Geochemical zones and environmental gradients for soils from the Central Transantarctic Mountains, Antarctica

Melisa A. Diaz^{1,2±*}, Christopher B. Gardner^{1,2}, Susan A. Welch^{1,2}, W. Andrew Jackson³, Byron J.
 Adams⁴, Diana H. Wall⁵, Ian D. Hogg^{6,7}, Noah Fierer⁸, W. Berry Lyons^{1,2}

⁵ School of Earth Sciences, The Ohio State University, Columbus, OH, USA

⁶ ²Byrd Polar and Climate Research Center, The Ohio State University, Columbus, OH, USA

7 ³Department of Civil, Environmental, & Construction Engineering, Texas Tech University, Lubbock, TX, USA

⁴Department of Biology, Evolutionary Ecology Laboratories, and Monte L. Bean Museum, Brigham Young University,
 Provo, UT, USA

- 10 ⁵Department of Biology and School of Global Environmental Sustainability, Colorado State University, Fort Collins, CO,
- 11 USA

12 ⁶Canadian High Arctic Research Station, Polar Knowledge Canada, Cambridge Bay, Nunavut, Canada

13 ⁷School of Science, University of Waikato, Hamilton, New Zealand

14 ⁸Department of Ecology and Evolutionary Biology and Cooperative Institute for Research in Environmental Science,

15 University of Colorado Boulder, Boulder, CO, USA

16

[†]Now at Departments of Geology and Geophysics and Applied Ocean Physics and Engineering, Woods Hole Oceanographic
 Institution, Woods Hole, MA, USA

19

20 Correspondence to: Melisa A. Diaz (<u>diaz.237@osu.edumdiaz@whoi.edu</u>) 21

22 Abstract. Previous studies have established links between biodiversity and soil geochemistry in the McMurdo Dry Valleys,

23 Antarctica, where environmental gradients are important determinants of soil biodiversity. However, these gradients are not

24 well established in the Central Transantarctic Mountains, which are thought to represent some of the least hospitable

25 Antarctic soils. We analyzed 220 samples from 11 ice-free areas along the Shackleton Glacier (~85 °S), a major outlet

26 glacier of the East Antarctic Ice Sheet. We established three zones of distinct geochemical gradients near the head of the

27 glacier (upper), central (middle), and at the mouth (lower). The upper zone had the highest water-soluble salt concentrations

28 with total salt concentrations exceeding $80,000 \ \mu g \ g^{-1}$, while the lower zone had the lowest water-soluble N:P ratios,

29 suggesting that, in addition to other parameters (such as proximity to water/ice), the lower zone likely represents the most

30 favorable ecological habitats. Given the strong dependence of geochemistry with on geographic parameters, we established

31 <u>developed</u> multiple linear regression and random forest models to predict soil geochemical trends given latitude, longitude,

32 elevation, distance from the coast, distance from the glacier, and soil moisture (variables which can be inferred from remote

33 measurements). Confidence in our random forest model predictions was moderately high, with R² values for total water-

34 soluble salts, water-soluble N:P, ClO₄⁻, and ClO₃⁻ of 0.851, 0.8842, 0.7840, and 0.7428, respectively. These modeling results

- 35 can be used to predict geochemical gradients and estimate salt concentrations for other Transantarctic Mountain soils,
- 36 information that can ultimately be used to better predict distributions of soil biota in this remote region.

37 1. Introduction

38 From an ecological standpoint, tThe least biologically diverse terrestrial systems are those found in extreme 39 physical and chemical environments. The abundance and diversity of life in soils is dependent on a number of environmental 40 environmental variablesparameters, including temperature, precipitation, organic matter content, and nutrient availability 41 (Wall et al., 2012). Hot deserts are typically viewed as one of the least biologically diverse environments, <u>--However, but</u> cold 42 deserts can often be even less diverse (Freckman and Virginia, 1998). Soils in Antarctica typically serve as end-members for 43 low habitat suitability due to their high salt concentrations, low organic carbon, low soil moisture, and low mean annual 44 temperatures (Courtright et al., 2001).

45 In the McMurdo Dry Valleys (MDV), organic matter and salt concentrations influence soil communities, where 46 soils with higher amounts of organic carbon, lower water-soluble N:P ratios, and lower total water-soluble salt 47 concentrations generally harbor the greatest biomass and biodiversity (Barrett et al., 2006; Bottos et al., 2020; Caruso et al., 48 2019; Magalhães et al., 2012). These Antarctic ecosystems are relatively simple and are among the only knownfew known 49 soil systems where nematodes and microarthropods (Collembola, Acari) are at the top of the food chain (Freckman and 50 Virginia, 1998; Hogg and Wall, 2012). Studies of soils in the MDV and Transantarctic Mountains (TAM) have been key to 51 understanding ecosystem structure and function in extreme terrestrial environments (e.g. Caruso et al., 2019; Collins et al., 52 2019, 2020; Convey and McInnes, 2005; Freckman and Virginia, 1998; Hodgson et al., 2010).

53 Biological processes in Antarctic soils are largely dependent on the availability, duration, and proximity of soils to 54 liquid water (Barrett et al., 2006). Due to the seasonality in freezing and of thawing events, liquid water acts as a pulse to the 55 ecosystem, providing water for organisms, but also wetting surface soils and dissolving soluble salts (Webster-Brown et al., 56 2010; Zeglin et al., 2009). Experiments of salt thresholds on Antarctic nematodes found that no individuals survived in 57 highly saline soils over (\sim 2,600 mg L⁻¹ TDS) (Nkem et al., 2006). Concentrations of soluble salts exist at these 58 concentrations or higher forat high elevation and inland locations in the TAM (Bockheim, 2008; Lyons et al., 2016). 59 Additionally, studies on TAM soils have found that increased salt concentrations lead to a decrease in soil biodiversity in 60 older soils compared to younger soils (Magalhães et al., 2012). Yet, despite these inhospitable conditions (e.g. high salt 61 concentrations and glacial advance and retreat), some organisms are postulated to have found suitable refugia in TAM soils 62 and persisted in isolation for millions of years and through glacial cycles (Beet et al., 2016; Collins et al., 2019, 2020; 63 Stevens et al., 2006; Stevens and Hogg, 2003).

It is generally accepted that habitat suitability for invertebrate species in Antarctic soils is driven by a combination of geochemical, geographic, <u>hydrologic</u>, and geomorphic variables (Bottos et al., 2020; Courtright et al., 2001; Freckman and Virginia, 1998; Magalhães et al., 2012). Geographic variables, such as elevation, can be measured with advanced mapping tools and satellite imagery; however, surface exposure ages, soil geochemistry and nutrient content require extensive logistical support and resource allocation for sample collection and analysis. <u>More efficient estimation toolsA</u>

69	better understanding of the relationship between geographic variables and on-the-ground measurements areis needed to aid
70	in our ability to understand and predict habitat suitability for invertebrates throughout the TAM.
71	With this study, we determined and evaluated geochemical patterns and gradients of water-soluble ions in soils

72 collected from 11 ice-free areas along the Shackleton Glacier, Central Transantarctic Mountains (CTAM). Particular 73 attention was given to total water-soluble salt concentrations, N:P ratios, and ClO₄⁻ and ClO₃⁻ concentrations, based on their 74 influence on biodiversity, as determined in previous studies (e.g. Ball et al., 2018; Barrett et al., 2006b; Courtright et al., 75 2001; Dragone et al., 2020; Nkem et al., 2006). The geochemical data were compared to geographic parameters to 76 understand how the physical environment influences the observed geochemical variability. Our results show that water-77 soluble ion concentrations and distributions are driven largely by soil geography and surface exposure age. Finally, we 78 implemented statistical and machine learning techniques to interpolate and predict the soil geochemistry across the region 79 using geographic variables. Our multiple linear regression and random forest models show that latitude, longitude, elevation, 80 distance from the coast, distance from the glacier, and soil moisture (all variables currently or soon to be remotely 81 measurable using maps and satellites) are moderately effective at estimating spatial patterns in TAM soil geochemistry, with 82 R² values as high as 0.87. These data will be particularly useful for ecologists seeking to understand refugia and habitat 83 suitability in Antarctica and similarly harsh, desert environments.

84 2. Study sites

The Shackleton Glacier (~84.5 to 86.4°S; ~130 km long and ~10 km wide) is a S-N trending outlet glacier of the East Antarctic Ice Sheet (EAIS) located to the west of the Beardmore Glacier and flows through the Queen Maud Mountains (CTAM) into the Ross Sea (Fig. 1). The elevations of exposed soils range from ~150 m.a.s.l. to >3,500 m.a.s.l. from the coast towards the Polar Plateau. Long-term climate data are not yet available, but the Shackleton Glacier region is a polar desert regime, similar to the Beardmore Glacier region, with average annual temperatures well below freezing and little precipitation (LaPrade, 1984).

91 During the Last Glacial Maximum (LGM) and glacial periods throughout the Pleistocene, the size and thickness of 92 the EAIS has been suggested to be was likely greater than current levels (Golledge et al., 2013; Nakada and Lambeck, 1988; 93 Talarico et al., 2012; Wilson et al., 2018). Outlet glaciers, such as the Shackleton Glacier, may have had the greatest 94 increases in extent, especially towards at the glacier terminus (Golledge et al., 2012; Golledge and Levy, 2011). The 95 behavior of local alpine and tributary glaciers is not well-constrained, but these glaciers are also believed to have advanced 96 and retreated over the last two million years (Diaz et al., 2020a; Jackson et al., 2018). As a result, currently exposed soils 97 were overlain and reworked by fluctuations of the Shackleton Glacier and other tributary and alpine glaciers in the region. 98 Exposure ages range from the early Holocene to the Miocene, and generally increase with distance from the coast and 99 distance from the glacier (Balter-Kennedy et al., 2020; Diaz et al., 2020a).

- 100 The soils contain a range of water-soluble salts derived primarily from atmospheric deposition and chemical
- 101 weathering (Claridge and Campbell, 1968; Diaz et al., 2020b). The major salts are typically nitrate and sulfate salts,
- 102 especially at higher elevations and further inland from the coast of the Ross Sea (Diaz et al., 2020b). The solubilities of the
- 103 salts vary, but nitrate salts are highly soluble and their occurrence at high elevation and inland locations suggests that those
- 104 soils have maintained persistent arid conditions.

105 **3. Methods**

106 3.1. Sample collection and preparation

107 During the 2017-2018 austral summer, 220 surface soil samples (~top 5 cm) were collected from 11 distinct ice-free 108 areas (Roberts Massif, Schroeder Hill, Mt. Augustana, Bennett Platform, Mt. Heekin, Thanksgiving Valley, Taylor Nunatak, 109 Mt. Franke, Mt. Wasko, Nilsen Peak, and Mt. Speed) along the Shackleton Glacier, including a subset of 27 samples 110 previously analyzed for S, N, and O isotopes in nitrate and sulfate (Diaz et al., 2020b). At each area, we collected samples in 111 transects (ranging from ~200 m to ~2,000 m in length) to maximize the geochemical variability. Our transects were also 112 designed to capture the LGM transition, with some soils exposed throughout the LGM and others exposed following glacier 113 retreat. GPS coordinates and elevations were recorded with each sample and later used to estimate the distance from coast 114 and distance from the glacier (defined as linear distance from the nearest glacier - Shackleton, tributary, or alpine). Once 115 collected, the samples were stored and shipped frozen (-20 °C) to The Ohio State University.

116Prior to geochemical analysis, the samples were dried at 50 °C for at least 72 hours with the loss in mass attributed117to soil moisture content. The dried soils were leached at a 1:5 soil to DI water ratio, and the leachate was filtered through 0.4118 μ m Nucleopore membrane filters (Diaz et al., 2018, 2020b; Nkem et al., 2006). Due to the low sediment to water ratio, this119leaching technique only dissolves the more water-soluble salts (Toner et al., 2013). These include salts with ClO₄⁻, NO₃⁻, Cl-,120SO₄²⁻, ClO₃⁻, and CO₃²⁻ + HCO₃⁻. Process blanks were generated and analyzed to account for any contamination from the121leaching process.

122 3.2. Analytical analysis of water-soluble anions, cations, and nutrients

123 The analytical techniques used here are similar to those reported by Diaz et al. (2020b). In brief, the analytes 124 included anions (F⁻, Cl⁻, Br⁻, and SO₄²⁻) which were measured on a Dionex ICS-2100 ion chromatograph, cations (K⁺, Na⁺, 125 Ca²⁺, Mg²⁺, and Sr²⁺) which were measured on a PerkinElmer Optima 8300 Inductively Coupled Plasma-Optical Emission 126 Spectrometer (ICP-OES), and nutrients (NO₃⁻ + NO₂⁻, PO₄³⁻, H₄SiO₄, and NH₃) which were measured on a Skalar San++ 127 Automated Wet Chemistry Analyzer at The Ohio State University. Perchlorate (ClO₄⁻) and chlorate (ClO₃⁻) were measured 128 using an ion chromatograph-tandem mass spectrometry technique (IC-MS/MS) at Texas Tech University (Jackson et al., 129 2012, 2015). All analytes are reported as listed. Total water-soluble salt concentration was calculated as the sum of all 130 measured cations and anions. The precision of replicated check standards and samples was typically better than 10% for all 131 major anions, cations and nutrients, and better than 20% for perchlorate and chlorate. Accuracy was typically better than 5% for all major anions, cations, and nutrients, as determined by the NIST 1643e external reference standard and the 2015 USGS
 interlaboratory calibration standard (M-216), and better than 10% for perchlorate and chlorate, as determined by spike

recoveries. Precision and accuracy for individual analytes are located in Table S1. Detection limits for the analytes have been
 previous reported (Diaz et al., 2018; Jackson et al., 2012).

3.3. Data interpolation and machine learning

136

137 Inverse distance weighted (IDW) interpolations were performed for Bennett Platform, Thanksgiving Valley, and

Roberts Massif using the Geostatistical Analyst tool in ArcMap 10.3. Since IDW is a deterministic method where unknown values are predicted based on proximity to known values, we chose those three sites as they had the most defined transects and relatively higher sample density. The interpolation parameters were constant with a power of 4, maximum neighbors of 15, minimum neighbors of 5, and 4 sectors, and a variable search radius. These parameters were chosen such that they optimize for the lowest mean absolute error.

143 Multiple linear regressions were generated for all geochemical analytes, except H4SiO4 (total of 15 dependent 144 variables), with latitude, longitude, elevation, distance from the coast, distance from the glacier, and soil moisture as 145 independent variables using built-in functions in R 3.6.3 (R Core Team, 2020). Random forest regression models were 146 similarly generated using the randomForest library. The random forest model is a machine learning algorithm that utilizes 147 supervised learning algorithms to predict values given input predictor variables (Breiman, 2001). Multiple decision trees are 148 run in parallel with a randomized subset of predictor variables, and the aggregate result of each tree is used to generate a 149 predicted outcome. Since each tree is generated using a random sample and random predictor variables, the random forest 150 model is effective at minimizing overfitting and handling outliers (Breiman, 2001). For both models, all geochemical data 51 were log-transformed to ensure the data were normally distributed (verified using a Jarque-Bera normality test). Missing 52 values were input as NA.

153Machine learning algorithms are widely used in variety of disciplines from finance (Patel et al., 2015) to ecology154(Davidson et al., 2009; Peters et al., 2007; Prasad et al., 2006), for both data prediction (regression) and classification.155Recently, these techniques have been used for Earth Science applications, including geologic mapping (Heung et al., 2014;156Kirkwood et al., 2016), air quality monitoring (Stafoggia et al., 2019), and water contaminant tracing (Tesoriero et al., 2017).157We developed a novel application of machine learning to predict concentrations and gradients of water-soluble salts in158Antarctic soils, given set geographic parameters, similar to the approaches developed for stock market and real estate159predictions (Antipov and Pokryshevskaya, 2012; Patel et al., 2015).

160For our random forest models, any sparse missing values in Table S2 were estimated by averaging the geochemistry161of the samples collected immediately before and after in the same transect. Missing values due to concentrations below the162detection limit were input as θ NA. The new imputed dataset was split into a training set representing 86% of the data (n =163189, Table S3) and a testing set representing the remaining 14% (n = 31, Table S4), based on ideal model parameters

164	described by Breiman (2001). The training dataset was used to generate the random forest models for each analyte. Each of
165	the models were run with 2000 decision trees (ntree = 2000) to minimize the mean squared errors. The number of random
166	variables used for each node split in the decision trees was set to the recommended regression default of variables/3 to
167	optimize the model randomness, which in our case was 2 (mtry = 2), following parameters described previously (Breiman,
168	2001). The scripts developed for both the multiple linear regression and random forest models are included in the
169	supplementary materials.

170 4. Results

171 4.1. Geochemistry of upper, middle, and lower zones

172The maximum, minimum, mean, standard deviation and coefficient of variation are reported in Table 1 for the173measured geographic and geochemical data. Concentrations of water-soluble ions span up to five orders of magnitude and174are variable across the region. Elevation, distance from the coast, distance from the glacier, and soil moisture are also175variable and span up to three orders of magnitude. The highest elevation samples (> 2,000 m.a.s.l.) were collected from176Schroeder Hill and the greatest soil moisture content is from Mt. Wasko at 12.3%, with a mean of 2.1% for all samples.177Shackleton Glacier region surface soils can be separated into three zones based on their water-soluble geochemistry:

178 an upper zone near the Polar Plateau, a middle zone near the center of the glacier, and a lower zone where the glacier flows 179 into the Ross Sea (Figs. 1; 2). The upper zone samples are characterized by the highest total water-soluble salt 180 concentrations, with the highest values greater than $80,000 \ \mu g \ g^{-1}$ at Schroeder Hill, while the lower zone samples have the 181 lowest total salt concentrations, with the lowest values near 10 µg g⁻¹ at Mt. Wasko (Fig. 2a-c). The middle zone has 182 intermediate values. Water-soluble N:P molar ratios generally follow a similar trend (Fig. 2d-f). The lowest N:P ratios are in 183 the lower zone soils, while the middle and upper zones have more variable values. Concentrations of CIO_4^- and CIO_3^- follow 184 similar trends as the total salts, with less distinction between middle and upper zones, though most concentrations in the 185 lower zone are below the detection limit (Fig. 2g-1; Table S2).

186 Observed trends between the zones appear to be driven, at least partially, by geography. Regressions of total water-187 soluble salt concentration, water-soluble N:P ratio, and ClO3⁻ concentration with elevation, distance from the coast, and 188 distance from the glacier are all positive (Fig. 2). The strongest relationships are between total salts and elevation, and Clos 89 N:P ratio and distance from the coastelevation, with R^2 values of 0.5926 and 0.5224, respectively, and p-values < 0.001 with 90 a Bonferroni Correction, which was applied to minimize the familywise type 1 error rate associated with multiple 91 comparisons (Fig. 2a;2dk). The weakest relationships are between $CIO_4CIO_4^-$ and distance from the coast, and CIO_3^- and 92 distance from the glacier, with R^2 values of 0.011 and 0.06, respectively4 (Fig. 2h; 2i). Distance from the glacier varies 193 widely between individual zones with frequent overlaps, but there appears to be a moderate relationship with N:P ratio and 194 total salts (Fig. 2c; 2f). Overall, total salt concentration has the strongest relationship with geography and ClO4⁻ has the 195 weakest relationships.

196 Ternary diagrams highlight the specific geochemical gradients within and between the zones. The anion ternary 197 diagram only includes SO_4^2 , NO_3^2 , and Cl², which are the major water-soluble salts in the region (Claridge and Campbell, 198 1968; Diaz et al., 2020b). Though carbonate and bicarbonate salts have been identified in both lacustrine sediments and soils 199 in Antarctica, previously measured concentrations in the Shackleton Glacier region were low, ranging from 0.07 to 2.5%, 200 and bicarbonate salts were not identified in the highest elevation and furthest inland soils (Claridge and Campbell, 1968: 201 Diaz et al., 2020b; Lyons et al., 2016). The most abundant anion for the upper zone is $SO_4^{2^-}$, which is greater than 99% of the 202 total anion budget in some Schroeder Hill and Roberts Massif samples, though other locations are dominated by NO3- (Fig. 203 3). The anions are more evenly distributed in the middle zone, though the majority of samples are most abundant in NO3⁻ and 204 Cl⁻. The lower zone has much lower SO_4^{2-} fractions than the upper zone and the dominant anion is generally Cl⁻. The cation 205 distribution is very similar for all three zones (Fig. 3). Na⁺ + K^+ is the most abundant cation pair representing over 90% of 206 the total cations for many upper and middle zone samples, while Ca^{2+} is the second most abundant. In general, Mg^{2+} is the 207 least abundant cation across all sampling locations.

208 4.2. Statistical geochemical variability

209 A principal component analysis (PCA) using the correlation matrix (i.e. scale = TRUE) was performed in R (using 210 factoextra (Kassambara and Mundt, 2017) and built in R software libraries) to determine which geochemical variables most 211 strongly differ across the samples. For the PCA, the first two principal components account for over 50% of the total dataset 212 variability at 44.2% and 11.6%, respectively. The different zones are correlated with different principal components (Fig. 4). 213 The samples from the middle zone are positively correlated with PC1 and PC2. In the biplot, they plot in the upper right 214 quadrant with high concentrations of Cl⁻, NO₃⁻, water soluble N:P ratio, and Ca²⁺, with a minor influence from soil moisture 215 and H₄SiO₄. The upper zone samples generally plot along PC1 and are most associated with Sr²⁺, SO₄²⁻, Mg²⁺, Na⁺, K⁺, F⁻, 216 ClO4, and ClO3. The samples from the lower, more coastal zone are negatively correlated with PC1 and are distinguished by 217 their higher PO43- concentrations. Most samples from all locations plot within the 95% confidence interval ellipses. 218 However, there are two strong outliers from Schroeder Hill and Mt. Heekin.

219 Similar to the PCA, we performed a simple Spearman's rank correlation for the entire dataset to visualize the 220 statistical dependence between all variables. Since a goal of this study is to relate water-soluble ion concentrations to 221 geography, we focused on latitude, longitude, distance from the coast, distance from the glacier, and soil moisture. The 222 strongest correlation coefficients are between Cl⁻ and latitude, elevation, and distance from the coast, and Sr^{2+} and soil 223 moisture (Fig. 5). Most other correlations are moderate to weak, though there appear to be notably stronger correlations 224 between ClO_3^- and latitude and distance from coast, Ca^{2+} and longitude, elevation, and distance from coast, NO_3^- and 225 latitude, and SO₄²⁻ with distance from glacier. Longitude, elevation, and distance from coast have the greatest number of 226 strong and moderate correlations with the geochemistry data. Outside of the geographic parameters, Na+ is highly correlated 227 with total water-soluble salts, likely representative of the high $Na^+ + K^+$ percentages (Fig. 3), and Sr^{2+} is highly correlated 228 with K⁺, likely reflecting a common ion source.

4.3. Spatial interpolation and machine learning model performance The total salt concentrations of individual samples at Bennett Platform produce the most defined interpolation gradient from the glacier front to further inland compared to Roberts Massif and Thanksgiving Valley (Fig. 6). Bennett

Platform also has the smoothest salt concentration contours suggesting that the interpolation model is the strongest and most robust at this location. The second strongest interpolation is Thanksgiving Valley. Contrary to the measurements at Bennett

234 Platform, Thanksgiving Valley has the highest salt concentrations in the center of the valley, with lower concentrations to

235 both the east and west. The lowest concentration contours are closest to the glacier for both Bennett Platform and

236 Thanksgiving Valley, which is likely related to glacial history since the soils near the glacier are relatively younger than

those further inland based on meteoric ¹⁰Be data (Diaz et al., 2020a). The interpolation from Roberts Massif does not have a
 distinguishable spatial trend.

239 The multiple linear regression and random forest models vary in their strength for the individual analytes. The 240 highest R² value from the linear regression is 0.655 for NaSr²⁺, while total water-soluble salts, water-soluble N:P ratio, 241 ClO_4 , and ClO_3 have values of 0.6137, 0.6037, 0.4440, and 0.5533, respectively (Table 2). The lowest R² value is for 242 CIPO₄³⁻ at 0.1705. The p-values for nearly all analytes are <<0.001, even with a Bonferroni Correction, with CI⁻ having the 243 only value above 0.05. The highest out-of-the-bag box explained variance values from the random forest models are for total 244 salts and KClO₃⁺⁺ and St²⁺ at 7562% and 63%, respectively for both analytes. The lowest explained variance is for Sr²⁺ at 245 37%. Values for NO₄⁻, PO₄³⁻, ClO₄⁻, and N:P ratio and ClO₄⁻ are negative52% and 48%, respectively. The explained variance 246 for total salts is 45% and the variance for ClO₃- is 43%. We also evaluated the most important and least important variables 247 in the random forest models based on node purity. The most important variable for the majority of analytes is elevation, 248 while distance from the glacier is most important for N:P ratio and latitude for ClO₃⁻ (Table 2). The least important variables 249 is are distance from the coast and latitude for every analyte, except ClO₃-and NH₂, for which distance from the glacier is 250 least important.

251 5. Discussion

252 5.1. Implications for ecological habitat suitability

253 By establishing geochemical zones for the Shackleton Glacier region, we can better understand the relationship 254 between geochemistry and geography, and ultimately biogeography. As stated in the introduction, we focused particularly on 255 total water-soluble salt concentrations, water-soluble N:P ratios, and ClO₄⁻ and ClO₃⁻ concentrations.

256 5.1.1. Elevation and moisture controls on total water-soluble salt gradients

257 The elevational trends of total salt concentrations at the Shackleton Glacier are similar to those previously described

in the TAM, where soils from higher elevation sites typically have higher salt concentrations (Bottos et al., 2020; Lyons et

- al., 2016; Magalhães et al., 2012). Our results are also consistent with those from Scarrow et al. (2014) who found that salt
- concentrations typically decreased with distance from the glacier in the Beardmore and Lennox King Glacier regions. Our

total water-soluble salt interpolation maps highlight the spatial variability in Shackleton Glacier region soils (Fig. 6). The most spatially variable location is Robert Massif, which does not appear to follow local elevational, latitudinal, and/or distance inland gradients. This heterogeneity is not necessarily due to currently active soil leaching, as the soil moisture values are not drastically different between the samples (Table S2). Though the variability in cation concentrations is likely due to weathering of tills, scree, and bedrock (Claridge and Campbell, 1968), recent work on the isotopic composition of water-soluble nitrate and sulfate, the major anions in the upper zone, suggests a common, atmospheric source (Diaz et al., 2020b).

268 We argue that the heterogeneity in the total salt concentrations at Roberts Massif (Figs. 2; 6) is probably related to 269 different and complex wetting history, where seasonal snow patch melt may pool in local depressions, transporting water-270 soluble salts from slightly higher elevations and/or from saline wet-patches (Levy et al., 2012). This is demonstrated on a 271 larger scale at Thanksgiving Valley, a glacially carved valley, where the higher concentrations of salts in the center of the 272 valley are likely due to the transport of salts from nearby higher elevation slopes during melting events. This is further 273 evidenced by the presence of two small, closed-basin ponds in the center of the valley, which likely formed from glacial melt 274 and may have been larger in size in the recent past (Diaz et al., 2019). Similarly, streams and meltwater tracks in the MDV 275 leach soils and carry salts into closed basin, brackish to hyper-saline lakes, where salts are cryoconcentrated over time 276 (Lyons et al., 1998). Our results suggest that elevation and wetting history are important contributors to total salt gradients in 277 the Shackleton Glacier region, as they influence the accumulation of salts and subsequent leaching from soils.

278 5.1.2. Influence of glacial history on water-soluble N:P ratios

279 Stoichiometric dependencies have been identified for Antarctic terrestrial organisms, where nutrient concentrations, 280 in addition to soil aridity, limit ecosystem development (Nkem et al., 2006). Since nitrate is primarily derived from 281 atmospheric deposition and phosphorus is primarily liberated from minerals by chemical weathering in the CTAM, many 282 inland and higher elevation soils have accumulated high concentrations of NO3-, resulting in stoichiometric imbalance with 283 soluble PO4³⁻ (Ball et al., 2018; Barrett et al., 2007; Diaz et al., 2020b; Lyons et al., 2016; Nkem et al., 2006). As in the 284 MDV, younger and coastal soils at lower elevations in the Shackleton Glacier region have the lowest water-soluble N:P 285 ratios, driven by relatively low concentrations of NO₃⁻ and high concentrations of PO₄³⁻ due to an increase in moisture 286 content and chemical weathering (Heindel et al., 2017) (Fig. 2; 4). It is not surprising that life was conspicuous in these soils, 287 with thick lichen growth on several rocks and the presence of both Collembola and mites at Mt. Speed and Mt. Wasko (Fig. 288 S1). However, despite overall elevational and latitudinal gradients, some inland locations in the middle and upper zones have 289 water-soluble N:P ratios near those from the lower zone (Fig. 2).

290 The interpolation model from Bennett Platform shows that some locations near the glacier have lower total water-291 soluble salt concentrations (Fig. 6), similar to soils surveyed in the MDV (Bockheim, 2002). However, the samples near the 292 glacier at Bennett Platform not only have lower total salt concentrations, they also have lower N:P ratios than samples 293 collected further inland. This is also the case for the middle zone locations (Fig. 2f). We argue this is due to differences in 294 glacial history between the locations. Our previous work showed that soils near the glacier are younger than soils further 295 inland in the Shackleton Glacier region (Diaz et al., 2020a). These soils are shielded from nitrate accumulation during glacial 296 periods, and the recently exposed rocks likely serve as fresh mineral weathering material for PO43- mobilization (Heindel et 297 al., 2017). Recently exposed and relatively nutrient rich soils might be important refugia for invertebrates. Previous 298 hypotheses have suggested that organisms may have persisted at higher elevations during glacial periods (Bennett et al., 299 2016; Stevens and Hogg, 2003). However, abiotic gradients in the Beardmore Glacier region suggest that higher elevation 300 soils have salt concentrations that would classify them as unsuitable habitats (Lyons et al., 2016). If few organisms survived 301 glaciations, the near-glacier, relatively P-rich soils may be important in helping communities recover and restructure post-

302 glaciation.

303 5.1.3. High and variable ClO₄⁻ and ClO₃⁻ concentrations

304 Our ClO4- and ClO3- concentrations include the highest measured in Antarctica to date and are comparable to 305 concentrations from the Atacama and Mojave Deserts (Jackson et al., 2015). Though not a strong correlation, the highest 306 elevation samples (upper zone) have the highest ClO₄⁻ and ClO₃⁻ concentrations (Fig. 2g; 2j). Similar to NO₃⁻, ClO₄⁻ and 307 ClO₃⁻ are derived from atmospheric deposition and because of their high solubilities, their accumulations are related to 308 wetting and glacial histories (Jackson et al., 2016, 2015). Therefore, soils which have been exposed for long periods of time 309 and have not experienced snow or ice melt, such as those from Schroeder Hill and Roberts Massif, are able to accumulate 310 high concentrations of ClO_4^- and ClO_3^- . Interestingly, our ClO_4^- concentrations are lower (maximum of ~1.9 g L⁻¹) than the 311 highest recorded tolerance (1.1M (~130 g L⁻¹) NaClO₄) for the extremotolerant bacteria *Planococcus halocryophilus*, yet a 312 recent study shows no detectable biomass for Schroeder Hill samples (Dragone et al., 2020). (Per)chlorates are strong 313 oxidizers and are well established as toxic, thus the concentrations of ClO₄⁻ and ClO₃⁻ might be additional, crucial indicators 314 of habitat suitability. However, the concentrations are highly heterogenous across our sampled locations (Fig. 2k-l), and 315 unlike ClO3, neither the multiple linear regression nor random forest models were able to adequately capture the variability 316 in ClO₄⁻ concentrations (Table 2).

317 5.2. Machine learning as a tool to predict soil geochemical trends

We sought to evaluate our multiple linear regression and random forest models using a testing dataset from the Shackleton Glacier region (n = 31) and a second dataset from the Darwin Mountains (~80°S) (n = 10) (Magalhães et al., 2012). Few published/available TAM dataset include sample GPS coordinates, soil moisture, and water-soluble ion geochemistry. As stated in Section 3.3, the Shackleton Glacier region test data were not included in the random forest model generation so we could evaluate our models with an independent dataset. For the Darwin dataset, distance from the glacier, distance from the coast, and elevation were determined using the Reference Elevation Model of Antarctica (REMA), while location, soil moisture and geochemistry were retrieved from the literature (Howat et al., 2019; Magalhães et al., 2012). We evaluated all 15 analytes from the original models with the Shackleton dataset and, due to a lack of data, only evaluated 7
 analytes from the Darwin soils (Figure 7).

327 Both the multiple linear regression and random forest model outputs are moderately well-correlated for the 328 Shackleton dataset, as determined by Pearson correlations between the measured and predicted values (Fig. 7a; Table 3). The 329 random forest models outperform the linear regression models for nearly every analyte, with the notable exceptions of FSr²⁺⁻, 330 Na^+ , NH₃, and PNO₃₄³⁻, and nearly all p-values are <0.001. For Cl⁻, in particular, the random forest model significantly 331 outperforms the multiple linear regression model, with R^2 values of 0.67 and 0.16, respectively. $Mg^{2+}N$: P molar ratio is the 332 most accurately predicted analyte, with R² values of 0.8879 and 0.5952 for the random forest and linear regression models, 333 respectively. However, the highest R² value for the multiple linear regression model is for Na⁺ at 0.64 (Fig. 7a Table 3). In 334 terms of our analytes of interest regarding habitat suitability, total salts have the second strongest correlation (following N:P 335 ratio) in-with the random forest model ($R^2 = 0.8151$), followed by water soluble N:P ratio ($R^2 = 0.42$), ClO₄⁻ ($R^2 = 0.4078$), 336 and ClO₃⁻ ($R^2 = 0.7428$). N:P ratio in particular performs significantly better than the linear regression model ($R^2 = 0.05$). 337 Mean absolute error (MAE) and root mean squared error (RMSE) values indicate that the random forest models also have a 338 smaller error compared to the multiple linear regression models (Table 4). MAE values are lower than RMSE values for both 339 models, indicating the strong presence influence of outliers in the testing dataset. This is unsurprising as the standard 340 deviation and coefficient of variation values for the entire dataset are relatively large for all analytes. Additionally, the strong 341 presence of outliers is are likely one reason why the random forest models are stronger than the multiple linear regression 342 models.

343 Similar to the model performance in the Shackleton Glacier region, the water-soluble ion predictions for the Darwin 344 Glacier region are more strongly correlated with measured values in the random forest models compared to the multiple **3**45 linear regressions (Fig. 7b). In fact, the linear regression models fail for nearly all the Darwin samples and all-most 346 concentration outputs are negative, which is likely due to overfitting during model generation. Here, Ca²⁺ and K⁺ are 347 exceptions and the multiple linear regression models outperform the random forest models in both cases. MAE and RSME 348 values for both models are much higher than those for the Shackleton dataset (Table 4). On the other hand, the random forest **3**49 models perform particularly well for some analytes. Though a small sample size, the R² values for Mg²⁺N:P molar ratio and 350 $\text{KCa}^{+2\pm}$ are 0.68 and 0.66, respectively 87, with p-values << 0.001. Total salts is moderately correlated ($R^2 = 0.474$) and N:P 351 ratio has an R² value of 0.01, indicating poor model performance. It is unclear why Mg²⁺ and K⁺ some analytes, such as N:P 352 molar ratio, are the most accurately predicted, though we suspect that this is due to 1) weathering trends of local lithology 353 across the TAM₇ since chemical weathering is probably the major source of these ions, and 2) deposition and accumulation 354 of atmospherically-derived ions at higher elevations (Diaz et al., 2020b).

355 It should be noted that the R² values simply measure the strength of the correlations between the measured and 356 predicted values. We performed slope tests by fitting bivariate lines using the standardized major axis (SMA) to further 357 understand the relationship between the two values using the smatr library in R (Warton et al., 2012). For this test, we 358 specifically evaluated the null hypothesis (H_0) where slope = 1, which would indicate whether an ideal, direct 1:1 359 relationship exists between the measured and predicted values. Test statistic values (t) were used to measure the sample 360 correlation between the residuals and fitted values (Warton et al., 2012). Test statistic values near 1 indicate that we reject 361 the null hypothesis. In other words, higher absolute test statistic values indicate a slope other than 1. Of the 15 analytes in the 362 Shackleton dataset, 57 analytes have slopes near 1 for the multiple linear regression models and 116 for the random forest, as 363 indicated by test statistic values less than 0.5. For the Darwin, only one analyte, NO3, has no analytes have a test statistic 364 values less than 0.5 (Fig. 7; Table 3).

365 These data indicate that while some analytes have high correlations between measured and predicted values, the 366 models perform best with the Shackleton Glacier region soils. Additionally, though the relationship may not be 1:1, the 367 random forest models are effective at predicting the measured geochemical gradients. For example, similar to our data, the 368 Darwin Glacier samples generally have greater water-soluble N:P ratios and total water-soluble salt concentrations further 369 from the glacier and at higher elevations (Magalhães et al., 2012), a trend that is reflected by our model results despite offset 370 values. Additionally, corrections for the offset of the model from a slope = 1 (i.e. multiplying the model output value by the 371 regression slope) can be made to better estimate specific concentrations, though the difference between modeled and 372 measured values can still be up to 2x greater. Our sample size for building the multiple linear regression and random forest 373 models is small. We anticipate that, as more data are collected throughout the CTAM, these data can be added to the model 374 training dataset, expanding our prediction capabilities and increasing model reliability.

375 6. Conclusions

376 The soil ecosystems found in the Transantarctic Mountains are among the least diverse on Earth and their structure **3**77 is influenced by environmental factors variables. We characterized environmental and geochemical gradients in the 378 Shackleton Glacier region, which aid in our understanding of the abiotic properties in soils governing biodiversity and 379 biogeography. The 220 samples we analyzed represent a wide range of soil environments: those with different elevation, 380 latitude, longitude, glacial history, and geochemistry. We determined three soil zones: an upper zone near the head of the 381 glacier which is characterized by high total water-soluble salt concentrations, high water-soluble N:P ratios, and high ClO₄-382 and ClO3⁻ concentrations, a lower zone with low total salt concentrations and higher PO4³⁻ concentrations, and a middle zone 383 with intermediate values. The zones help elucidate the geographic influences on soil geochemistry. In addition, our total 384 water-soluble salt interpolations at Roberts Massif, Bennett Platform, and Thanksgiving Valley reflect the local small-scale 385 variability of salt concentrations and possible influences from soil age and wetting history.

386 Similar to previous studies, our results suggest that high elevation and inland soils, such as those from the upper 387 zone, were likely unsuitable candidates for refugia during the Last Glacial Maximum. However, glacial advance and retreat 388 and climate shifts may leach soils, lowering otherwise toxic total water-soluble salt concentrations and N:P ratios. These 389 more recently exposed soils may be particularly important in maintaining and reviving contemporary and past biological 390 communities.

391 Five geographic variables (latitude, longitude, elevation, distance from the coast, and distance from the glacier) and 392 soil moisture were correlated with soil geochemistry. We used these variables to develop multiple linear regression and 393 random forest models to predict ion concentrations and geochemical gradients. The model results generally reflected the 394 measured geochemical variability across the region. Test datasets from the Shackleton and Darwin Glacier regions showed 395 that the random forest models typically outperformed the multiple linear regression models when correlating measured and 396 predicted values, especially for the Darwin region. Though most correlations did not exhibit a 1:1 relationship and had 397 varying slopes, the random forest models were able to adequately predict geochemical gradients, as demonstrated by 398 moderate to high R² values between measured and model predicted concentrations. As terrestrial Antarctic geochemical 399 databases expand and are included in the random forest model training dataset, we anticipate the model's predictive 400 capabilities will expand and improve as well. While these results are currently most applicable for Central Transantarctic 401 Mountain soils, similar techniques can be applied to other hyper-arid environments (e.g. Namib and Atacama Deserts, Mars) 402 to inform patterns of biodiversity and biogeography.

403 **Author Contributions**

404 The project was designed and funded by BJA, DHW, IDH, NF, and WBL. Fieldwork was conducted by BJA, DHW, IDH, 405 NF, and MAD. CBG, SAW, and MAD prepared and analyzed the samples for water-soluble ion and nutrient analyses. WAJ 406 prepared and analyzed the samples for ClO₄⁻ and ClO₃⁻. MAD generated the scripts and performed the analyses for the IDW 407 interpolations, multiple linear regression, and random forest models, MAD wrote the article with contributions and edits 408 from all authors.

409 **Data Availability Statement**

410 The datasets generated for this study are included in the article or supplementary materials.

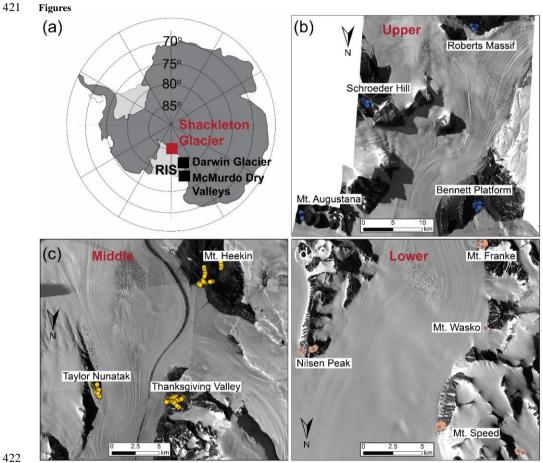
411 **Competing Interests**

412 The authors declare that they have no conflict of interest.

413 Acknowledgments

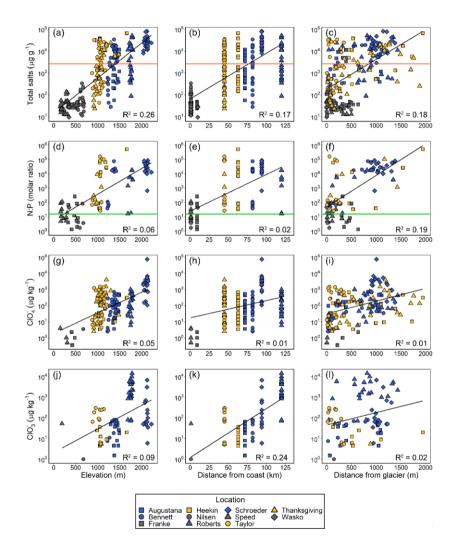
414 We thank the United States Antarctic Program (USAP), Antarctic Science Contractors (ASC), Petroleum Helicopters Inc.

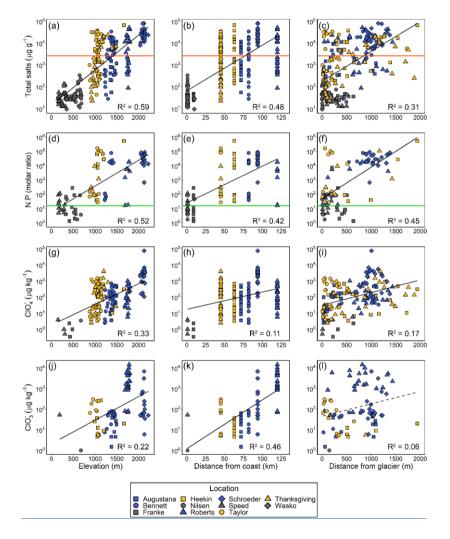
- 415 (PHI), and Marci Shaver-Adams for logistical and field support. Additionally, we thank Daniel Gilbert for help with initial
- 416 laboratory analyses at The Ohio State University. We are grateful to Dr. Peter Convey and Dr. Natasja van Gestel for
- 417 thoughtful reviews which strengthened our results and broader implications. This work was supported by NSF OPP grants 418
- 1341631 (WBL), 1341618 (DHW), 1341629 (NF), 1341736 (BJA), and NSF GRFP fellowship 60041697 (MAD).
- 419 Geospatial support for this work provided by the Polar Geospatial Center under NSF OPP grants 1043681 and 1559691.
- 420





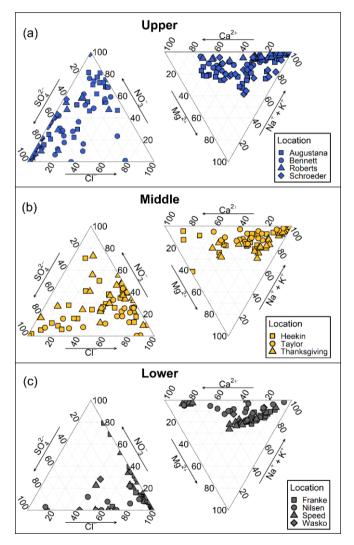
423 424 Figure 1. Samples were collected and analyzed from the exposed soils along the Shackleton Glacier, a major outlet glacier of the EAIS (a), in three zones. The upper zone (b) was located at the head of Shackleton Glacier, the middle zone (c) was the 425 central portion, and the lower zone (d) was at the mouth of the glacier where it drains into the Ross Sea. Satellite images 426 were provided courtesy of the Polar Geospatial Center (PGC).



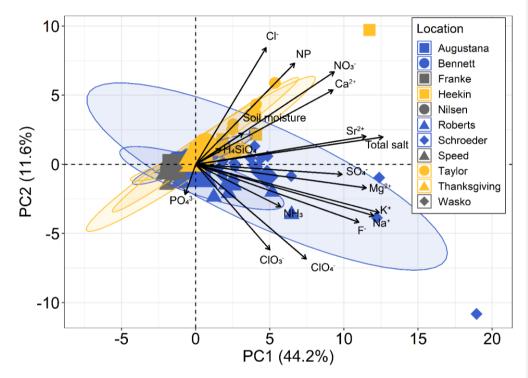




- Figure 2. Total water-soluble salts, water-soluble N:P molar ratio, and ClO₄⁻ and ClO₃⁻ concentrations (log scale) were
- compared to elevation, distance from the coast, and distance from the glacier for samples from the three geographic zones
- (blue for upper, yellow for middle, grey for lower zones). Linear regression lines are plotted, where dashed lines represent
- 430 431 432 433 434 435 436 regressions where p > 0.05 with a Bonferroni Correction, and R^2 values are reported for each relationship. The horizontal
- orange lines represents nematode salt tolerance of ~2,600 (Nkem et al., 2006) and the green lines represents the Redfield
- ratio, N:P = 16 for phytoplankton in the ocean.

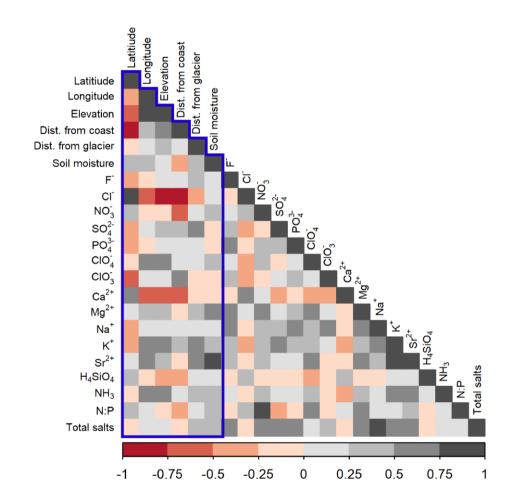


438 Figure 3. Anion and cation ternary diagrams for the three geographic zones.

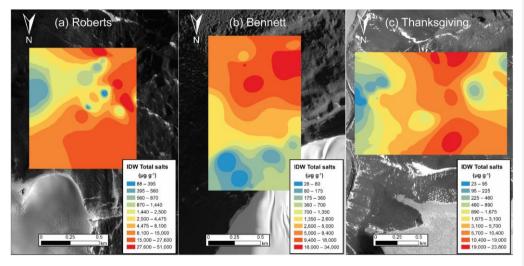


441 442 443 444 445 Figure 4. Principal component analysis (PCA) biplot generated in R using factoextra and built in R software libraries with all anions, cations, nutrients, and soil moisture for the three geographic zones. The PCA is based on the correlation matrix (i.e.

scale = TRUE). Principal component 1 and principal component 2 are plotted on the x and y axes, respectively. Shaded ellipses represent 95% confidence intervals.



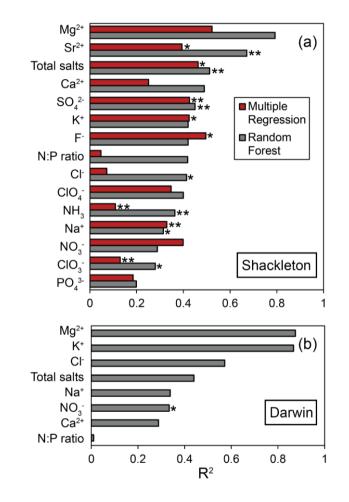
448 449 450 451 452 Figure 5. Spearman's rank correlation matrix generated in R using the corrplot library. The colors represent correlation coefficients, indicating the strength and magnitude of the correlation. The blue box indicates the geographic variables and soil moisture, which were variables used in the multiple linear regression and random forest models. Familywise type 1 error corrections were not applied for this analysis.





454 Figure 6. Inverse distance weighted (IDW) interpolations of total salt concentration for Roberts Massif (a), Bennett Platform
 455 (b), and Thanksgiving Valley (c). The color scale represents the 10 natural breaks in the data. Interpolations were created and

456 mapped using the Geostatistical Analyst tool in ArcMap 10.3.



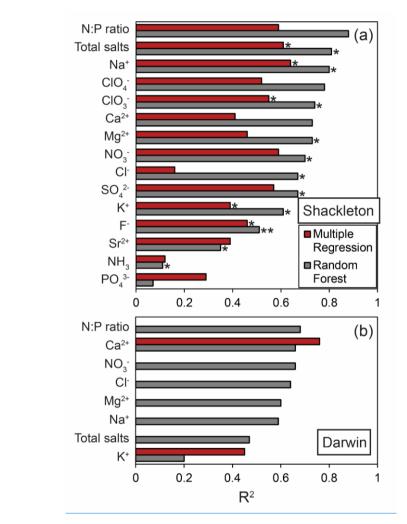


Figure 7. \mathbb{R}^2 values for the multiple linear regression and random forest model predicted and measured values for the different analytes (Table 3). Test datasets include the Shackleton Glacier region (n = 31) and the Darwin Glacier region (n = 10) (Magalhães et al., 2012). Analytes with slopes near 1, indicating good agreement between measured and predicted values, are indicated (* t < 0.5; ** t < 0.20).

•	Max	Min	Mean	STD	CV
Elevation (m)	2,220	150	1,130	551	48
Distance from coast (km)	120	1	55	38	68
Distance from glacier (m)	1,940	1	519	472	90
Soil moisture (%)	12.3	0.1	2.1	2.1	102
$F^{-}(\mu g g^{-1})$	120	0.39	8.87	11.78	133
$Cl^{-}(\mu g g^{-1})$	13,600	1.59	615	1,780	289
NO_{3}^{-} (µg g ⁻¹)	38,400	0.10	1,470	3,450	235
SO_4^{2-} (µg g ⁻¹)	55,300	0.08	4,390	8,080	184
PO_4^{3-} (µg kg ⁻¹)	4,200	76.09	381	560	147
ClO ₄ ⁻ (µg kg ⁻¹)	75,000	0.35	985	6,020	611
ClO ₃ ⁻ (µg kg ⁻¹)	14,500	1.00	1,170	2,500	214
Ca ²⁺ (µg g ⁻¹)	4,400	0.55	839	1,160	139
Mg^{2+} (µg g ⁻¹)	6,280	0.12	293	705	240
$Na^{+} (\mu g g^{-1})$	25,300	0.39	1,140	2,880	252
$K^{+}(\mu g g^{-1})$	440	0.86	28.31	51.61	182
$Sr^{2+}(\mu g g^{-1})$	46.61	0.01	8.63	10.31	119
H ₄ SiO ₄ (µg g ⁻¹)	60.78	1.14	21.78	11.03	50.67
NH ₃ (µg kg ⁻¹)	5,080	18.85	324	587	181
N:P ratio (molar)	526,000	0.29	23,600	62,700	266

9.46

7,932

13,300

167

80,500

STD, standard deviation; CV, coefficient of variation

465 Table 1. Overview of geography, soil moisture, and water-soluble ions from the Shackleton Glacier region. The minimum values reported are those within the detection limits. Individual sample concentrations are detailed in Table S2.

466

Total salt (µg g⁻¹)

468	Table 2. Out-of-the-bag-box multiple linear regression and random forest model statistics generated in R. All geochemical
469	data were log-transformed.

	Multip	le regression		Random forest	
	\mathbf{R}^2	p-value	Variance explained (%)	Most important variable	Least important variable
F	0. 27<u>47</u>	<<0.001	<u>57</u> 36	Elevation	Distance from coast
Cl	0. 05<u>19</u>	<u><<0.001</u> 0.082	20<u>60</u>	Elevation	Distance from coastLongitude
NO ₃ ⁻	0. 18<u>52</u>	<<0.001	-4 <u>60</u>	Distance from glacierElevation	Distance from coastLongitude
SO4 ²⁻	0. 37<u>53</u>	<<0.001	44 <u>62</u>	Elevation	Distance from coastLongitude
PO4 ³⁻	0.1 <u>7</u> 6	<<0.0010.017	<u>-74</u>	LatitudeElevation	Distance from coast
ClO ₄ -	0. <u>44</u> 1	<u><<0.001</u> 0.010	-3<u>48</u>	Elevation	Distance from coast
ClO ₃ -	0. <u>55</u> 33	<<0.001	4 <u>3</u> 63	Latitude	Distance from glacier
Ca ²⁺	0. <u>2644</u>	<<0.001	4 <u>660</u>	Soil moistureElevation	Distance from coast
Mg ²⁺	0. 29<u>49</u>	<<0.001	<u>2261</u>	Elevation	Distance from coastLongitude
Na ⁺	0. <u>2165</u>	<<0.001	<u>3875</u>	Elevation	Distance from coastLongitude
\mathbf{K}^+	0.4<u>0.48</u>	<<0.001	62 60	Elevation	Distance from coast
Sr ²⁺	0. 55 34	<<0.001	62 <u>37</u>	Elevation	Distance from coast
NH ₃	0.29	<<0.001	54<u>38</u>	Elevation	Distance from glaciercoast
N:P	0. 37<u>60</u>	<<0.001	- <u>352</u>	Distance from glacier	Distance from coastLongitude
Total salts	0. <u>3761</u>	<<0.001	4 5 75	Elevation	Distance from coastLongitude

Table 3. Multiple linear regression and random forest statistics between predicted and measured concentrations from the

472 473 474 Shackleton and Darwin Glacier regions. R² and p-values are reported for the correlations between measured and predicted

concentrations. Regression slopes and test statistic values (t) were calculated using the smatr library (Warton et al., 2012) in R to evaluate the null hypothesis (H_0) of slope = 1. Higher test statistic values (closer to one) indicate that we reject the null

76	hypothesis.	All	geochemical	data	were	log-transformed.	

		Multiple Linea	ar Regress	ion	n Random Forest				
Analyte	R ²	p-value	Reg. slope	Test statistic (t) for H_0 slope = 1	R ²	p-value	Reg. slope	Test statistic (t) for H ₀ slope = 1	
		I		Shackleton		I		1	
N:P	0.59 0.5	<0.001 << 0.001	0.58 0.5	-0.720-0.711	0.88 0.7	<<0.001 <<0.00	0.64 0.5	-0.792 0.780	
ratioMg ²	2		2	0.720 0.711	9	4	8	0.7720.700	
Total	0.610.3	<<0.001<0.001	0.711.2	Ţ.	0.810.6	<<0.001<<<0.00	0.86 0.9	-0.324*-	
salts Sr²⁺	9		2	0.483*0.247*	7	4	4	0.166**	
Na ⁺ Total	0.640.4	<<0.001 <<0.00	0.760.7	-0.424*-	0.800.5	<<0.001 <<0.00	0.890.9	-0.262*-	
salts	6	- +	6	0.343*	- +	- +	3	0.107**	
<u>ClO4</u>	<u>0.520.2</u>	<0.0010.004	0.600.4	<u>-0.614</u> -0.747	<u>0.78</u> 0.4	<u><<0.001</u> <<0.00	<u>0.710.6</u>	<u>-0.590</u> -0.586	
Ca ²⁺	5		2		9	+	+		
<u>ClO₃=</u>	<u>0.55</u> 0.4	<u>0.009</u> <<0.001	<u>0.72</u> 1.0	=	0.74 0.4	<u><0.001</u> <<0.001	0.861.1	Ē	
SO 4 ²⁻	3		7	<u>0.454*0.093*</u> *	5		θ	0.284*0.130* *	
$Mg^{2+}K^+$	0.460.4	<<0.001<<<0.00	<u>0.63</u> 1.5	<u>-0.550</u> 0.504*	<u>0.730.4</u>	<<0.001<<<0.00	<u>0.76</u> 1.7	<u>-0.469*0.629</u>	
	2	- +	4		2	+	9		
$Ca^{2+}F$	<u>0.410.5</u>	<u><0.001</u> <<0.001	<u>0.57</u> 1.2	<u>-0.613</u> 0.267*	<u>0.73</u> 0.4	<u><<0.001</u> <0.001	<u>0.741.7</u>	<u>-0.512</u> 0.617	
	0		2		2		8		
<u>NO3⁻N:P</u>	<u>0.59</u> 0.0	<<0.001 <u>0.241</u>	0.620.5	<u>-0.615-0.517</u>	0.700.4	<u><<0.001</u> <<0.00	<u>0.75</u> 0.3	-0.465*-	
ratio	5		9		2	4	5	0.867	
Sr ²⁺ Cl	<u>0.35</u> 0.0	0.0260.144	0.540.2	<u>-0.631-0.867</u>	<u>0.67</u> 0.4	<u><0.001</u> <<0.001	<u>0.82</u> 0.7	<u>-0.326*</u> -	
	7		8		1		0	0.424*	
<u>SO4</u> 2-	<u>0.57</u> 0.3	<u><<0.001</u> <0.001	0.632.0	<u>-0.584</u> 0.685	<u>0.67</u> 0.4	<u><<0.001</u> <0.001	<u>0.83</u> 3.4	<u>-0.310*</u> 0.897	
ClO ₄	5		+		θ		θ		
<u>Cl⁻NH₃</u>	0.160.1	0.0280.070	0.381.0	Ē.	<u>0.67</u> 0.3	<u><<0.001</u> <0.001	<u>0.76</u> 1.0	<u> </u>	
	+		4	<u>0.773</u> 0.037**	6		9	0.428*0.106* *	
<u>K⁺Na⁺</u>	<u>0.39</u> 0.3	<u><0.001</u> <0.001	<u>0.69</u> 0.9	<u>-0.429*-</u>	<u>0.610.3</u>	<u><<0.001</u> 0.001	<u>0.83</u> 1.5	E .	
	3		1	0.112**	1		4	<u>0.291*0.473*</u>	
<u>F'NO3</u>	0.460.4	<u><0.001</u> <0.001	<u>0.76</u> 0.4	-0.352*-	<u>0.510.2</u>	<u><<0.001</u> 0.002	<u>0.910.5</u>	-0.141**-	
	θ		7	0.725	9		6	0.594	
<u>NH₃ClO</u>	<u>0.12</u> 0.1	0.0520.043	<u>0.57</u> 1.2	Ē.	0.110.2	0.0680.002	0.610.7	-0.475*-	
3	3		θ	<u>0.528</u> 0.197**	8		4	0.382*	
$\underline{PO_4^{3-}}$	0.290.1	<u>0.070</u> 0.016	0.320.5	<u>-0.857-0.645</u>	0.070.2	0.4080.022	0.380.1	<u>-0.764-0.967</u>	
PO4 ³⁻	8		0	Darwin	θ		5		
NLD.				Dai will	0.690.9	0.021 < <0.001	0.54 0.3	-0.765 -0.948	
<u>N:P</u> ratio Mg²	<u> </u>	F	<u> </u>	<u> </u>	<u>0.68</u> 0.8 7	<u>0.021</u> <<0.001	<u>0.540.3</u> 9	<u>-0.703-0.948</u>	
-					F		~	1	

Formatted
Formatted
Formatted
Formatted Table
Formatted
Formatted Formatted
Formatted
Formatted
Formatted
Formatted
Formatted
Formatted
Formatted
Formatted
Formatted
Formatted
Formatted
Formatted
Formatted
Formatted
Formatted
Formatted
Formatted
Formatted
Formatted
Formatted
Formatted
Formatted
Formatted
Formatted
Formatted
Formatted

$Ca^{2+}K^+$	0.76-	0.001-	0.66-	-0.645-	0.660.8	0.004 << 0.001	0.500.4	-0.785-0.89
					7		9	
NO ₃ -Cl	<u></u>	F	<u></u>	<i>E</i>	0.660.5	0.0040.011	0.490.1	-0.794-0.98
				-	7		3	
<u>Cl⁻Total</u>	—		_	F	0.640.4	0.0050.001	0.223.2	<u>-0.962</u> 0.940
salts					4		5	
Mg ²⁺ Na ⁺		—	-	F	0.600.3	<u>0.1400.078</u>	<u>0.670.2</u>	-0.544-0.93
					4		3	
Na ⁺ NO₃⁻	F	F	F	F	<u>0.59</u> 0.3	<u>0.160</u> 0.080	0.420.6	-0.836-
					3		5	0.476*
Total	_	F	F	F	<u>0.470.2</u>	<u>0.028</u> 0.110	0.420.1	<u>-0.802</u> -0.96
saltsCa2+				_	9		7	
<u>K⁺N:P</u>	0.45-	0.070-	0.62-	-0.550-	<u>0.20</u> 0.0	<u>0.2670.765</u>	0.288.0	<u>-0.882</u> 0.97
ratio					4		4	

	Formatted
	Formatted
	Formatted
$\overline{}$	Formatted
$\langle \rangle$	Formatted
())	Formatted
///	Formatted
	Formatted Formatted
	Formatted
	Formatted
	Formatted
	Formatted
	Formatted
	Formatted
	Formatted
	Formatted
	Formatted
Y	Formatted
	Formatted
V	Formatted
	Formatted

478 479 Table 4. Multiple linear regression and random forest model mean absolute error (MAE) and root mean squared error (RMSE). All geochemical data were log-transformed for the analysis.

	Multiple Linear Regression nalyte MAE		Randor	n Forest				
Analyte			MAE	RMSE				
Shackleton								
<u>N:P ratio</u> Mg ²⁺	<u>2.19</u> 300	<u>2.73</u> 461	<u>1.75</u> 204	<u>2.11</u> 347				
Total saltsSr ²⁺	<u>1.45</u> 3.74	<u>1.69</u> 4 .96	<u>0.86</u> 1.83	<u>1.17</u> 2.90				
<u>Na⁺Total salts</u>	<u>1.23</u> 5,640	<u>1.52</u> 7,070	<u>0.83</u> 4,400	<u>1.13</u> 7,030				
<u>ClO4</u> -Ca ²⁺	<u>1.33</u> 797	<u>1.62</u> 1,100	<u>0.91</u> 554	<u>1.12</u> 912				
<u>ClO3</u> - SO 4 ²⁻	<u>1.07</u> 3,310	<u>1.67</u> 3,890	<u>1.01</u> 2,200	<u>1.26</u> 3,780				
$\underline{Mg^{2+}}K^{+}$	<u>1.78</u> 15.86	<u>2.08</u> 21.16	<u>1.07</u> 13.48	<u>1.48</u> 25.61				
<u>Ca²⁺</u> F	<u>1.84</u> 3.14	<u>2.21</u> 4 .19	<u>1.18</u> 3.13	<u>1.53</u> 6.31				
<u>NO3-N:P ratio</u>	<u>1.96</u> 39,700	<u>2.29</u> 59,300	<u>1.56</u> 7,310	<u>1.93</u> 17,210				
<u>Sr²⁺</u> Cl ⁻	<u>1.05</u> 936	<u>1.17</u> 1,540	<u>0.59</u> 658	<u>0.82</u> 1,240				
<u>SO4</u> ²⁻ ClO4 ⁻	<u>1.58</u> 1,180	<u>1.94</u> 1,560	<u>1.35</u> 875	<u>1.67</u> 2,960				
<u>Cl-</u> NH ₃	<u>2.11</u> 214	<u>2.39</u> 301	<u>1.07</u> 158	<u>1.5</u> 244				
<u>K</u> ⁺ Na ⁺	<u>0.73</u> 883	<u>0.89</u> 1,170	<u>0.56</u> 918	<u>0.72</u> 1,730				
<u>F</u> NO ₃ -	<u>0.48</u> 1,200	<u>0.6</u> 1,910	<u>0.46</u> 1,130	<u>0.58</u> 2,040				
<u>NH₃ClO₃-</u>	<u>0.67</u> 1,110	<u>0.83</u> 1,630	<u>0.65</u> 343	<u>0.86</u> 1,050				
<u>PO₄³-PO₄³⁻</u>	<u>0.75</u> 428	<u>0.96</u> 690	<u>0.78</u> 261	<u>1.14</u> 742				
		Darwin						
<u>N:P ratio</u> Mg ²⁺	<u>2676,300</u>	<u>2676,320</u>	<u>1.48</u> 302	<u>1.73</u> 475				
<u>Ca²⁺</u> K ⁺	<u>5.79</u> 1,060	<u>5.85</u> 1,060	<u>2.63</u> 13.33	<u>2.83</u> 15.84				
NO3-Cl-	<u>261</u> 206,000	<u>261</u> 206,000	<u>3.19</u> 2,140	<u>3.52</u> 3,330				
Cl ⁻ Total salts	<u>372</u> 215,000	<u>372</u> 215,000	<u>2.99</u> 5,540	<u>3.32</u> 7,590				
Mg ²⁺ Na ⁺	<u>460</u> 8,330	<u>460</u> 8,530	<u>2.89</u> 1,500	<u>3.06</u> 2,600				
<u>Na⁺NO₃-</u>	<u>245</u> 128,000	<u>245</u> 128,000	<u>2.57</u> 3,260	<u>2.88</u> 4,870				
Total saltsCa2+	<u>13970,300</u>	<u>139</u> 70,300	<u>1.22</u> 1,410	<u>1.67</u> 2,070				
<u>K⁺N:P ratio</u>	<u>30.8</u> 18,100,000	<u>30.8</u> 18,100,000	<u>1.00</u> 18,700	<u>1.19</u> 46,900				
IAE, mean absolu	te error; RMSE, root 1	nean squared error						

	Formatted	Table
--	-----------	-------

Formatted Table

481 References

- 482 Antipov, E. A. and Pokryshevskaya, E. B.: Mass appraisal of residential apartments: An application of Random forest for
- 483 valuation and a CART-based approach for model diagnostics, Expert Syst. Appl., 39(2), 1772–1778,
- 484 doi:10.1016/j.eswa.2011.08.077, 2012.
- Ball, B. A., Adams, B. J., Barrett, J. E., Wall, D. H. and Virginia, R. A.: Soil biological responses to C, N and P fertilization
 in a polar desert of Antarctica, Soil Biol. Biochem., 122, 7–18, doi:10.1016/J.SOILBIO.2018.03.025, 2018.
- Balter-Kennedy, A., Bromley, G., Balco, G., Thomas, H. and Jackson, M. S.: A 14.5-million-year record of East Antarctic
 Ice Sheet fluctuations from the central Transantarctic Mountains, constrained with cosmogenic 3He, 10Be, 21Ne, and 26Al,
 Cryosph., 14(8), 2647–2672, doi:10.5194/tc-2020-57, 2020.
- Barrett, J. E., Virginia, R. A., Wall, D. H., Cary, S. C., Adams, B. J., Hacker, A. L. and Aislabie, J. M.: Co-variation in soil
 biodiversity and biogeochemistry in northern and southern Victoria Land, Antarctica, Antarct. Sci., 18(4), 535–548,
 doi:10.1017/S0954102006000587, 2006.
- Barrett, J. E., Virginia, R. A., Lyons, W. B., McKnight, D. M., Priscu, J. C., Doran, P. T., Fountain, A. G., Wall, D. H. and
 Moorhead, D. L.: Biogeochemical stoichiometry of Antarctic Dry Valley ecosystems, J. Geophys. Res., 112(G1), G01010,
 doi:10.1029/2005JG000141, 2007.
- Beet, C. R., Hogg, I. D., Collins, G. E., Cowan, D. A., Wall, D. H., Adams, B. J., Beet, C., Hogg, I., Collins, G., Cowan, D.
 and Adams, B.: Genetic diversity among populations of Antarctic springtails (Collembola) within the Mackay Glacier
 ecotone 1, Genome, 59, 762–770, doi:10.1139/gen-2015-0194, 2016.
- Bennett, K. R., Hogg, I. D., Adams, B. J. and Hebert, P. D. N.: High levels of intraspecific genetic divergences revealed for
 Antarctic springtails: evidence for small-scale isolation during Pleistocene glaciation, Biol. J. Linn. Soc., 119(1), 166–178,
 doi:10.1111/bij.12796, 2016.
- Bockheim, J. G.: Landform and Soil Development in the McMurdo Dry Valleys, Antarctica: A Regional Synthesis, Arctic,
 Antarct. Alp. Res., 34(3), 308–317, doi:10.1080/15230430.2002.12003499, 2002.
- Bockheim, J. G.: Functional diversity of soils along environmental gradients in the Ross Sea region, Antarctica, Geoderma,
 144(1–2), 32–42, doi:10.1016/j.geoderma.2007.10.014, 2008.
- 506 Bottos, E. M., Laughlin, D. C., Herbold, C. W., Lee, C. K., McDonald, I. R. and Cary, S. C.: Abiotic factors influence
- patterns of bacterial diversity and community composition in the Dry Valleys of Antarctica, FEMS Microbiol. Ecol., 96(5),
 42, doi:10.1093/femsec/fiaa042, 2020.
- 509 Breiman, L.: Random forests, Mach. Learn., 45(1), 5–32, doi:10.1023/A:1010933404324, 2001.
- Caruso, T., Hogg, I. D., Nielsen, U. N., Bottos, E. M., Lee, C. K., Hopkins, D. W., Cary, S. C., Barrett, J. E., Green, T. G.
 A., Storey, B. C., Wall, D. H. and Adams, B. J.: Nematodes in a polar desert reveal the relative role of biotic interactions in
- 512 the coexistence of soil animals, Commun. Biol., 2(63), doi:10.1038/s42003-018-0260-y, 2019.
- 513 Claridge, G. G. C. and Campbell, I. B.: Soils of the Shackleton glacier region, Queen Maud Range, Antarctica, New Zeal. J.
 514 Sci., 11(2), 171–218, 1968.
- 515 Collins, G. E., Hogg, I. D., Convey, P., Barnes, A. D. and McDonald, I. R.: Spatial and temporal scales matter when
- 516 assessing the species and genetic diversity of springtails (Collembola) in Antarctica, Front. Ecol. Evol., 7, 76,
- 517 doi:10.3389/fevo.2019.00076, 2019.
- 518 Collins, G. E., Hogg, I. D., Convey, P., Sancho, L. G., Cowan, D. A., Lyons, W. B., Adams, B. J., Wall, D. H. and Green, T.
- G. A.: Genetic diversity of soil invertebrates corroborates timing estimates for past collapses of the West Antarctic Ice Sheet,
 Proc. Natl. Acad. Sci. U. S. A., 117(36), 22293–22302, doi:10.1073/pnas.2007925117, 2020.

- Convey, P. and McInnes, S. J.: Exceptional tardigrade-dominated ecosystems in Ellsworth Land, Antarctica, Ecology, 86(2),
 519–527, doi:10.1890/04-0684, 2005.
- 523 Courtright, E. M., Wall, D. H. and Virginia, R. A.: Determining habitat suitability for soil invertebrates in an extreme 524 environment: The McMurdo Dry Valleys, Antarctica, Antarct. Sci., 13(1), 9–17, doi:10.1017/S0954102001000037, 2001.
- 525 Davidson, A. D., Hamilton, M. J., Boyer, A. G., Brown, J. H. and Ceballos, G.: Multiple ecological pathways to extinction in mammals., 2009.
- 527 Diaz, M. A., Adams, B. J., Welch, K. A., Welch, S. A., Opiyo, S. O., Khan, A. L., McKnight, D. M., Cary, S. C. and Lyons,
- W. B.: Aeolian Material Transport and Its Role in Landscape Connectivity in the McMurdo Dry Valleys, Antarctica, J.
 Geophys. Res. Earth Surf., 123, 3323–3337, doi:10.1029/2017JF004589, 2018.
- 530 Diaz, M. A., Corbett, L. B., Bierman, P. R., Adams, B. J., Wall, D. H., Hogg, I. D., Fierer, N. and Lyons, W. B.: Relative terrestrial exposure ages inferred from meteoric 10Be and NO3- concentrations in soils along the Shackleton Glacier, Antarctica, Earth Surf. Dyn., in review, doi:https://doi.org/10.5194/esurf-2020-50, 2020a.
- 533 Diaz, M. A., Li, J., Michalski, G., Darrah, T. H., Adams, B. J., Wall, D. H., Hogg, I. D., Fierer, N., Welch, S. A., Gardner, C.
 534 B. and Lyons, W. B.: Stable isotopes of nitrate, sulfate, and carbonate in soils from the Transantarctic Mountains, Antarctica:
 535 A record of atmospheric deposition and chemical weathering, Front. Earth Sci., 8(341), doi:10.3389/feart.2020.00341,
- 536 2020ь.
- Dragone, N. B., Diaz, M. A., Hogg, I., Lyons, W. B., Jackson, W. A., Wall, D. H., Adams, B. J. and Fierer, N.: Exploring the
 boundaries of microbial habitability in soil, bioRxiv, doi:https://doi.org/10.1101/2020.08.03.234583, 2020.
- Freckman, D. W. and Virginia, R. A.: Soil Biodiversity and Community Structure in the Mcmurdo Dry Valleys, Antarctica, in Ecosystem dynamics in a polar desert; the McMurdo dry valleys, Antarctica, edited by J. C. Priscu, pp. 323–335,
 American Geophysical Union (AGU)., 1998.
- 542 Golledge, N. R. and Levy, R. H.: Geometry and dynamics of an East Antarctic Ice Sheet outlet glacier, under past and 543 present climates, J. Geophys. Res., 116(F3), F03025, doi:10.1029/2011JF002028, 2011.
- Golledge, N. R., Fogwill, C. J., Mackintosh, A. N. and Buckley, K. M.: Dynamics of the last glacial maximum Antarctic ice sheet and its response to ocean forcing, Proc. Natl. Acad. Sci. U. S. A., 109(40), 16052–16056, doi:10.1073/pnas.1, 2012.
- 546 Golledge, N. R., Levy, R. H., McKay, R. M., Fogwill, C. J., White, D. A., Graham, A. G. C., Smith, J. A., Hillenbrand, C.
- 547 D., Licht, K. J., Denton, G. H., Ackert, R. P., Maas, S. M. and Hall, B. L.: Glaciology and geological signature of the Last
 548 Glacial Maximum Antarctic ice sheet, Quat. Sci. Rev., 78, 225–247, doi:10.1016/j.quascirev.2013.08.011, 2013.
- Heindel, R. C., Spickard, A. M. and Virginia, R. A.: Landscape-scale soil phosphorus variability in the McMurdo Dry
 Valleys, Antarct. Sci., 29(3), 252–263, doi:10.1017/S0954102016000742, 2017.
- Heung, B., Bulmer, C. E. and Schmidt, M. G.: Predictive soil parent material mapping at a regional-scale: A Random Forest approach, Geoderma, 214–215, 141–154, doi:10.1016/j.geoderma.2013.09.016, 2014.
- Hodgson, D. A., Convey, P., Verleyen, E., Vyverman, W., McInnes, S. J., Sands, C. J., Fernández-Carazo, R., Wilmotte, A.,
 De Wever, A., Peeters, K., Tavernier, I. and Willems, A.: The limnology and biology of the Dufek Massif, Transantarctic
- 555 Mountains 82° South, Polar Sci., 4(2), 197–214, doi:10.1016/j.polar.2010.04.003, 2010.
- Hogg, I. D. and Wall, D. H.: Polar deserts, in Life at Extremes: Environments, Organisms and Strategies for Survival, edited
 by E. M. Bell, pp. 176–195, CAB International., 2012.
- Howat, I. M., Porter, C., Smith, B. E., Noh, M.-J. and Morin, P.: The Reference Elevation Model of Antarctica, Cryosph.,
 13, 665–674, doi:10.5194/tc-13-665-2019, 2019.

- 560 Jackson, A., Davila, A. F., Böhlke, J. K., Sturchio, N. C., Sevanthi, R., Estrada, N., Brundrett, M., Lacelle, D., McKay, C. P.,
- 561 Poghosvan, A., Pollard, W. and Zacny, K.: Deposition, accumulation, and alteration of Cl-, NO3-, ClO4- and ClO3- salts in
- 562 a hyper-arid polar environment: Mass balance and isotopic constraints, Geochim. Cosmochim. Acta, 182, 197-215,
- 563 doi:10.1016/j.gca.2016.03.012, 2016.
- 564 Jackson, M. S., Hall, B. L. and Denton, G. H.: Asynchronous behavior of the Antarctic Ice Sheet and local glaciers during 565 and since Termination 1, Salmon Valley, Antarctica, Earth Planet. Sci. Lett., 482, 396-406, doi:10.1016/j.epsl.2017.11.038, 566 2018
- Jackson, W. A., Davila, A. F., Estrada, N., Lyons, W. B., Coates, J. D. and Priscu, J. C.: Perchlorate and chlorate 567
- 568 biogeochemistry in ice-covered lakes of the McMurdo Dry Valleys, Antarctica, Geochim. Cosmochim. Acta, 98, 19-30, 569 doi:10.1016/j.gca.2012.09.014, 2012.
- Jackson, W. A., Böhlke, J. K., Andraski, B. J., Fahlquist, L., Bexfield, L., Eckardt, F. D., Gates, J. B., Davila, A. F., McKay, 570
- 571 C. P., Rao, B., Sevanthi, R., Rajagopalan, S., Estrada, N., Sturchio, N., Hatzinger, P. B., Anderson, T. A., Orris, G.,
- 572 Betancourt, J., Stonestrom, D., Latorre, C., Li, Y. and Harvey, G. J.: Global patterns and environmental controls of
- 573 perchlorate and nitrate co-occurrence in arid and semi-arid environments, Geochim. Cosmochim. Acta, 164, 502-522, 574 doi:10.1016/J.GCA.2015.05.016, 2015.
- 575 Kassambara, A. and Mundt, F.: Package "factoextra," Extr. Vis. results Multivar. data Anal., 76 [online] Available from: 576 https://github.com/kassambara/factoextra/issues (Accessed 11 August 2020), 2017.
- Kirkwood, C., Cave, M., Beamish, D., Grebby, S. and Ferreira, A.: A machine learning approach to geochemical mapping, J. 577 578 Geochemical Explor., 167, 49-61, doi:10.1016/j.gexplo.2016.05.003, 2016.
- 579 LaPrade, K. E.: Climate, geomorphology, and glaciology of the Shackleton Glacier area, Queen Maud Mountains, 580 Transantarctic Mountains, Antarctica, Antarct. Res. Ser., 36(9), 163-196, doi:10.1029/ar036p0163, 1984.
- 581 Levy, J. S., Fountain, A. G., Welch, K. A. and Lyons, W. B.: Hypersaline "wet patches" in Taylor Valley, Antarctica, Geophys. Res. Lett., 39(5), n/a-n/a, doi:10.1029/2012GL050898, 2012. 582
- 583 Lyons, W. B., Welch, K. A., Neumann, K., Toxey, J. K., McArthur, R., Williams, C., McKnight, D. M. and Moorhead, D. 584 L.: Geochemical linkages among glaciers, streams and lakes within the Taylor Valley, Antarctica, Ecosyst. Dyn. a polar 585 desert; McMurdo dry Val. Antarct., 72, 77-92, doi:10.1029/AR072p0077, 1998.
- 586 Lvons, W. B., Deuerling, K., Welch, K. A., Welch, S. A., Michalski, G., Walters, W. W., Nielsen, U., Wall, D. H., Hogg, I. 587 and Adams, B. J.: The Soil Geochemistry in the Beardmore Glacier Region, Antarctica: Implications for Terrestrial
- 588 Ecosystem History, Sci. Rep., 6, 26189, doi:10.1038/srep26189, 2016.
- 589 Magalhães, C., Stevens, M. I., Cary, S. C., Ball, B. A., Storey, B. C., Wall, D. H., Türk, R. and Ruprecht, U.: At Limits of 590 Life: Multidisciplinary Insights Reveal Environmental Constraints on Biotic Diversity in Continental Antarctica, edited by F. 591
- de Bello, PLoS One, 7(9), e44578, doi:10.1371/journal.pone.0044578, 2012.
- 592 Nakada, M. and Lambeck, K.: The melting history of the late Pleistocene Antarctic ice sheet., 1988.
- 593 Nkem, J. N., Virginia, A. R. A., Barrett, A. J. E., Wall, D. H. and Li, A. G.: Salt tolerance and survival thresholds for two 594 species of Antarctic soil nematodes, Polar Biol., 29, 643-651, doi:10.1007/s00300-005-0101-6, 2006.
- 595 Patel, J., Shah, S., Thakkar, P. and Kotecha, K.: Predicting stock market index using fusion of machine learning techniques, 596 Expert Syst. Appl., 42(4), 2162-2172, doi:10.1016/j.eswa.2014.10.031, 2015.
- 597 Peters, J., De Baets, B., Verhoest, N. E. C., Samson, R., Degroeve, S., Becker, P. De and Huybrechts, W.: Random forests as
- 598 a tool for ecohydrological distribution modelling, Ecol. Modell., 207(2-4), 304-318, doi:10.1016/j.ecolmodel.2007.05.011, 599 2007.

- Prasad, A. M., Iverson, L. R. and Liaw, A.: Newer classification and regression tree techniques: Bagging and random forests
 for ecological prediction, Ecosystems, 9(2), 181–199, doi:10.1007/s10021-005-0054-1, 2006.
- 602 R Core Team: R: A language and environment for statistical computing, 2020.
- 603 Scarrow, J. W., Balks, M. R. and Almond, P. C.: Three soil chronosequences in recessional glacial deposits near the polar 604 plateau, in the Central Transantarctic Mountains, Antarctica, Antarct. Sci., 26(5), 573–583,
- 605 doi:10.1017/S0954102014000078, 2014.
- 606 Stafoggia, M., Bellander, T., Bucci, S., Davoli, M., de Hoogh, K., de' Donato, F., Gariazzo, C., Lyapustin, A., Michelozzi,
- 607 P., Renzi, M., Scortichini, M., Shtein, A., Viegi, G., Kloog, I. and Schwartz, J.: Estimation of daily PM10 and PM2.5
- 608 concentrations in Italy, 2013–2015, using a spatiotemporal land-use random-forest model, Environ. Int., 124, 170–179,
- 609 doi:10.1016/j.envint.2019.01.016, 2019.

Stevens, M. I. and Hogg, I. D.: Long-term isolation and recent range expansion from glacial refugia revealed for the endemic
 springtail Gomphiocephalus hodgsoni from Victoria Land, Antarctica, Mol. Ecol., 12(9), 2357–2369, doi:10.1046/j.1365 294X.2003.01907.x, 2003.

613 Stevens, M. I., Greenslade, P., Hogg, I. D. and Sunnucks, P.: Southern Hemisphere Springtails: Could Any Have Survived 614 Glaciation of Antarctica?, Mol. Biol. Evol., 23(5), 874–882, doi:10.1093/molbev/msj073, 2006.

Talarico, F. M., McKay, R. M., Powell, R. D., Sandroni, S. and Naish, T.: Late Cenozoic oscillations of Antarctic ice sheets revealed by provenance of basement clasts and grain detrital modes in ANDRILL core AND-1B, Glob. Planet. Change, 96– 97, 23–40, doi:10.1016/j.gloplacha.2009.12.002, 2012.

- Tesoriero, A. J., Gronberg, J. A., Juckem, P. F., Miller, M. P. and Austin, B. P.: Predicting redox-sensitive contaminant
 concentrations in groundwater using random forest classification, Water Resour. Res., 53(8), 7316–7331,
- 620 doi:10.1002/2016WR020197, 2017.
- Toner, J. D., Sletten, R. S. and Prentice, M. L.: Soluble salt accumulations in Taylor Valley, Antarctica: Implications for
 paleolakes and Ross Sea Ice Sheet dynamics, J. Geophys. Res. Earth Surf., 118(1), 198–215, doi:10.1029/2012JF002467,
 2013.
- Wall, D. H., Bardgett, R. D., Behan-Pelletier, V., Herrick, J. E., Jones, H., Ritz, K., Six, J., Strong, D. R. and van der Putten,
 W. H., Eds.: Soil Ecology and Ecosystem Services, Oxford University Press, Oxford., 2012.
- Warton, D. I., Duursma, R. A., Falster, D. S. and Taskinen, S.: smatr 3- an R package for estimation and inference about
 allometric lines, Methods Ecol. Evol., 3(2), 257–259, doi:10.1111/j.2041-210X.2011.00153.x, 2012.
- Webster-Brown, J., Gall, M., Gibson, J., Wood, S. and Hawes, I.: The biogeochemistry of meltwater habitats in the Darwin
 Glacier region (80°S), Victoria Land, Antarctica, Antarct. Sci., 22(6), 646–661, doi:10.1017/S0954102010000787, 2010.
- Wilson, D. J., Bertram, R. A., Needham, E. F., van de Flierdt, T., Welsh, K. J., McKay, R. M., Mazumder, A., Riesselman,
 C. R., Jimenez-Espejo, F. J. and Escutia, C.: Ice loss from the East Antarctic Ice Sheet during late Pleistocene interglacials,
 Nature, 561(7723), 383–386, doi:10.1038/s41586-018-0501-8, 2018.
- 633 Zeglin, L. H., Sinsabaugh, R. L., Barrett, J. E., Gooseff, M. N. and Takacs-Vesbach, C. D.: Landscape Distribution of 634 Microbial Activity in the McMurdo Dry Valleys: Linked Biotic Processes. Hydrology. and Geochemistry in a Cold Desert
- 635 Ecosystem, Ecosystems, 12(4), 562–573, doi:10.1007/s10021-009-9242-8, 2009.
- 636