0.005 vs.
455 SeaWIFS: 0.065 ± 0.007 mg Chl m^{-3}) while slightly underestimating the observed annual
456 maximum (Model: 0.47 ± 0.15 vs. SeaWIFS: 0.75 ± 0.55 mg Chl m^{-3}). When model estimates for
457 the entire oligotrophic region are taken into account (i.e. not restricted to satellite measurement
458 locations and times), the annual minimum develops in early September, and the maximum
459 develops in late January (**Table 1**). In the shelf region, greater model-data mismatch exists for
460 surface Chl, with the model overestimating the observed annual minimum by 15% (Model: $0.23 \pm$
461 0.09 vs. SeaWIFS: 0.20 ± 0.07 mg Chl m^{-3}) and the observed annual maximum by 102% (Model:
462 8.09 ± 1.31 vs. SeaWIFS: 4.01 ± 1.23 mg Chl m^{-3}). Here, we find the annual surface Chl seasonal
463 cycle almost completely out of phase with the oligotrophic region, with the annual minimum
464 developing in early February and the annual maximum developing at the end of July (**Table 1**).



465

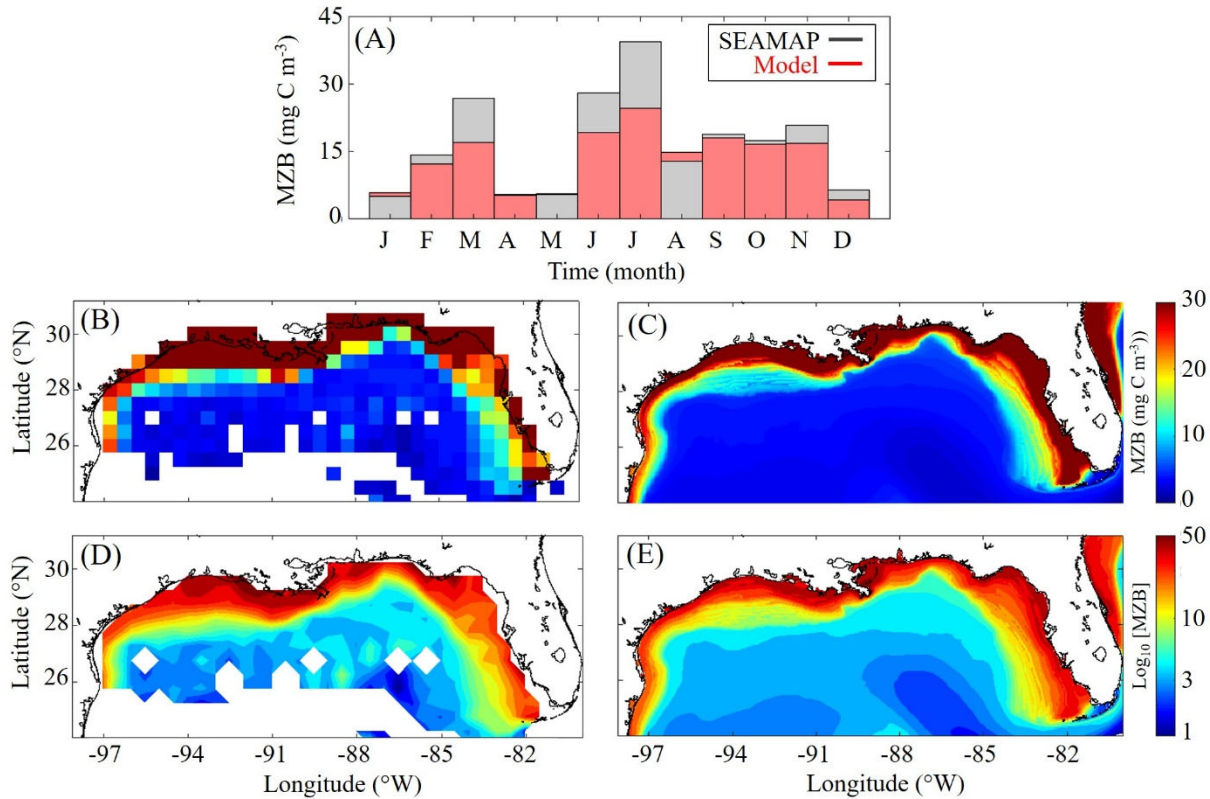
466 **Figure 2 (A-F):** Comparison of surface chlorophyll (mg m^{-3}) between SeaWIFS observations and
 467 model from 4 September 1997 to 10 December 2010. Average SeaWIFS chlorophyll (A). Average
 468 model estimated surface chlorophyll (B). Log_{10} of the average SeaWIFS chlorophyll (C). Log_{10} of
 469 the average model estimated surface chlorophyll (D). Time series of simulated 30-day average
 470 surface chlorophyll (red) and SeaWIFS observations (black) for bottom depths ≥ 1000 m (E) and
 471 bottom depths < 1000 m (F). The 1000 m isobaths and coastline are denoted by black lines.

472 3.2 Regional mesozooplankton biomass model-data comparisons

473 Model mesozooplankton biomass (i.e. LZ + PZ) fields also agree closely with observations in both
 474 the oligotrophic and shelf region (**Fig. 3**). Spatial covariance between SEAMAP climatology and
 475 model climatology of depth-averaged mesozooplankton biomass is statistically significant ($p <$
 476 0.01) with a ρ value of 0.90. When model estimates are compared to SEAMAP tows at

477 corresponding sample times and locations for the 6,835 measurements in the simulation period,
478 the ρ value is 0.55 ($p < 0.01$). In the oligotrophic region, the model slightly overestimates
479 mesozooplankton biomass (Model: $4.09 \pm 1.82 \text{ mg C m}^{-3}$ vs. SEAMAP: $3.52 \pm 3.44 \text{ mg C m}^{-3}$)
480 with ρ value of 0.23 ($p < 0.01$) with a bias of 0.57 mg C m^{-3} , equivalent to 16% of the observed
481 mean. Conversely, in the shelf region the model underestimates mesozooplankton biomass
482 (Model: $17.40 \pm 13.58 \text{ mg C m}^{-3}$ vs. SEAMAP: $20.91 \pm 24.62 \text{ mg C m}^{-3}$), with a ρ value of 0.49
483 ($p < 0.01$) and a bias of -3.5 mg C m^{-3} , equivalent to 17% of the observed mean. Model estimates
484 and SEAMAP measurements also compare well with total mesozooplankton biomass
485 measurements (0.2-5 mm) collected in the oligotrophic region during the process study cruises
486 (Model: $5.55 \pm 2.87 \text{ mg C m}^{-3}$ vs. Cruise: $4.33 \pm 2.28 \text{ mg C m}^{-3}$).

487 Although seasonal cycles in the oligotrophic and shelf regions could not be derived from the
488 SEAMAP dataset given the significant differences in sampling locations over the course of a year,
489 we investigated model-data mismatches for each month. The model closely matches or slightly
490 underestimates mesozooplankton biomass for most of the year, with the exception of January, May
491 and August (**Fig. 3A**). The largest model-data mismatch occurs during March, June, July and
492 December, when the model underestimates mesozooplankton biomass by approximately 35%.
493 Unlike surface Chl, the total mesozooplankton biomass (i.e. depth-integrated) seasonality is
494 similar in both regions of the GoM. In the oligotrophic region, the annual biomass minimum
495 (maximum) occurs at the beginning of January (middle of May), while in the shelf region, the
496 annual minimum (maximum) occurs in late December (end of May) (**Table 1**).



497

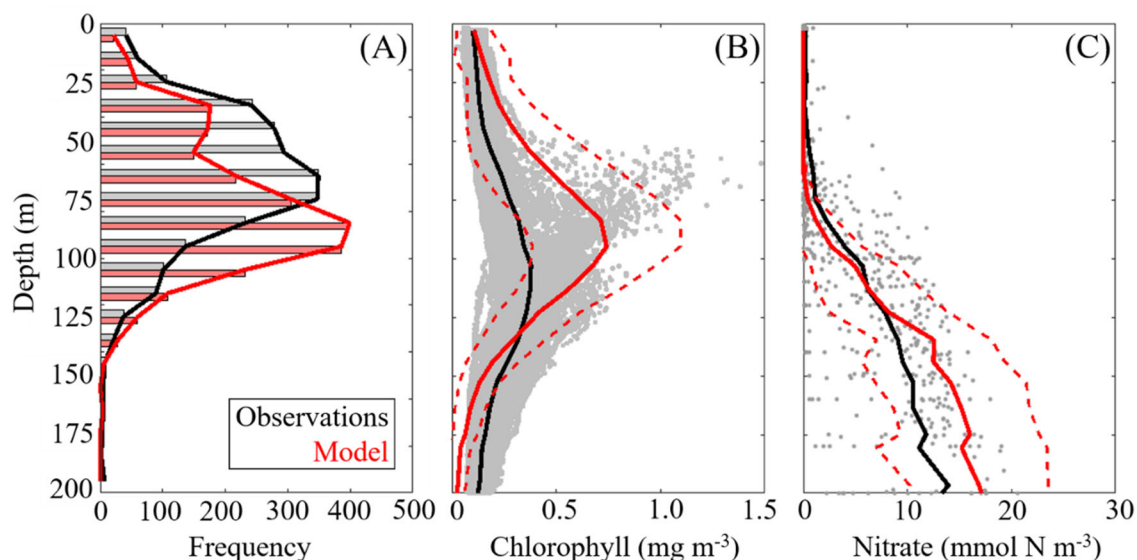
498 **Figure 3 (A-E):** Comparison of climatological depth-averaged (200 m) mesozooplankton biomass
 499 (MZB, mg C m⁻³) between SEAMAP observations (left) and model output (right). Monthly
 500 average MZB samples organized by month (A). Monthly variability is not representative of
 501 seasonality as sampling locations change between months. MZB from all SEAMAP tows (B).
 502 MZB 20-year model average (C). Log₁₀ of SEAMAP MZB (D). Log₁₀ of model MZB (E).

503 3.3 Chlorophyll and nitrate profile model-data comparisons

504 To validate the vertical structure of the simulated ecosystem, we utilized observed profiles of
 505 fluorescence, Chl and nitrate. When simulated DCM depths were compared to all 2,435 SEAMAP
 506 fluorescence profiles, we find a statistically significant correlation ($\rho = 0.59$, $p < 0.01$) with the
 507 observed maximum fluorescence depth. The maximum fluorescence depth ranged from the surface
 508 to 143 m while model values show a similar variability ranging from the surface to 163 m (**Fig.**
 509 **4A**). In the oligotrophic region, the model overestimates the DCM depth (Model: 95 ± 20 m vs.
 510 SEAMAP: 80 ± 25 m) and has a ρ value of 0.38 ($p < 0.01$) with a bias of 15 m, equivalent to 19%
 511 of the observed mean. In the shelf region, the model also overestimates DCM depth (Model: $63 \pm$

512 26 m vs. SEAMAP: 53 ± 23 m) and has a ρ value of 0.49 ($p < 0.01$) with a bias of 10 m, equivalent
513 to 19% of the observed mean.

514 In contrast, the model slightly underestimated the DCM depth when compared to calibrated
515 fluorescence profiles collected during the process cruises (Model: 100 ± 18 m vs. Observed: $107 \pm$
516 21 m) (**Fig. 4B**). In terms of magnitude, the model overestimates DCM Chl (Model: 0.74 ± 0.35
517 mg Chl m^{-3} vs. Observed: $0.38 \pm 0.13 \text{ mg Chl m}^{-3}$), although most of the observations fall within
518 one standard deviation of the model average. Despite this model-data mismatch, simulated nitrate
519 profiles closely match profiles from the World Ocean Database (WOD). In both model and
520 observations, the mean nitracline occurred at approximately 75 m (**Fig. 4C**). On average, model
521 nitrate tended to be lower at the surface and higher at depth relative to observations. Above the
522 nitracline, model nitrate was $0.071 \pm 0.39 \text{ mmol N m}^{-3}$ while observed nitrate was 0.55 ± 1.29
523 mmol N m^{-3} . Below 200 m, model and data show better agreement, with deep nitrate in the model
524 of $24.92 \pm 3.28 \text{ mmol N m}^{-3}$ compared to $23.55 \pm 5.21 \text{ mmol N m}^{-3}$ in WOD profiles.

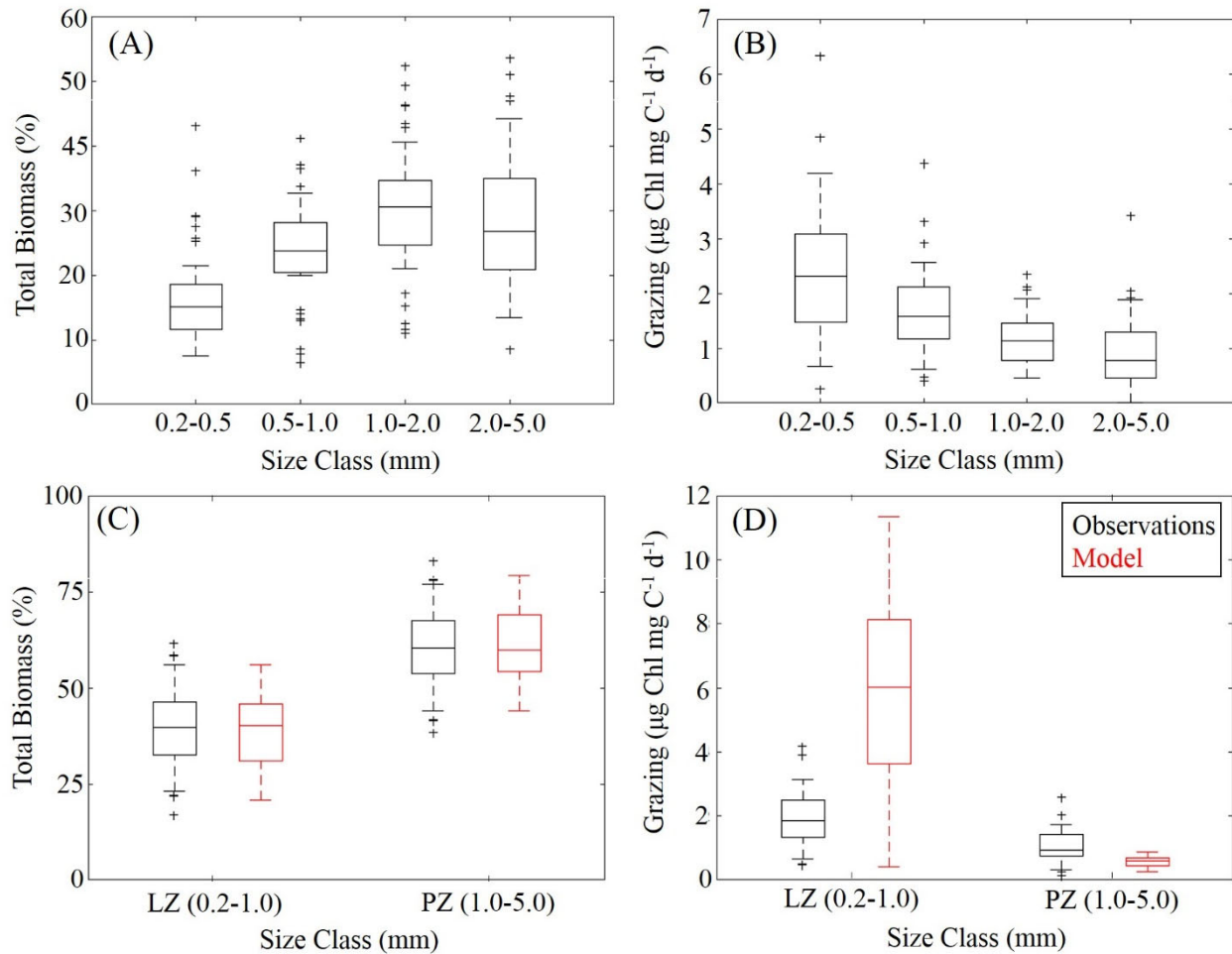


525
526 **Figure 4 (A-C):** Model-data comparisons of DCM depth (A) chlorophyll profiles (B) and nitrate
527 profiles (C). DCM depth was evaluated using un-calibrated fluorescence profiles obtained during
528 SEAMAP cruises. Chlorophyll profiles were collected during the May 2017 and 2018 Lagrangian
529 process cruises. For comparisons, the model and data were sampled at corresponding locations and
530 time of the year for all simulated years. Nitrate values from World Ocean Database that overlapped

531 with the simulation period and were located in the oligotrophic GoM (>1000 m) were used for
532 model-data comparisons.

533 **3.4 Size fractionated mesozooplankton biomass and grazing model-data comparisons**

534 To further constrain the phytoplankton and zooplankton community simulated by NEMURO-
535 GoM, we utilized in situ measurements collected during the process study cruises. First, we
536 compared the relative proportions of LZ and PZ biomass to four discrete size classes measured at
537 sea (**Fig. 5A, C**). In both model and observations, we find nearly identical size distributions
538 assuming that LZ approximates the smallest two size classes of mesozooplankton sampled (“small
539 mesozooplankton”, 0.2-1.0-mm) and PZ approximates the largest two size classes (“large
540 mesozooplankton”, 1.0-5.0 mm). In the field data, small mesozooplankton biomass varied from
541 33 to 46 % (median = 40%, at 95% C.I.), while model estimates of LZ biomass vary from 31 to
542 46% (median = 40%). Large mesozooplankton biomass in the field data varied from 54 to 67%
543 (median = 60%), while model estimates of PZ biomass vary from 54 to 69% (median = 60%).



544

545 **Figure 4 (A-D):** A summary of field (black) and model (red) estimates of mesozooplankton size-
 546 fractioned biomass and grazing rates. Mesozooplankton size-fractioned biomass as a percent of
 547 total biomass for each of the four size classes measured at sea in May, 2017 and 2018 (A).
 548 Corresponding mesozooplankton specific grazing rates for each of the four size classes (B). Field
 549 data aggregated into two size classes for direct comparison with model biomass estimates for large
 550 (LZ) and predatory (PZ) mesozooplankton (C). Similarly, model data comparison of specific
 551 grazing rates by large and predatory zooplankton to aggregated field estimates (D). Whiskers
 552 extend to 95% confidence interval. Outliers for model estimates are not shown.

553 Mesozooplankton specific grazing rates measured during the process study cruises were also used
 554 to validate the simulated mesozooplankton community. Field measurements showed that specific
 555 grazing rates ($\mu\text{g Chl mg C}^{-1} \text{d}^{-1}$), decreased consistently with increasing mesozooplankton size-
 556 class (**Fig. 5B**). For model-data comparisons, we computed grazing on LP by LZ and PZ at each

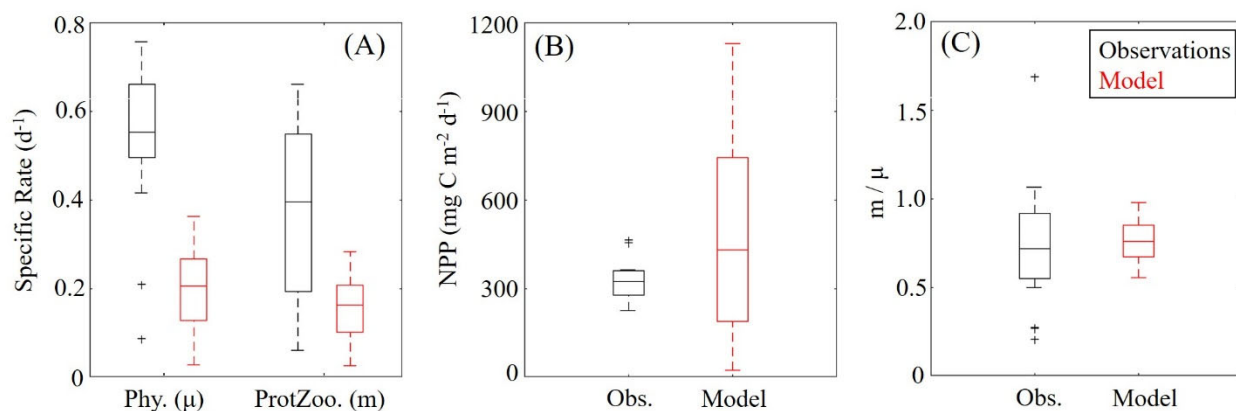
557 depth. Grazing terms were converted into units of Chl using the model estimated C:Chl ratio for
558 LP before being depth-integrated to the corresponding net tow depth and normalized to simulated
559 depth-integrated LZ and PZ biomasses. We find that model mesozooplankton grazing estimates
560 capture the general trend of decreased specific grazing rates with increasing mesozooplankton size
561 (**Fig. 5D**). However, the model overestimates grazing by small mesozooplankton while
562 underestimating grazing by large mesozooplankton. In the field data, small mesozooplankton
563 grazing ranges from 1.34 to 2.51 $\mu\text{g Chl mg C}^{-1} \text{d}^{-1}$ (median = 1.85) while model estimates of LZ
564 grazing rates vary from 3.64 to 8.14 $\mu\text{g Chl mg C}^{-1} \text{d}^{-1}$ (median = 6.01). Field measurements of
565 large mesozooplankton grazing range from 0.76 to 1.44 $\mu\text{g Chl mg C}^{-1} \text{d}^{-1}$ (median = 0.94), while
566 model estimates of PZ grazing vary from 0.44 to 0.70 $\mu\text{g Chl mg C}^{-1} \text{d}^{-1}$ (median = 0.58). In terms
567 of total grazing, the model average is considerably higher ($2.99 \pm 2.20 \mu\text{g Chl mg C}^{-1} \text{d}^{-1}$) than
568 found in the field measurements ($1.38 \pm 0.59 \mu\text{g Chl mg C}^{-1} \text{d}^{-1}$) (see Discussion).

569 **3.5 Phytoplankton growth and microzooplankton grazing model-data comparisons**

570 Measurements of specific phytoplankton growth rates, phytoplankton mortality due to
571 microzooplankton grazing, and net primary production (NPP) were used to evaluate protistan
572 dynamics in the model. We find the model underestimates phytoplankton growth and
573 microzooplankton grazing while overestimating NPP (**Fig. 6A, B**). Phytoplankton specific growth
574 rates from dilution experiments range from 0.50 to 0.66 d^{-1} (median = 0.55 d^{-1}) while model
575 estimates of phytoplankton (SP+LP) specific growth rates vary from 0.13 to 0.27 d^{-1} (median =
576 0.21 d^{-1}). In terms of microzooplankton grazing rates, field data range from 0.19 to 0.55 d^{-1} (median
577 = 0.39 d^{-1}) while model estimates of SZ vary from 0.10 to 0.21 d^{-1} (median = 0.16 d^{-1}). NPP
578 estimates show better agreement between model and data, with rates from 276 to 360 $\text{mg C m}^{-2} \text{d}^{-1}$
579 1 (median = 321 $\text{mg C m}^{-2} \text{d}^{-1}$) in field data while model estimates vary from 190 to 741 mg C m^{-2}
580 d^{-1} (median = 431 $\text{mg C m}^{-2} \text{d}^{-1}$).

581 Although the model underestimates phytoplankton growth and microzooplankton grazing rates,
582 the relative proportion of NPP consumed by protists in the model (67 - 85%; median = 76%)
583 compares reasonably well to field measurements (55 - 92%; median = 72%) (**Fig. 6C**). Notably,
584 the model average proportion of phytoplankton production consumed by protists closely matches
585 the mean for all tropical waters reported by Calbet & Landry (2004). When phytoplankton
586 mortality due to mesozooplankton grazing was evaluated in the model at cruise sample locations

587 we find mesozooplankton grazing accounts for $13 \pm 8 \%$, which also closely agrees with the global
 588 average (Calbet et al., 2001).



589
 590 **Figure 5 (A-C):** Specific phytoplankton growth (μ , d^{-1}) and microzooplankton grazing (m , d^{-1})
 591 between model (red) and field data (black) (A). Depth-integrated net primary production ($mg\ C$
 592 $m^{-2}\ d^{-1}$) (B). The fraction of phytoplankton growth that is grazed by protists in the model and field
 593 data (C). Whiskers extend to the 95% confidence intervals. Outliers for model estimates are not
 594 shown.

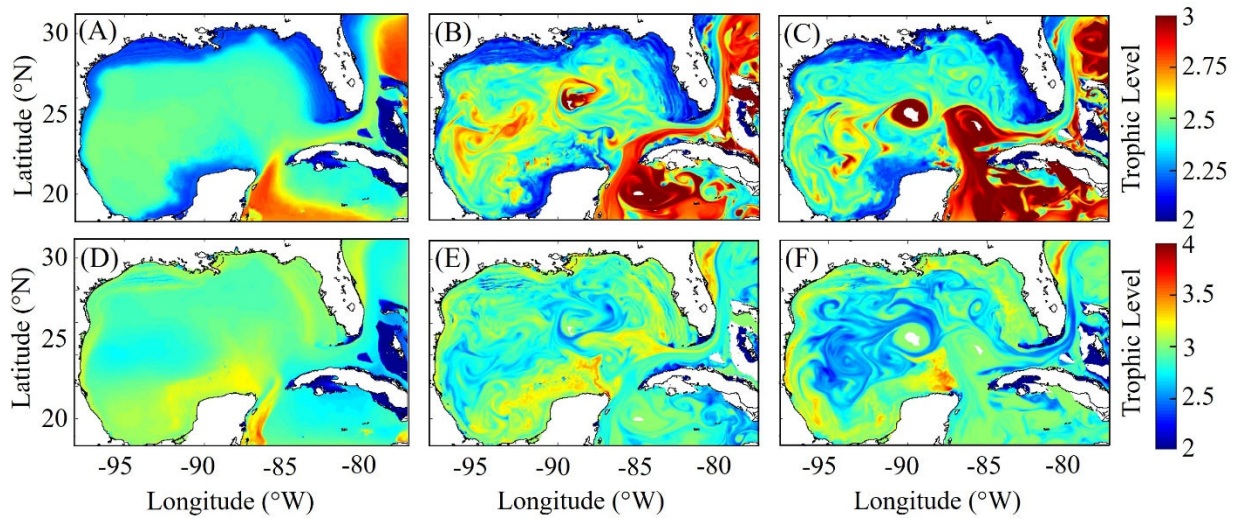
595 3.6 Simulated mesozooplankton diet

596 After model tuning and validation, we utilized NEMURO-GoM to investigate spatiotemporal
 597 variability in diet and secondary production of the GoM mesozooplankton community. First, we
 598 examined the trophic level of LZ and PZ in the model, which provides a measure of their
 599 cumulative diet. Trophic level is calculated by computing the dietary contributions of each prey in
 600 LZ (i.e. LP and SZ) and PZ diets (i.e. LP, SZ, and LZ), assuming that the trophic level of LP = 1
 601 and SZ = 2. In the oligotrophic region, both LP and SZ contribute approximately 50% to LZ diet,
 602 as indicated by a mean trophic level near 2.5 (2.54 ± 0.02) for LZ (**Fig. 7A**). In the same region,
 603 PZ have a trophic level of 2.78 ± 0.04 indicating a higher contribution of zooplankton to their diet
 604 (i.e. SZ and/or LZ) (**Fig. 7B**). In the shelf region, LZ are more herbivorous, as indicated by a
 605 decrease in trophic level to 2.31 ± 0.01 , while PZ are more carnivorous, as indicated by an increase
 606 in trophic level to 2.90 ± 0.04 .

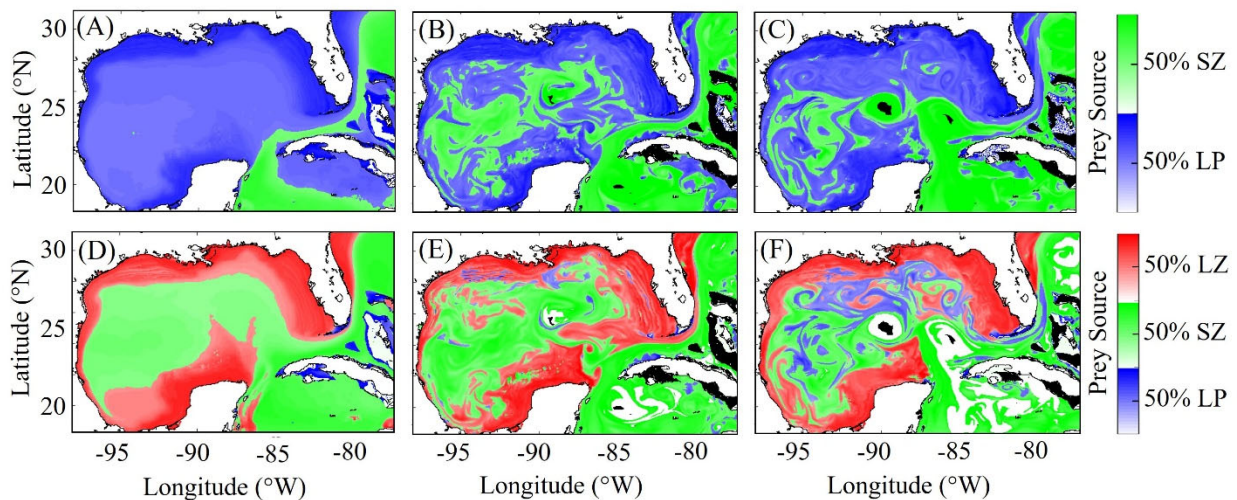
607 Despite little evidence for LZ diets dominated by zooplankton in the annual average (in contrast
 608 to PZ, which often have a trophic level ~ 3), we commonly find regions in instantaneous fields

609 during both winter and summer conditions where SZ are the dominant prey source for LZ (**Fig.**
610 **7C, E**). These regions, typically in the Loop Current or Loop Current Eddies (LCEs), highlight the
611 episodic importance of heterotrophic protists as prey sources for small mesozooplankton in the
612 GoM. High proportions of SZ in LZ diets can be attributed to the competitive advantage of SP
613 over LP in extremely low nutrient environments such as in the Loop Current, resulting in high
614 abundances of SP and their predators (SZ) relative to LP. Instantaneous fields also reveal that
615 phytoplankton can be important prey for PZ as well, particularly during summer, as indicated by
616 trophic levels of around 2.5 in the western GoM (**Fig. 7F**). In addition to strong variability in
617 trophic positions, there are also regions in the oligotrophic GoM, most clearly in the centers of
618 LCEs during summer, where the model predicts no feeding by mesozooplankton (**Fig. 8E**). The
619 convergent anti-cyclonic circulation of LCEs is typically associated with low phytoplankton
620 biomass, which at times may fall near or below feeding thresholds in the NEMURO grazing
621 formulation. This formulation is intended to simulate suppression of feeding activity for
622 zooplankton at mean prey densities that cannot support the energy expended while searching for
623 prey.

624 To investigate which prey contributes most to LZ and PZ diets, we computed each prey source
625 term for both LZ and PZ at each grid cell (**Fig. 8**). As we would expect, the dominant prey for LZ
626 and PZ align closely with spatial variability in their respective trophic positions. For LZ diet,
627 herbivory dominates throughout the GoM, except for the Loop Current (**Fig. 8A**). LP contribution
628 to LZ diet is highest on the shelf, where LP biomass is also high due to the competitive advantage
629 of LP over SP in high nutrient conditions. In contrast, PZ diet varies with the relative availability
630 of SZ and LZ prey. In the oligotrophic region, PZ feed mainly on SZ (heterotrophic protists)
631 because LZ biomass is relatively low. On the shelf, they consume primarily LZ (**Fig. 8D**). Despite
632 the significant change in dominant prey between the shelf and oligotrophic regions, PZ trophic
633 positions remain fairly consistent (**Fig. 7D**) because SZ in the oligotrophic region and LZ in the
634 shelf region both feed predominantly on phytoplankton and hence occupy similar trophic levels.
635 In the instantaneous fields for winter (**Fig. 8B, E**) and summer (**Fig. 8C, F**), the dominant prey for
636 both LZ and PZ show substantial mesoscale variability, indicating that oceanographic features
637 such as fronts and eddies influence not only biomass but also zooplankton ecological roles (see
638 Discussion).



639
 640 **Figure 7 (A-F):** Trophic levels of simulated large zooplankton (LZ, top) and predatory
 641 zooplankton (PZ, bottom). Annual-average trophic positions of LZ (A) and PZ (D). Instantaneous
 642 trophic positions of LZ (B) and PZ (E) for winter conditions on 4 February 2012. Instantaneous
 643 trophic positions of LZ (C) and PZ (F) for summer conditions on 5 August 2011.



644
 645 **Figure 8 (A-F):** Dominant prey source for simulated large zooplankton (LZ, top) and predatory
 646 zooplankton (PZ, bottom). Colors indicate dominant prey. Brightness indicates percent of
 647 dominant prey in the zooplankton diet. Annual averaged field for LZ (A) and PZ (D). Instantaneous
 648 winter condition for LZ (B) and PZ (E) on simulated day 4 February 2012. Instantaneous summer
 649 conditions for LZ (C) and PZ (F) on 4 August 2011.

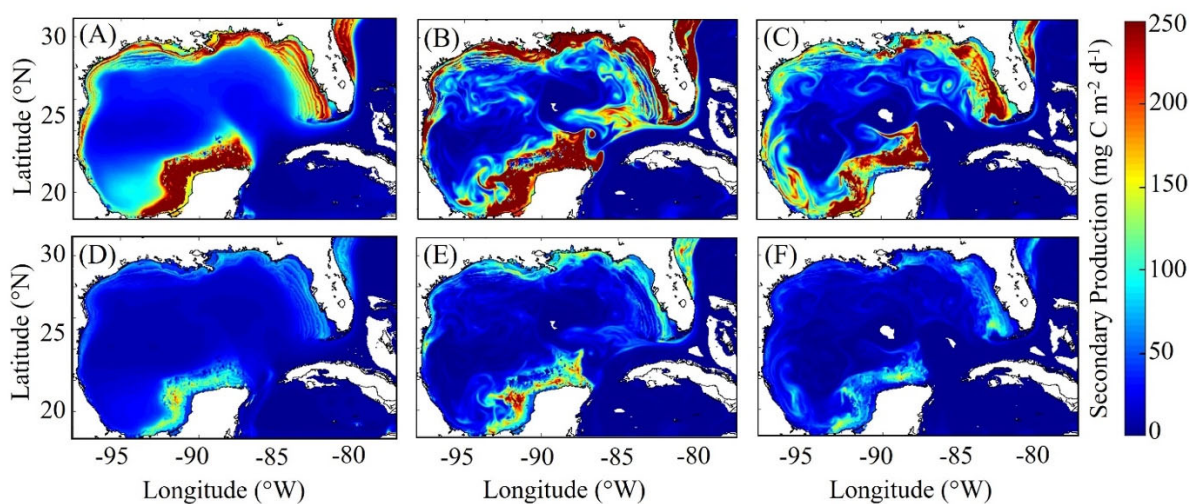
650 3.7 Simulated mesozooplankton secondary production

651 To our knowledge, regional secondary production for the GoM has not been quantified previously.
652 Based on our model, secondary production due to mesozooplankton averages $66 \pm 8 \times 10^6$ kg C yr⁻¹,
653 and ranged from a minimum of 51×10^6 kg C (in 1999) to a maximum of 82×10^6 kg C (in
654 2011). In the oligotrophic region, LZ secondary production averages 35 ± 5 mg C m⁻² d⁻¹ while
655 PZ secondary production is 11 ± 2 mg C m⁻² d⁻¹ (**Fig. 9**). The annual secondary production
656 minimum develops at the end of December while the annual maximum occurs at the beginning of
657 June (**Table 1**). In this region, mesozooplankton are responsible for $14 \pm 2 \times 10^6$ kg C yr⁻¹,
658 equivalent to 6% of NPP. On the shelf, secondary production is about 4-fold higher, with LZ
659 production of 146 ± 17 mg C m⁻² d⁻¹ and PZ production of 42 ± 5 mg C m⁻² d⁻¹. Here, the annual
660 minimum also occurs at the end of December but the maximum occurs later at the end of July
661 (**Table 1**). On the shelf, secondary production constitutes a higher proportion of NPP (13%) and
662 averages $51 \pm 6 \times 10^6$ kg C yr⁻¹.

663 In addition to differences in total secondary production, significant differences were found in the
664 mesozooplankton community response to changes in total phytoplankton biomass on the shelf and
665 in the oligotrophic region. On the shelf, the average ratio between LZ and PZ secondary production
666 is 3.51 and remains almost constant with increasing phytoplankton biomass ($\rho = 0.13$, $p < 0.01$).
667 Although we find a similar average value in the oligotrophic region (3.14), ratios are more variable
668 and strongly dependent on phytoplankton biomass ($\rho = 0.52$, $p < 0.01$). Ratios of LZ to PZ
669 secondary production reached values of ~ 2.5 in the lowest phytoplankton biomass regions of the
670 open ocean GoM and increased to ~ 4.0 during times and places where local phytoplankton biomass
671 was high. These differences likely stem from the longer turnover times of PZ, which make them
672 less sensitive to variability in bottom-up drivers and allows them to have a proportionally greater
673 role in oligotrophic settings.

674 As witnessed in the instantaneous fields of diet and secondary production, mesoscale eddies are
675 common features in the GoM and hence important to quantify for regional zooplankton dynamics. To
676 investigate secondary production inside cyclonic and anticyclonic eddies we
677 implement the TOEddies eddy detection algorithm (Laxenaire et al., 2018) which uses surface
678 velocities along closed contours of sea surface height (SSH) for detection of mesoscale eddies
679 (Chaigneau et al., 2011; Laxenaire et al., 2019; Pegliasco et al., 2015). Grid cells located inside each

680 eddy are defined within the SSH contour associated with the maximum mean surface velocity (interior
 681 grid cells). Grid cells located between the outer most closed contour and within 1.5 radius of the
 682 eddy center and not within another eddy were used to define background conditions outside of
 683 eddies (exterior grid cells). Only eddies with areas larger than an equivalent circular diameter of
 684 100km and not within the Loop Current were considered in the analysis. On average, 3.78 cyclonic
 685 and 3.33 anticyclonic eddies were identified in each daily velocity field. We find that cyclonic
 686 eddies were associated with 10% higher secondary production relative to exterior grid cells and the
 687 ratio of secondary production in interior cells to exterior cells ranged from 0.4 to 3.37 (95% CI). In
 688 contrast, secondary production was substantially lower inside anticyclonic eddies accounting for only
 689 46% of the average secondary production in exterior cells (0.03 - 1.87 (95% CI)). In addition to their
 690 convergent nature that dampens nutrient input, lower rates of secondary production in anticyclonic
 691 eddies can likely be attributed to the presence of highly oligotrophic Loop Current water trapped within
 692 large anticyclonic LCE.



693
 694 **Figure 9 (A-F):** Vertically integrated secondary production ($\text{mg C m}^{-2} \text{d}^{-1}$) by simulated large
 695 zooplankton (LZ, top) and predatory zooplankton (PZ, bottom). Annual average of secondary
 696 production for LZ (A) and PZ (D). Instantaneous model output of secondary production in winter
 697 for LZ (B) and PZ (E) on simulated day 4 February 2012. Instantaneous model output for
 698 secondary production in summer for LZ (C) and PZ (F) on 2 August 2011.

699 **Table1:** Average seasonal minimum and maximum values in the model (1993-2012) and the day
 700 of year in which they occur for surface chlorophyll (mg m^{-3}) and depth-integrated estimates of
 701 phytoplankton biomass (mg C m^{-2}), net primary production ($\text{mg C m}^{-2} \text{d}^{-1}$), mesozooplankton

702 biomass (mg C m^{-2}), and mesozooplankton secondary production ($\text{mg C m}^{-2} \text{ d}^{-1}$) calculated by
 703 spatially averaging daily fields over the oligotrophic (upper half of table) and shelf regions (lower
 704 half of table). Day of year values are in the format “day/month \pm days.”

	Daily Field Value		Day of Year	
Diagnostic (Oligotrophic)	Annual Min.	Annual Max.	Day of Min.	Day of Max.
Surface Chlorophyll	0.09 ± 0.005	0.27 ± 0.06	$9/9 \pm 23$	$1/29 \pm 13$
Phytoplankton Biomass	2300 ± 130	3600 ± 140	$12/26 \pm 7$	$4/29 \pm 17$
Net Primary Production	290 ± 70	1000 ± 120	$12/31 \pm 12$	$7/6 \pm 27$
Mesozooplankton Biomass	1000 ± 40	1400 ± 90	$1/1 \pm 4$	$5/19 \pm 18$
Secondary Production	18 ± 4	68 ± 10	$12/31 \pm 10$	$6/4 \pm 15$
Diagnostic (Shelf)	Annual Min.	Annual Max.	Day of Min.	Day of Max.
Surface Chlorophyll	1.96 ± 0.15	3.00 ± 0.30	$2/8 \pm 37$	$7/31 \pm 58$
Phytoplankton Biomass	3200 ± 290	5200 ± 440	$1/1 \pm 9$	$7/18 \pm 11$
Net Primary Production	750 ± 120	2000 ± 220	$12/31 \pm 8$	$7/21 \pm 14$
Mesozooplankton Biomass	670 ± 70	1100 ± 90	$12/29 \pm 7$	$5/23 \pm 25$
Secondary Production	94 ± 17	270 ± 28	$12/31 \pm 6$	$7/20 \pm 16$

705

706 4.0 Discussion

707 Many parameters in biogeochemical models are poorly constrained by observations and laboratory
 708 studies and/or highly variable in the environment. The numbers and uncertainties around these
 709 parameters allow PBMs with varying degrees of tuning to reproduce a single ecosystem attribute
 710 (e.g., surface Chl) even if multiple processes are inaccurately represented (Anderson, 2005;
 711 Franks, 2009). Once validated, one of the main values of coupling physical and biogeochemical
 712 models (i.e. PBMs) is their utility for making inferences about portions of the lower trophic level
 713 that are under sampled and/or difficult to measure in the field. If PBMs are to be utilized to explain
 714 variability rather than simply fit an observational dataset, multiple ecosystem attributes, must be
 715 validated and the underlying model structure and assumptions critically evaluated. In the section
 716 below, we further justify changes to model structure by evaluating the underlying assumptions in
 717 default NEMURO and discuss model-data mismatch before drawing conclusions about the GoM
 718 zooplankton community and the implications of its dynamics for higher trophic levels.

719 4.1 Justification for NEMURO modifications

720 The phytoplankton community in the North Pacific (NP) domain where NEMURO was originally
721 designed is largely composed of nanoplankton (original SP) and microplankton (original LP). By
722 default, SP is assumed to represent coccolithophores and autotrophic nanoflagellates, which can
723 be important prey of copepods and other mesozooplankton in temperate and subpolar regions
724 (Kishi et al., 2007). However, in tropical regions such as the GoM, smaller picophytoplankton taxa
725 typically dominate, particularly in highly oligotrophic regions. Common picophytoplankton of the
726 GoM include cyanobacteria and picoeukaryotes, which are too small for most mesozooplankton
727 to feed on. Hence, the SP to LZ grazing pathway was removed in our model. We found that
728 removal of this grazing pathway allowed the model to simulate a more realistic phytoplankton
729 community on the shelf region. Despite intuition, SP largely dominated the shelf region in the
730 model when LZ were allowed to graze on SP. After closer inspection, we found that grazing of SP
731 sustained LZ biomass on the shelf to levels where top-down pressure constrained LP standing
732 stocks. This prevented large blooms of LP, leading to a competitive advantage for SP even in
733 highly eutrophic conditions (the Mississippi river delta), which was observed for a wide range of
734 LP maximum growth rates, LP half-saturation constants, and LZ/PZ grazing rates. Thus, removal
735 of the SP to LZ grazing pathway added ecological realism and improved the model solution.

736 During the model tuning process (outlined in the supplemental), we also found that, despite a wide
737 range of tested parameter sets, the model was unable to simulate mesozooplankton biomass low
738 enough to match SEAMAP observations in the oligotrophic region. Even with unrealistically low
739 phytoplankton biomass, equivalent to approximately 50% of surface Chl observed in SeaWIFS
740 images, the model overestimated mesozooplankton biomass. To achieve realistic levels of
741 mesozooplankton biomass in the oligotrophic region, default LZ and PZ mortality parameter
742 values needed to be increased by an order of magnitude. However, this produced unrealistically
743 high loss rates on the shelf region, leading to mesozooplankton biomass estimates that were
744 substantially lower than SEAMAP shelf observations. Implementation of linear mortality on all
745 biological state variables (except PZ) resolved this issue by providing the model with a greater
746 dynamic range. In NEMURO, and other biogeochemical models, quadratic mortality is often used
747 to increase model stability and/or is mechanistically justified as representing the impact of
748 unmodeled predators that co-vary in abundance with prey (Gentleman and Neuheimer, 2008;
749 Steele and Henderson, 1992). However, grazing losses of all state variables (except PZ), are
750 already explicitly modeled in NEMURO by default. Hence, removal of quadratic mortality also

751 added ecological realism and improved the model solution. Quadratic mortality was retained for
752 PZ to account for the implicit predation pressure of un-modeled predators (e.g. planktivorous fish).

753 **4.2 Model-data mismatch**

754 **4.2.1 Surface chlorophyll discrepancies**

755 Within our model-data comparisons of surface Chl we find that NEMURO-GoM reproduces
756 important patterns in both the oligotrophic and shelf region. The latter of which, apart from the
757 northern shelf, has not been well resolved by previous PBMs (e.g., Gomez et al., 2018; Xue et al.,
758 2013). The absence of a shelf Chl signature may, in some cases, be overly attributed to bias in
759 satellite measurement due to high concentrations of colored dissolved organic matter (CDOM).
760 While a clear shelf signature is well resolved in NEMURO-GoM, the model-data mismatch is
761 greater on the shelf compared to oligotrophic regions. This is an expected result considering that
762 the model incorporates climatological river forcing while actual variability is much more complex.
763 Furthermore, the absence of CDOM in the model likely contributes to the overestimation of
764 phytoplankton biomass on the shelf.

765 In future studies, the inclusion of daily nutrient data like that produced for the Mississippi River
766 by USGS starting in 2011 is needed for PBMs to better resolve variability on the shelf. Including
767 benthic processes, such as denitrification (Fennel et al., 2006), may also reduce model-data
768 mismatch in shelf regions. Implementing more realistic light attenuation (e.g. wavelength-specific
769 light attenuation or inclusion of CDOM) could further improve estimates of phytoplankton
770 biomass on the shelf as primary production can be sensitive to different light attenuation
771 formulations (Anderson et al., 2015). In our model, it was difficult to simulate deep DCMs in the
772 oligotrophic GoM while also simulating DCMs on the shelf that were shallow enough to maintain
773 high nitrate. This may reflect the need for more realistic light attenuation in the model. Quantifying
774 uncertainty in C:Chl ratios is also an important task moving forward, which may decrease model-
775 data mismatch on the shelf as well as other regions. Future PBMs will likely continue to depend
776 heavily on satellite Chl for the bulk of model validation and hence more in situ samples are needed
777 to assess changes in phytoplankton light harvesting pigments along gradients from coastal to
778 oligotrophic regions and from the surface to the DCM. Without these observations, it is difficult
779 to gauge mismatches between model and satellite ocean color products or in situ profiles of Chl.

780 In our model, the most noticeable surface Chl model-data mismatch occurs on the southern GoM
781 shelf (Campeche Bank (CB)), where the model consistently overestimates surface Chl. This bias
782 was also notable in the GoM PBM implemented by Damien et al. (2018), particularly in winter.
783 We believe this discrepancy may be driven by a combination of errors involving overestimation
784 of shelf mixing by the hydrodynamic model, entrainment of high Chl water (given the
785 overestimated DCM magnitude in the model), or errors in the open boundary conditions which
786 result in an overestimation of upwelled nutrients/biomass near the YP that are transported
787 westward by shelf currents. We found that the CB model-data mismatch was reduced when open
788 boundary conditions included nitracline depths of greater than 100 m. This may reflect realistic in
789 situ conditions considering that Caribbean water entering the GoM is highly oligotrophic. During
790 our process cruises, nitrate was often undetectable above 100 m in samples collected near the Loop
791 Current (A. Knapp, pers. comm.).

792 Although modifying the boundary conditions may be justified, deepening the nitracline at the
793 boundaries made it increasingly difficult to sustain realistic surface phytoplankton biomass in the
794 oligotrophic GoM. This may point to the importance of nitrogen fixing cyanobacteria, which
795 provide an alternative source of new nitrogen (other than upwelling and mixing) that could be
796 supporting phytoplankton at the surface given the strong stratification and deep nitraclines in the
797 GoM. In the process of model tuning, we noticed that increasing the DON pool by increasing the
798 PON to DON decomposition rate was necessary to maintain both relatively deep nitraclines and
799 realistic surface Chl by providing a slow leeching of ammonium near the surface through bacterial
800 communities. The need for this slow production of ammonium in surface layers may compensate
801 for nitrogen fixation, which is not included in NEMURO (Holl et al., 2007; Mulholland et al.,
802 2006). In future studies, including diazotrophs as a separate phytoplankton functional type would
803 be essential for evaluating the importance of nitrogen fixation in the GoM.

804 Despite the model-data mismatch on the CB, this discrepancy appears to have little impact on the
805 rest of the GoM. However, the model overestimates surface Chl in the southwestern GoM, which
806 can likely be attributed to entrainment of high Chl water originating from the CB. Locally, the
807 ecological impact is likely more significant. Higher phytoplankton biomass would be expected to
808 support higher mesozooplankton grazing rates and secondary production. Indeed, some of the

809 highest model rates of secondary production occur on the CB. Hence, the surface Chl model-data
810 mismatch may lead to an overestimation of secondary production for this region.

811 **4.2.2 Deep chlorophyll maximum discrepancies**

812 Since most PBMs focus on validating against satellite derived surface chlorophyll, the dynamics
813 of the DCM is often insufficiently investigated. Consequently, many models predict DCM depths
814 that are far too shallow. Identifying this issue in the literature proved to be difficult because most
815 studies do not provide profiles of simulated Chl (an exception is the recent GoM PBM by Damien
816 et al. (2018)). We note that DCM depths in the DIAZO model (Stukel et al., 2014) were often quite
817 shallow or completely nonexistent in the portion of the domain that included the oligotrophic GoM
818 region. Underestimates of DCM depth in the unmodified COBALT biogeochemical model has
819 also been identified (Moeller et al., 2019). In our investigation of the PBM implemented by Gomez
820 et al. (2018), we found that DCMs in the oligotrophic region were commonly shallow and weak.
821 In the default NEMURO simulation, DCM depths in the oligotrophic region were typically at a
822 depth of 25 m, which is much shallower than observed (SEAMAP: 80 ± 25 m, Process cruises:
823 107 ± 21 m). While this issue may seem insignificant, particularly if a study is focused on mixed-
824 layer dynamics, accurate placement of the DCM can have profound impacts on PBM behaviors,
825 because the DCM is typically co-located with the nitracline. Unrealistically shallow DCMs and
826 nitraclines permit unrealistically high nitrate fluxes into the surface layer following mixing events;
827 thus, validating the DCM in PBMs is critical.

828 For these reasons, we devoted substantial effort to tuning phytoplankton dynamics in the DCM.
829 Modifications to α (the slope of the photosynthesis-irradiance curve) and attenuation coefficients
830 allowed us to move the DCM down to realistic depths. Inclusion of a variable C:Chl module was
831 also implemented to better resolve the DCM. However, an additional issue was present in the
832 default NEMURO simulations, the NEMURO-GoM, and every simulation that we attempted. In
833 all simulations that formed DCMs, the location of the DCM was always co-located with a
834 maximum in phytoplankton specific growth rate, even though field measurements indicate that
835 phytoplankton growth rates and NPP are either relatively constant with depth or decline in the
836 DCM. This is not surprising, given the low photon flux at the base of the euphotic zone and the
837 energetic demands required to up-regulate cellular density of light harvesting pigments.
838 Additionally, our field measurements show that the DCM was not associated with a biomass

839 maximum (biomass was fairly constant with depth), suggesting that DCM formation in the GoM
840 is physiologically driven.

841 We believe this DCM dynamical issue was responsible, in part, for the underestimation of specific
842 phytoplankton growth and microzooplankton grazing rates by the model despite estimating higher
843 NPP (**Fig. 4D**). The model-data mismatch at the DCM may also be associated with an
844 overestimation of phytoplankton biomass which would decrease the vertical transport of nitrate
845 and hence inhibit primary and secondary production at the surface. Underestimates of primary and
846 secondary production would be expected to be greatest with a deep DCM and shallow MLD, which
847 occur during the summer months in the GoM. Indeed, we found that nitrate concentrations above
848 the nitracline were considerably lower in the model relative to observations. Future PBM studies
849 need to focus more effort on resolving dynamics of the DCM.

850 **4.2.3 Mesozooplankton grazing discrepancies**

851 Novel to this study, model estimates of mesozooplankton biomass are shown to agree closely with
852 observations on the shelf and in the oligotrophic GoM. To our knowledge, this study includes the
853 first quasi-regional zooplankton biomass model validation in a PBM. Our model also provides the
854 first model-data comparisons of size-specific zooplankton biomass and grazing rates for the GoM.
855 Such comparisons provide valuable insights into the potential biases of traditional functional group
856 biogeochemical models pertaining to zooplankton dynamics (Everett et al., 2017). While
857 NEMURO-GoM shows broad agreement with zooplankton observations, some model-data
858 mismatch occurs, particularly for mesozooplankton grazing rates.

859 We identify three factors that may explain the model-data mismatch for mesozooplankton grazing
860 rates. The first and most obvious factor is the temporal sampling discrepancy as measurements
861 were collected outside our model simulation period. Model-data mismatch may also arise from
862 inaccuracies in the field measurements. During our process cruises, the zooplankton gut pigment
863 measurements were based solely on phaeopigment content due to phytodetrital aggregates and
864 *Trichodesmium* colonies found in our zooplankton net tows, which can lead to substantial
865 contamination. Thus, true mesozooplankton grazing rates were likely underestimated because
866 undegraded Chl can be abundant in the foreguts of mesozooplankton. Furthermore, the gut pigment
867 approach assumes that any group of mesozooplankton has a constant gut throughput time (as a

868 function of temperature), which is an oversimplification. Uncertainties in model grazing
869 formulations could also contribute to model-data mismatch (Gentleman et al., 2003a; Sailley et al.,
870 2015). Future in situ grazing measurements are needed to enable an objective selection of grazing
871 formulations and parameter values. In particular, field studies that shed light on prey selectivity
872 would be useful for parameterizing PBMs with multiple mesozooplankton functional groups, such
873 as NEMURO-GoM.

874 Clear model-data mismatch is also evident in the proportion of grazing mediated by PZ and LZ.
875 This may be due to the fact that PZ is by default explicitly defined and parameterized as a higher
876 trophic level mesozooplankton that can feed on LZ. In reality, while there is a correlation between
877 size and trophic level in the ocean, many predatory zooplankton are <1 mm, and many suspension-
878 feeding zooplankton are >1 mm; hence, the overlap of taxonomic groups with different functional
879 roles and sizes makes it difficult to directly compare model categories to field data. For example,
880 shelf suspension-feeding zooplankton are likely larger than their counterparts in the oligotrophic
881 GoM although their functional role in the ecosystem does not change between environments.

882 The ecological impact of the model's potential overestimation of LZ grazing rates is most likely
883 to manifest through an increase in the ratio of secondary production to mesozooplankton biomass.
884 Since both LZ and PZ biomasses are accurately modeled by NEMURO-GoM, the overestimation
885 of grazing rates suggests that LZ turnover times may be too high, thus leading to higher estimates
886 of secondary production. However, this interpretation may oversimplify the complex interactions
887 within pelagic protistan communities. In the oligotrophic region where our model overestimates
888 LZ grazing rates, the model indicates that heterotrophic protists comprise approximately half of
889 the LZ diet. Thus, overestimates of grazing on LP do not necessarily lead to overestimates in total
890 consumption if <1-mm zooplankton derive substantial nutrition from non-phototrophic sources in
891 the field. Furthermore, the model's construction (i.e., LZ and PZ are functional groups, while the
892 field data are size classes) suggests that part of the model-data mismatch in Fig. 5d may result
893 from the presence of some suspension-feeders (i.e., LZ) in the >1-mm zooplankton and some
894 carnivorous zooplankton (i.e., PZ) in the <1-mm zooplankton. In this case, the model may simply
895 attribute too high of a LP:SZ prey ratio to LZ. If this is the issue, the model's estimate of LZ
896 secondary production may be accurate, but its trophic level too low (or, conversely, the trophic
897 level of PZ too high). Direct assessments of zooplankton trophic position (e.g., by compound

898 specific isotopic analysis of amino acids,(Chikaraishi et al., 2009; Décima et al., 2017) may help
899 resolve these issues.

900 **4.3 Mesozooplankton dynamics in the open-ocean oligotrophic Gulf of Mexico**

901 Despite its nutrient-poor conditions, the open-ocean GoM ecosystem is a key region for spawning
902 and larval development of many commercially important fishes, including Atlantic bluefin tuna,
903 yellowfin tuna, skipjack tuna, sailfish and mahi mahi (Cornic and Rooker, 2018; Kitchens and
904 Rooker, 2014; Lindo-Atichati et al., 2012; Muhling et al., 2017; Rooker et al., 2012, 2013). Why
905 so many species choose such oligotrophic waters as habitat for their larval stages is unknown, but
906 may be due to reduced predation risk (Bakun, 2013; Bakun and Broad, 2003). Regardless, rapid
907 growth and survival through the larval period depends on mesozooplankton prey that are suitably
908 abundant and appropriately sized for these larval fishes. These prey taxa may be especially
909 sensitive to increased stratification and oligotrophication associated with climate change, making
910 investigation of their dynamics and production an important topic of research.

911 Mesozooplankton biomass in the oligotrophic GoM was found to be strikingly low in both
912 observations and model estimates, approximately an order of magnitude less than on the shelf.
913 Model results clearly show that this low biomass condition arises from bottom-up resource
914 limitation. Our results suggest that low phytoplankton biomass in oligotrophic regions, and
915 particularly within Loop Current Eddies, may even lead to localized and episodic regions where
916 phytoplankton concentrations approach thresholds low enough that trigger collapse of
917 mesozooplankton grazing. Prey limiting conditions for mesozooplankton and their predators
918 would be expected more frequently in the GoM during warmer ocean conditions. Higher sea
919 surface temperatures and increased thermal stratification could suppress vertical mixing, resulting
920 in lower phytoplankton biomass.

921 Despite extreme oligotrophy and dominance of picophytoplankton, our model shows that both PZ
922 and LZ can be sustained at modest abundances in the oligotrophic GoM. Indeed, the substantial
923 abundances of large (>1 mm) mesozooplankton, equivalent to 60% of total mesozooplankton in
924 both observations and model results (**Fig. 4A, C**), is an important result that helps explain the
925 success of larval fish in the region. Our results show that large mesozooplankton (PZ) occupy a
926 trophic position of approximately 3.0 in the open ocean GoM, which is marginally lower than on

927 the shelf where they feed primarily on small mesozooplankton (LZ). The change in trophic position
928 is associated with a switch from carnivory to feeding predominantly on heterotrophic protists in
929 the oligotrophic region. This result highlights the importance of intermediate protistan trophic
930 levels in sustaining mesozooplankton communities in oligotrophic regions. Indeed, both LZ and
931 PZ ingest proportionally more SZ in the open ocean than on the shelf. Notably, these protistan
932 trophic steps cannot be quantified by routine field techniques because they have no pigment
933 signature to make them visible in gut pigment measurements and may not enrich in bulk ^{15}N
934 leading to isotopic invisibility from a trophic perspective (Gutiérrez-Rodríguez et al., 2014).
935 Despite their importance to phytoplankton grazing, they are sometimes missing from GoM
936 ecosystem models (e.g., Fennel et al., 2011) and severely underrepresented or even absent in
937 complex mass-balance constrained models (Arreguin-Sanchez et al., 2004; Geers et al., 2016).
938 New insights may arise from focused investigations of phytoplankton → protist → crustacean
939 linkages in oligotrophic regions in both model and experimental studies. This will likely require
940 the use of next-generation technologies such as compound specific isotopic analyses of specific
941 amino acids that have been shown to enrich in protists (Décima et al., 2017) or DNA
942 metabarcoding to assess zooplankton gut contents (Cleary et al., 2016).

943 Another robust finding from this study is the dynamic mesoscale variability in zooplankton
944 abundance, diet, and trophic position. These model results highlight the impact of Loop Current
945 Eddies and mesoscale fronts and other features in modifying the biogeochemistry and food web of
946 the GoM. The existence of hot spots of productivity in the GoM has been noted in observational
947 studies (Biggs and Ressler, 2001), and the importance of GoM mesoscale features to fish larvae
948 has been hypothesized (Domingues et al., 2016; Lindo-Atichati et al., 2012; Rooker et al., 2012).
949 Indeed, cyclonic eddies were found to have enhanced secondary production in our model, while
950 secondary production was depressed within anticyclonic eddies. Our results further suggest that
951 these mesoscale structures may not only modify zooplankton abundances, but also their trophic
952 roles in the ecosystem, with implications for the transfer efficiencies of carbon and nitrogen in the
953 pelagic food web.

954 **5.0 Conclusions**

955 We used an extensive suite of in situ measurements to validate zooplankton dynamics simulated
956 by a PBM of the GoM. The model was able to capture broad patterns in phytoplankton and

957 mesozooplankton abundances, depth of the DCM and nutricline, as well as growth and grazing
958 patterns. Using the validated model to investigate characteristics of the GoM mesozooplankton
959 community, our results suggest that small mesozooplankton are largely herbivorous and large
960 mesozooplankton largely carnivorous on the GoM shelf. However, distinct changes in diet were
961 noted in the oligotrophic GoM, where both groups rely more on protistan prey. Changes in diet
962 and secondary production highlighted in this study have the potential to impact food availability
963 to higher trophic levels, such as pelagic larval fishes. In future work, we plan to couple our model
964 to an individual-based model of larval fish to evaluate the extent to which mesozooplankton
965 abundance limits larval fish feeding and growth along their transport pathways in the GoM.
966 Insights from this ecosystem-based approach may help to better resolve stock-recruitment
967 relationships that are needed for sustainable fisheries management and improved stock-assessment
968 models.

969

970 *Code and data availability.*

971 The model code and model validation data used in this study can be downloaded from GitHub at
972 <https://github.com/tashrops/NEMURO-GoM>. An idealized one-dimensional version of
973 NEMURO-GoM written in Matlab is also provided. The three-dimensional NEMURO-GoM
974 model outputs used in the study are available on the FSU-COAPS server in a Network Common
975 Data Form (NetCDF format).

976 *Author Contribution.*

977 TAS conducted all numerical simulations and model analysis. EPC, SLM, and AB provided
978 expertise on the hydrodynamic modeling. MRS and VJC provided expertise on the biogeochemical
979 model coding and tuning. RS, MRL, and GZ processed and provided data that was central to
980 NEMURO-GoM's validation. TAS wrote the manuscript with contributions from all authors.

981 *Competing interest.*

982 The authors declare that they have no conflict of interest

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