



Operational implementation of the burned area component of the Copernicus Climate Change Service: from MODIS 250m to OLCI 300m data

Joshua Lizundia-Loiola¹, Magí Franquesa¹, Martin Boettcher², Grit Kirches², M. Lucrecia Pettinari¹,
5 Emilio Chuvieco¹

¹Environmental Remote Sensing Research Group, Department of Geology, Geography and the Environment, Universidad de Alcalá, Calle Colegios 2, Alcalá de Henares, 28801, Spain

²Brockmann Consult GmbH, Chrysanderstr 1, 21029 Hamburg, Germany

Correspondence to: Joshua Lizundia-Loiola (joshua.lizundia@uah.es)

10 **Abstract.** This paper presents a new global, operational burned area (BA) product at 300 m, called C3SBA10, generated from Sentinel-3 Ocean and Land Colour Instrument (OLCI) near-infrared (NIR) reflectance and Moderate Resolution Imaging Spectroradiometer (MODIS) thermal anomaly data. This product was generated within the Copernicus Climate Change Service (C3S). Since C3S is a European service, it aims to use extensively the European Copernicus satellite missions, named Sentinels. Therefore, one of the components of the service is adapting previous developed algorithms to the
15 Sentinel sensors. In the case of BA datasets, the precursor BA dataset (FireCCI51), which was developed within the European Space Agency's (ESA) Climate Change Initiative (CCI), was based on the 250m-resolution NIR band of the MODIS sensor, and the effort has been focused on adapting this BA algorithm to the characteristics of the Sentinel-3 OLCI sensor, which provides similar spatial and temporal resolution to MODIS. As the precursor BA algorithm, the OLCI's one combines thermal anomalies and spectral information in a two-phase approach, where first thermal anomalies with a high
20 probability of being burned are selected, reducing commission errors, and then a contextual growing is applied to fully detect the BA patch, reducing omission errors. The new BA product includes the full time-series of S3 OLCI data (2017-present). Following the specifications of the FireCCI project, the final datasets are provided in two different formats: monthly full-resolution continental tiles, and monthly global files with aggregated data at 0.25-degree resolution. To facilitate the use by global vegetation dynamics and atmospheric emission models several auxiliary layers were included, such as land cover and
25 cloud-free observations. The C3SBA10 product detected 3.77 Mkm², 3.59 Mkm², and 3.63 Mkm² of annual BA from 2017 to 2019, respectively. The quality and consistency assessment of C3SBA10 and the precursor FireCCI51 was done for the common period (2017-2019). The global spatial validation was performed using reference data derived from Landsat-8 images, following a stratified random sampling design. The C3SBA10 showed commission errors between 14-22% and omission errors from 50 to 53%, similar to those presented by the FireCCI51 product. The temporal reporting accuracy was
30 also validated using 4.7 million active fires. 88% of the detections were made within 10 days after the fire by both products. The spatial and temporal consistency assessment performed between C3SBA10 and FireCCI51 using four different grid



sizes (0.05°, 0.10°, 0.25°, and 0.50°) showed global, annual correlations between 0.93 and 0.99. This high consistency between both products ensures a global BA data provision from 2001 to present. The datasets are freely available through the Copernicus Climate Data Store (CDS) repository (DOI: <https://doi.org/10.24381/cds.f333cf85>, Lizundia-Loiola et al. (2020a)).

1 Introduction

Interactions between climate and fire are bidirectional. On the one hand, climate impacts fire regimes (Marlon et al., 2013;Turco et al., 2019b), mostly by modifying temperature and precipitation patterns, which in turn impact fire ignition and behaviour through changes in soil and fuel moisture, vegetation productivity, and fuel availability (Andela and van der Werf, 2014;Daniau et al., 2013;Enright et al., 2015;Moritz et al., 2012;Bowman et al., 2020). On the other hand, biomass burning is a critical source of aerosols and greenhouse gases that directly affect atmospheric chemistry (van der Werf et al., 2017;Ward et al., 2012) and carbon budgets (Yue et al., 2020;Poulter et al., 2015) and carbon stocks (Friedlingstein et al., 2020). Emissions from fires increase aerosol optical depth, modifying the radiation budget and, thus, warming the lower atmosphere, which affects regional temperature, clouds, and precipitation patterns (Tosca et al., 2010;Tosca et al., 2014).

Fires have important ecological implications as well. They are a critical component supporting biodiversity (Kelly and Brotons, 2017), since they alter vegetation composition, structure, and succession (Bowman et al., 2009). Furthermore, wildfires have significant societal impacts, particularly when extreme events occur as a result of climate anomalies. Large damages on people's lives, health and infrastructures have been observed in recent catastrophic fire seasons of Southern Europe, Australia and Western USA (Bowman et al., 2020;Turco et al., 2019b).

The Global Climate Observing System (GCOS) programme promotes international efforts to generate a wide range of observations required for monitoring the Earth's climate system (<https://gcos.wmo.int/en/home>, last accessed December 2020). These observations are commonly categorized under the umbrella of Essential climate variables (ECV) (Bojinski et al., 2014), which critically contribute to the characterisation of Earth's climate system. The European Space Agency (ESA) responded to the GCOS needs with the Climate Change Initiative (CCI) in 2009. The CCI programme was initially focused on 13 GCOS ECVs (now extended to 21 ECVs), each of them addressed within a dedicated project that encompassed algorithm development, validation, uncertainty characterisation, and large-scale EO data processing among other tasks (Plummer et al., 2017).

Fire disturbance is one of the initial ECVs tackled within the CCI programme through the FireCCI project (<https://climate.esa.int/en/projects/fire/>, last accessed December 2020). GCOS identified burned area (BA) as the key parameter for defining the Fire Disturbance ECV, complemented by active fires and fire radiative power (FRP) (GCOS, 2016). Therefore, the principal objective of FireCCI is to produce long-term time series of global BA data. Although BA information is required for a wide range of applications (Chuvieco et al., 2019;Mouillot et al., 2014), the FireCCI project is mainly oriented towards climate modellers and, hence, BA data have been mainly used for the characterisation of fire



emissions for atmospheric chemistry modelling (Seiler and Crutzen, 1980;van der Werf et al., 2017) and the
65 parameterization of fire-enabled Dynamic Global Vegetation Models (DGVMs) (Forkel et al., 2017;Forkel et al.,
2019;Hantson et al., 2016;Hantson et al., 2020). These applications strongly benefit from an extended temporal coverage of
BA datasets, while assuring their temporal and spatial consistency (Heil, 2019).

The FireCCI project has developed three BA algorithms to generate BA data from moderate-resolution sensors. The three
resulting BA products were named FireCCI41, based on Envisat satellite's Medium Resolution Imaging Spectrometer
70 (MERIS) data (Alonso-Canas and Chuvieco, 2015), covering the period 2005-2011 at 300m spatial resolution; the
FireCCI50 product, based on 250m resolution Moderate Resolution Imaging Spectroradiometer's (MODIS) bands (Chuvieco
et al., 2018), and an improved version of this product, named FireCCI51, which extends from 2001 to 2019 (Lizundia-Loiola
et al., 2020b). These three algorithms are based on a hybrid approach, combining Near-Infrared (NIR) and Red reflectance
with active fires detections obtained from thermal anomalies. The three algorithms also followed a two-phase approach (seed
75 detection + region growing), although the newest one overcame the previous version's weaknesses by improving BA
detection using active fire cluster-based thresholds (Lizundia-Loiola et al., 2020b).

A few years after the ESA CCI programme started, the European Commission launched the Copernicus programme, which
included different services mostly based on Earth observation data. One of them is the Copernicus Climate Change Service
(C3S, <https://climate.copernicus.eu/>, last accessed December 2020). C3S has been developed as an operational service that
80 routinely produces datasets for more than 20 ECVs (Thépaut et al., 2018). It builds upon the research and development
carried out by the worldwide scientific community, especially the European one. In this regard, the C3S programme is now
in charge of the operational processing of several ECVs, many of which were generated within the CCI programme (Thépaut
et al., 2018). Since C3S is a European service, it aims to use extensively the European Copernicus satellite missions, named
Sentinels. Therefore, one of the components of the service is adapting previous developed algorithms to the Sentinel sensors.
85 In the case of BA datasets, the precursor BA dataset (FireCCI51) was based on 250m resolution bands of the MODIS sensor
(Lizundia-Loiola et al., 2020b) and, hence, the effort has been focused on adapting this BA algorithm to the characteristics of
the Ocean and Land Colour Instrument (OLCI) on board Sentinel-3 (S3), which provides similar spatial and temporal
resolution to MODIS. Built upon the experience of the Envisat's MERIS sensor, OLCI was initially designed for ocean
monitoring, but it has been successfully used in several land applications (Zhang et al., 2020;Pastor-Guzman et al., 2020).
90 MERIS was the basis of the abovementioned FireCCI41 BA dataset (Alonso-Canas and Chuvieco, 2015), and therefore, it
was expected that OLCI would show similar BA detection capabilities. In fact, due to its better temporal resolution (3-day
revisit time for MERIS and 1 day for OLCI, when the two S3 are available), it was expected that OLCI would detect BA
more accurately than MERIS.

This paper aims to present the new operational, global BA product OLCI C3S Burned Area v1.0 (hereafter C3SBA10)
95 developed under the framework of the C3S, based on OLCI data. The manuscript briefly describes the original algorithm
(FireCCI51) in Section 2.1, while the pre-processing, adaptation, uncertainty characterisation, generation, and distribution of
the operational BA dataset are explained in Sections 2.2-2.5. The operational dataset (2017-present) was designed to be as



consistent as possible with its predecessor FireCCI51 (2001-2019) to ensure that multi-decadal analyses can benefit from both datasets uninterrupted. Hence, quality assessment was performed in two stages. First, the period 2017-2019 was spatially validated using reference data obtained from multitemporal Landsat-8 images and temporally using active fire information (Section 2.6). Then, an inter-comparison was performed between FireCCI51 and C3SBA10 datasets to provide users with insights about the consistency between the two products, pointing out similarities and discrepancies (Section 2.7). The results of the BA detections for the common time series (2017-2019) as well as the quality assessment of the product are presented in Section 3, followed by the discussion and conclusion (Section 4 and Section 5, respectively).

105 2 Methods

2.1 Original burned area algorithm

The original FireCCI51 algorithm was initially developed for the Terra MODIS NIR band at 250m spatial resolution and, therefore, some minor changes were introduced to adapt it to the new sensor (Section 2.3). A detailed description of the original algorithm can be found on Lizundia-Loiola et al. (2020b), although a simplified description is provided in the following paragraphs and in Figure 1.

The FireCCI51 algorithm follows a hybrid approach that takes advantage of combining active fires and spectral information. It first creates monthly composites of NIR reflectance by searching for minimum NIR reflectance while maximising the proximity to the fire date determined by the nearest active fire. Then, a second variable is computed from the relative NIR drop of consecutive composites. The third variable is generated from neighbour active fires that are close enough in space (<1875m) and time (<4 days) to be aggregated into spatio-temporal clusters (STC) (Lizundia-Loiola et al., 2020b).

These variables are the basis for the two-phase hybrid approach used by the FireCCI51 algorithm. The aim of the first step is to minimise commission errors by selecting seed-pixels with a high probability of being burned (seed-phase). NIR and relative NIR drop thresholds are obtained for each STC based on specific burned and unburned samples. The burned sample is defined by the active fires that belong to the STC. The unburned one is composed by pixels from the vicinity of the STC, i.e. those located within a strip between 10 and 20 km from the STC. Active fires are filtered out using in each case the thresholds of the corresponding STC to obtain seed-pixels. Then, a contextual growing is applied from each seed-pixel, being the growing stop criteria the thresholds of the STC to which the seed-pixel belongs (growing-phase). The purpose of this second step is to minimise omission errors by detecting the full burned patch.

The final result is a monthly classification where each burned pixel is labelled by its day of detection, i.e. the day of the year from which the NIR value of the monthly composite was extracted.

The adaptation of FireCCI51 to C3SBA10 BA algorithm was done in two steps. First, it was checked whether any parameter that controlled the original BA algorithm needed to be adapted. Second, the most suitable input band among the different NIR bands of the OLCI sensor was selected. Both aspects depend on the datasets used as input for the BA algorithm and, hence, Section 2.2 provides information about them. This is used in Section 2.3 to justify how the adaptation was performed.



130 2.2 Input data

2.2.1 OLCI Surface directional reflectances

The S3 satellites were designed to provide operational ocean and land observation services (<https://sentinel.esa.int/web/sentinel/missions/sentinel-3>, last accessed December 2020). They represent the continuity of the European Remote Sensing satellites (ERS-1 and ERS-2), the Envisat and the Satellite Pour l'Observation de la Terre (SPOT). S3 have a low (814.5 km), high inclination (98.65°), sun-synchronous Earth-orbit. With a period of ~101 minutes, S3 need 27 days to complete a full cycle (385 orbits). The first satellite, called S3A, was launched in February 2016 and the second one, called S3B, was launched two years later. The orbit of S3B is identical to S3A but flies $\pm 140^\circ$ out of phase to improve revisit time. The design life of S3 is 7.5 years (consumables for 12 years), but the mission is expected to last longer, as two additional satellites will be launched during the first years of the current decade.

140 The OLCI sensor, one of S3's main payloads, is a push-broom imaging spectrometer composed of five cameras that are 12.6° tiled in western direction to mitigate potential sun-glint effects (<https://sentinel.esa.int/web/sentinel/technical-guides/sentinel-3-olci>, last accessed December 2020). It has a swath width of 1270 km (Field-of-View = 68.6°) and it offers global coverage at 300m every ~2 or ~1 day depending on the number of satellites available (S3A or S3A+S3B, respectively) and application (ocean or land). The OLCI instrument includes the same 15 spectral bands of MERIS (400-145 1020 nm), and six extra channels that were included to improve the atmospheric and aerosol corrections (Table 1). All this information is included in the OL_1_EFR Level 1B product, which provides full resolution, calibrated, ortho-geolocated and spatially re-sampled Top-of-Atmosphere (TOA) radiances (<https://sentinel.esa.int/web/sentinel/user-guides/sentinel-3-olci/product-types/level-1b>, last accessed December 2020).

This product was converted to surface directional reflectance (SDR) using an automated pre-processing chain that generates 150 OLCI Level 3 SDR. The first module of the chain converted the TOA spectral radiance to the apparent SDR. Then, a pixel identification module calculates a set of pixel classification attributes, like clear, cloud, snow/ice, cloud shadows, etc. The third module is an atmospheric correction algorithm, which included the correction for the absorbing and scattering effects of atmospheric gases, in particular ozone, oxygen and water vapour, of the scattering of air molecules (Rayleigh scattering) and the correction of absorption and scattering due to aerosol particles. The atmospheric correction including aerosol retrieval 155 was first developed in the GlobAlbedo project (GlobAlbedo, 2013). The last processing module of the pre-processing chain (compositing and mosaicking) was applied to retrieve the final Level 3 products. These 1-day SDR composites were derived from an input set of single satellite observations (i.e., SDR and pixel classification data described previously). Thus, those single observations were (i) reprojected onto a Plate Carrée grid, (ii) temporally aggregated for given binning cells (tiles), and (iii) mosaicked based on the binning cells onto a Level 3 product of 10x10 degrees tiles (Figure 2).



160 2.2.2 Auxiliary data

Information on active fires was used to guide both the seed and the growing phase of the BA algorithm. The input product was the MODIS MCD14ML Collection 6, which provides global monthly information on the location of thermal anomalies using both Terra and Aqua MODIS bands (Giglio et al., 2018b) at 1 km spatial resolution. The MCD14ML product is provided in ASCII format and one of its attributes specifies the presumed origin of the thermal anomaly: 0 for presumed
165 vegetation fire, 1 for active volcano, 2 for other static land sources, and 3 for offshore thermal anomalies. Among these categories, only those anomalies labelled as 0 were considered. This product was selected as it was the one used by the original FireCCI51 algorithm (Lizundia-Loiola et al., 2020b).

Global, annual land cover (LC) maps, which were provided by the Land Cover C3S, were used as well (<https://cds.climate.copernicus.eu/cdsapp#!/dataset/satellite-land-cover?tab=overview>, last accessed December 2020). LC
170 production was transferred from CCI to C3S at the same time as the BA datasets, which meant that the same working group was the responsible of adapting the algorithms. So, although the FireCCI51 BA algorithm used the latest LC v2.0.7 of CCI project (ESA, 2017), the LC version v2.1.1 from C3S was considered fully consistent with its predecessor (Defourny et al., 2020). In fact, the annual land cover maps from 1992 to 2015 developed within the CCI were brokered to C3S and extended
175 of 2016, 2017, and 2018. Data at 333m spatial resolution from PROBA-V was used to generate the global, annual LC maps of 2016, 2017, and 2018. The use of the LC products had two main objectives. On the one hand, they were used to mask unburnable areas, i.e. bare areas, urban areas, water bodies, and permanent snow and ice classes, to decrease the amount of data to be processed. On the other hand, they were used to report which land cover class was burned. Since the idea was to characterise the situation before the fire, the LC map that was applied in each year was the one prior to the BA analysis, e.g. LC of 2016 was used for the BA processing of 2017. This product was selected to 1) ensure consistency with the original
180 algorithm, 2) ensure consistency within the C3S, and 3) use a product that offered LC information at a similar spatial resolution of the input OLCI SDR.

2.3 Adapted burned area algorithm

The adaptation was based on the analysis of the same study sites that were used for the FireCCI algorithm (Lizundia-Loiola et al., 2020b) (Figure 2) as they encompassed a wide variety of vegetation types and fire regimes. This study sites were the
185 thirteen equivalent 10x10° degree OLCI tiles of the MODIS standard, sinusoidal tiles used in the development of the original FireCCI51 algorithm (Lizundia-Loiola et al., 2020b). Six of those tiles were distributed in regions where fire perimeters from different official services were available. The western coast of US, in California, was chosen as representative of temperate forests, where fire perimeters were obtained through the Fire and Resource Assessment Program (FRAP, <http://frap.fire.ca.gov/> last accessed on December 2020). Fire perimeters from the Northern Australian Fire Information
190 (NAFI, <http://www.firenorth.org.au/nafi3/>, last accessed on December 2020) served to assess the adapted algorithm's performance in tropical savanna. Finally, as an example of the boreal forests, perimeters of fires larger than 200 ha were



downloaded from the Canadian National Fire Database (CNFDB, <http://cwfis.cfs.nrcan.gc.ca/ha/nfdb>, last accessed
December 2020). The remaining seven tiles were used to visually check that no major problems arise when applying the
adapted algorithm to the new sensor, e.g. border effects in Central Africa, problems on the thresholds due to contrasted fire
regimes within the same tile in Angola or region-growing problems due to high BA/active fire ratios found near Kazakhstan.
The adaptation of the algorithm to OLCI and its integration in the C3S operational system were carried out during 2018 and,
hence, the training dataset's temporal coverage was limited to 2017, when only S3A was available.

Most of the parameters that control the FireCCI51 BA algorithm were independent from the base sensor used to detect BA,
such as the time-gap used to temporally aggregate active fires or the absolute thresholds that were fixed for global detection
(Lizundia-Loiola et al., 2020b). However, parameters linked to a distance might be affected by the spatial resolution of both
the surface reflectance product and the active fire product. The fixed distance used to spatially aggregate active fires into a
same cluster (Section 2.1), for example, might depend on the spatial resolution of the input active fires product. Higher
spatial resolution thermal bands can lead to higher densities of active fire pixels within a same burned patch (Oliva and
Schroeder, 2015; Schroeder et al., 2014) and, therefore, the distance that is needed to spatially aggregate them could be
smaller. In this case, there was no difference between the input active fire product used in the development of the original
algorithm and the adapted one (Section 2.2.2). Similarly, the distances used to define the unburned region around active fire
clusters (Section 2.1) depend on both the spatial resolution of the input surface reflectance and the density of active fires per
burned patch. The spatial resolution of the base sensor directly affects the number of surface reflectance pixels that fall into
the unburned strip and, hence, could be used to estimate the thresholds. The same distance could imply the selection of much
more pixels when the spatial resolution of the input sensor was high while in coarser resolutions could lead to a number of
pixels that may be insufficient. The difference of the spatial resolution of the input sensors used in the original algorithm
(250m) and the adapted one (300m) were considered sufficiently similar to keep the parameters that defined the unburned
strip.

Since the original FireCCI51 BA algorithm and, hence, the adapted one, needed only one input NIR band, the most suitable
OLCI band (Table 1) had to be selected. In the case of MODIS it was not necessary to make such selection because it offers
only one NIR band (Lizundia-Loiola et al., 2020b). Based on a sensitivity analysis of the MERIS bands carried out by Oliva
et al. (2011) during the development of FireCCI41, two NIR bands were considered candidates to be the input of the adapted
BA algorithm: Oa12 centred at 753.75nm and Oa17 centred at 865nm (Table 1). These spectral bands were the most
sensitive bands for BA detection, although no significant differences were found between them (Oliva et al., 2011).

The adapted BA algorithm was run two times with the training dataset, i.e. thirteen study sites and year 2017, using in each
process one of the mentioned bands. The assessment against the services' fire perimeters showed slightly more accurate
results of the 865nm band over the 753.75nm one (same commission error and ~1% lower omission error). None of the
bands showed unexpected anomalies in the rest of the tiles. Taking into account these results and the similarity between the
OLCI 865 ± 10 nm band and the MODIS 858.5 ± 17.5 nm band used in the original FireCCI51 BA algorithm, which implied



225 a greater consistency between the algorithms, the OLCI's 865nm NIR band was selected as main input for the adapted algorithm.

2.4 Product generation

The adapted algorithm was processed in a monthly basis for each 10°x10° tile and, hence, the raw outputs were provided per month and tile. Pre-processing and BA processing took advantage of the independence of tiles and were produced in parallel
230 for multiple tiles. BA processing needed a time series of two prior composites to detect the NIR drops, and it required active fire data as seeds for the growing step, which introduced constraints for sequential processing of the time series. These raw outputs were formatted, following mainly the specifications of the FireCCI project (Chuvieco et al., 2017), to obtain the pixel and grid products that were delivered to the final users in NetCDF4 format. The former provided the original BA classification at the original spatial resolution (300m) distributed in six continental tiles (Figure 3). The latter computed a
235 global aggregated summary of the raw results at 0.25° spatial resolution. Both the pixel and grid products provided auxiliary layers suggested by the project end-users, including the land cover that was burned, and the uncertainty, fraction of observed area and burnable area (Table 2).

A total of 273 tiles that encompass the main burnable areas of the Earth, except from Antarctica, some small islands and regions northern than 80° latitude, were processed. The C3SBA10 extended the spatial coverage provided by the FireCCI51
240 dataset. The northern latitudes between 70° and 80° N of North America and Greenland not processed before were included since rare fire events were observed in the last years in far northern latitudes (Amos, 2019;Evangelidou et al., 2019). This change affected the original bounding box of the Area 1 of the pixel product defined by Chuvieco et al. (2017) within the FireCCI, whose eastern border was extended from 50° W to 26° W in C3SBA10 (Figure 3).

The operational dataset was processed since January 2017, ensuring up to three years of overlap (2017-2019) between the
245 FireCCI51 and C3SBA10 products. The delivery frequency of the operational product ranged from three to six months depending on the availability of the MODIS active fire MCD14ML product.

2.5 Uncertainty characterisation

Uncertainty characterisation is increasingly demanded by the end-users of BA products since it greatly helps the parameterization of climate models (Heil, 2019). Although important advances have been made in the last years regarding
250 uncertainty characterisation of ECVs (Merchant et al., 2017;Mittaz et al., 2019;Sayer et al., 2019), methods to standardise the generation of this variable are still needed (Chuvieco et al., 2019). Therefore, the CS3BA10 algorithm follows the same approach used in the FireCCI51 product. Four input variables related to different phases of the original algorithm were used to compute the estimation of BA detection uncertainty: monthly NIR composite (NIR), monthly relative NIR drop (RelΔNIR), number of valid observations in the first 10 post-fire days (obs) and the distance to the nearest seed-pixel (dist).
255 The uncertainty was computed using a logistic regression. Based on those four variables and the results of 2008 of the study sites that were used to develop the original algorithm, the final coefficients of the regression were obtained.



In the case of the adapted C3SBA10 algorithm, the coefficients of the logistic regression were updated (Eq. 1) using the four input variables and results of the training dataset described in Section 2.3:

$$pb = \frac{1}{1+e^c} \quad c = -(4.068 - 0.002926 \cdot NIR + 0.003942 \cdot Rel\Delta NIR - 0.01303 \cdot obs - 17.29 \cdot dist) \quad (1)$$

260 2.6 Validation

2.6.1 Spatial assessment

The spatial accuracy assessment was conducted for the period 2017-2019 for both the FireCCI51 and C3SBA10 products. An independent reference BA dataset derived from Landsat-8 OLI imagery was produced in compliance with the CEOS LPVS stage 3 validation requirements (Morissette et al., 2006), which implies a rigorous statistical selection of reference sites
265 representing diverse global conditions. Hence, for each calendar year 100 sample units were randomly selected using a stratified random sampling design. The sampling units were spatially defined by the Thiessen scene areas (TSAs) constructed by Cohen et al. (2010) and Kennedy et al. (2010) from the Landsat-frames World Reference System 2 (WRS-2). TSAs enable the partition of the earth's surface into non-overlapping spatial units, allowing the computation of unbiased estimators (Gallego, 2005).

270 To ensure that sampled units were distributed across the main biogeographic regions, a first level of stratification was applied based on the biomes defined by the ecoregions2017 map (Dinerstein et al., 2017). The original 14 biomes were aggregated into 8 major biomes, i.e. boreal forest, deserts and xeric shrublands, Mediterranean, temperate forest, temperate savanna, tropical forest, tropical savanna, and tundra (Figure 4). A second level of stratification was applied based on the FireCCI51 BA extent for each calendar year. Thus, each TSA within each biome and year was assigned to a high or low BA
275 stratum following the approach of Padilla et al. (2014). Then, the sample's allocation within the 16 resulting strata (8 biomes x 2 high-low strata) was established proportionally to $N_h \sqrt{\overline{BA}_h}$, where N_h is the number of units in stratum h and \overline{BA}_h is the mean mapped BA for stratum h (Padilla et al., 2017).

The reference data for the selected sites were obtained from pairs of Landsat-8 images using a semi-automatic classification algorithm (Bastarrika et al., 2014), followed by a visual inspection to confirm the correct identification of burned patches.
280 Consecutive pairs of images were used to obtain long temporal reference data, covering several months, the entire fire season or the whole year, depending on the image availability. These long temporal reference data allow for an extended temporal overlapping between reference and BA datasets, which makes it possible to improve the spatial accuracy assessment of the BA products, minimizing the impact of the product's temporal reporting accuracy (Section 2.6.2) in the spatial accuracy estimates (Franquesa et al., 2020b).

285 Accuracy metrics were computed based on the error matrix, which is widely used by the scientific community for thematic accuracy assessment (Congalton, 1991). This matrix was derived from the cross tabulation of reference data and the BA products. The metrics used to assess the product accuracy were the Dice coefficient (DC) (Dice, 1945), the relative bias



(relB) and the commission (Ce) and omission (Oe) errors. DC measures the similarity between the two sets of data; the higher its value, the more the accuracy of the burned category. The relB informs about the bias of the BA product relative to the reference BA, negative relB means BA product underestimation and positive relB, overestimation. Ce refers to the burned area mapped by the BA product but not classified as burned in the reference data, and Oe to the burned area in the reference data not mapped in the BA product. All the accuracy metrics and their associated standard errors were estimated for the whole population applying the formulas described in Padilla et al. (2014) and Padilla et al. (2015).

2.6.2 Temporal reporting accuracy assessment

A validation of the temporal reporting accuracy of the products was performed as well. Unlike the assessment of the spatial accuracy, which is already addressed in the previous section, the assessment of the temporal accuracy aims to show how accurately the day-of-burn (JD in Table 2) is determined by the BA product. Then, following the approach of Boschetti et al. (2010), the day-of-detection reported by the active fires (Section 2.2.2) was used as ground truth and compared with the day-of-detection reported by the corresponding pixel of the global BA product. These differences were then analysed per biome as well.

2.7 Consistency assessment

The estimates of accuracy metrics provided by the validation exercise provide an objective, valuable information regarding the spatial and temporal accuracy of the validated products. However, along with the spatial and temporal accuracy, consistency between products is also of primary interest for multi-decadal analyses, e.g. climate modelling (Chuvieco et al., 2019;Heil, 2019). From this perspective, the C3SBA10 operational product cannot be understood as a unique, independent dataset, but as a continuation of its predecessor FireCCI51 product, ensuring uninterrupted global BA estimations from 2001 to present. To account for the similarities and differences that can arise between both datasets spatial and temporal trends of BA were compared.

A fundamental step when comparing global BA datasets is to define both spatially and temporally the comparison grid that will be used to aggregate the information. Although different spatial and temporal comparison units can be found in the literature (Boschetti et al., 2004;Giglio et al., 2010;Humber et al., 2018;Turco et al., 2019a), in this case the requirements of some of the most relevant applications of global BA datasets were used to determine the comparison framework (Chuvieco et al., 2019;Heil, 2019). However, in most cases, different applications use different spatial and temporal analysis units. The “fire model intercomparison project” (FireMIP), for example, which systematically compares different fire-enabled DGVMs, uses global, monthly BA data aggregated at 0.5° spatial resolution for model benchmarking (Hantson et al., 2016). Statistical analyses that target at identifying factors that control fire patterns at global and regional scales commonly require monthly BA data aggregated at 0.25° or coarser spatial resolutions (Andela and van der Werf, 2014;Bistinas et al., 2013;Forkel et al., 2019;Knorr et al., 2014). The Global Fire Emission Database (van der Werf et al., 2017), one of the most widely used inventories in atmospheric chemistry modelling, provides global emissions derived from BA in monthly, daily and 3-hourly



320 temporal scales at 0.25° resolution. However, several global fire emission databases are already available at 0.1° (e.g. Kaiser et al. (2012) updated) and even 0.05° (e.g. Kuenen et al. (2014)) following the increasing spatial resolution of some global atmospheric chemistry models (Aleksankina et al., 2018;Hu et al., 2018). Thus, it was decided to aggregate BA information into four different geographic grid cell sizes, i.e. 0.05°, 0.1°, 0.25°, and 0.5°, in a monthly and annual basis.

The results were aggregated spatially into eight biomes as well, following the first stratification of the validation exercise. 325 Four comparison cases were identified: biome-specific monthly BA, biome-specific annual BA, global monthly BA and global annual BA. A total of 1404 scatterplots (3 years x (12 monthly +1 annual) x (eight biomes + 1 global) x 4 spatial resolutions) were generated based on these comparison cases and the four comparison grids defined above for the analysed years (2017-2019). A linear regression was calculated for each one (Humber et al., 2018;Turco et al., 2019a). Three metrics were used to determine the agreement of the products for a given case. The slope of the regression line was used to analyse 330 the bias, the correlation coefficient (r) was used to determine the agreement between products and the root mean square error (RMSE) was used to account for the dispersion of the estimations.

3 Results

3.1 Trends on global burned area

More than three years (2017 – present) of S3 OLCI data were processed to generate the C3SBA10 global BA product. A 335 total of 273 10°x10° tiles were needed to cover almost all the burnable areas on the Earth. The algorithm results were summarised and published in pixel (300 m resolution) and grid (0.25-degree resolution) products, using in both the NetCDF4 format. Figure 5 shows the annual accumulated grid BA product for the year 2019 while Figure 6 shows the latitudinal and seasonal contribution of BA for the same year. The amount of BA detect per biome for the three complete natural years can be found in Table 3.

340 The C3SBA10 product detected 3.77, 3.68, and 3.59 Mkm² of BA for the years 2017, 2018 and 2019, respectively. The main land cover contributor to global BA was tropical savanna, which had burned 2.8 Mkm² per year (Table 3), representing between 74 and 78% of global burns. Tropical forest had between 9 and 11% of total BA (average 0.35 Mkm²). Almost all the BA of these biomes was located on a relatively small latitudinal band between 20°N and 20°S (Figure 5 and Figure 6) that covers part of South America, Central America, Africa, South East Asia, Indonesia and the northern part of Australia. 345 Among these tropical regions several studies have shown that the African savannas are responsible of around 70% of global BA, followed by the Australian ones, the Brazilian Cerrado and the Orinoquia region in Colombia and Venezuela (Chuvieco et al., 2018;Chuvieco et al., 2016;Giglio et al., 2018a;Lizundia-Loiola et al., 2020b;Roy et al., 2008;van der Werf et al., 2017). In the case of tropical forest, the impacts of deforestation fires on agricultural frontiers can be observed in South America and SE Asia (Chen et al., 2013;van der Werf et al., 2008;van der Werf et al., 2010). A lesser contribution, but still 350 important, comes from the peatland areas of Kalimantan and Sumatra in Indonesia.



The influence of the dry season in these tropical biomes is quite evident, with a marked fire period from November to March in the Northern Hemisphere, and from June to October in the Southern hemisphere (Figure 6). This trend slightly differs in South East Asia where the fire season continues until May.

Another significant belt of fire activity can be found in the temperate forest and grasslands of the Northern hemisphere (Figure 5), with relative maximum around 50° N (Figure 6). The C3SBA10 detected that on average the accumulated contribution to the global fire activity of these biomes was 0.27 Mkm² (7.5%). Most of this BA comes from the grasslands on the Asian steppe, which is dominated by large fires, although there is also an important agricultural activity (Hall et al., 2016;Zhu et al., 2017). The central plain of United States also exhibits fire activity linked to croplands as well as the forest areas of the north-western territories. In the southern hemisphere, the BA derived from the extreme wildfires of 2019-2020 in the temperate forest of south-eastern Australia (Bowman et al., 2020) is noticeable in Figure 5. The monthly variability of BA in temperate areas is marked by both human agricultural practices and dry summer conditions, leading to two separated fire activity periods: one in March-April and the other in July-August.

The deserts and xeric shrublands' amount of BA detected by the C3SBA10 product showed a similar contribution to that of temperate areas, although with much more inter annual variability. The BA on 2017 was 0.26 MKm² while on 2019 this number decreased by a half (0.11 Mkm²) (Table 3). Two areas on the Earth are responsible of most of this BA: central and southern Kazakhstan are affected by large fires in summer (July-August) while the xeric shrublands and grasslands of central and western Australia show fire activity from October to January.

In the highest latitudes (>60°N), the boreal forest showed a clear increasing trend in BA for the three-year period when the C3SBA10 product was computed, from 61,173 km² in 2017 to 90,503 km² in 2019 (50% increase). In the Tundra region, BA showed even higher increases, as in 2019 the product estimated 6 times more BA than the previous year (13,044 km² versus 1,697 km², respectively) (Table 3). In fact, in 2019 the contribution of these two biomes was almost as much as that of temperate forest. Figure 6 shows how the fire season in the northern latitudes has a clear seasonal pattern, with highest activity in the Summer months (July-August).

Finally, BA on Mediterranean areas showed a significant contrast between the BA of 2018 (8,884 km²) and the other two years (24,975 and 27,421 km² in 2017 and 2019, respectively). This inter-annual variation seems to be related to the extreme fire seasons of Portugal in 2017 and in 2019.

3.2 Validation

3.2.1 Spatial assessment

Table 4 shows the estimated accuracy measures and their standard errors (SE) for the validation dataset. The Dice coefficient varied from 0.59 in 2017 (SE=0.03) and 2019 (SE=0.04) to 0.64 (SE=0.03) in 2018. These values were similar to those obtained for the FireCCI51 product, although this one showed higher accuracy for the years 2017 and 2018, but not significantly higher for 2019 (Table 4). The availability of more images due to the presence of two S3 satellites can explain



why 2019 was the most similar year. In general terms, C3SBA10 showed similar (2017 and 2019) or lower (2018) commission errors (Ce) and higher omission errors (Oe) than FireCCI51 through the three-year period. Both products showed negative relative bias for all the years, indicating a systematic underestimation of BA, which is common to other global BA datasets (Boschetti et al., 2019).

Although three years are not enough to extract significant conclusions about the temporal trends of accuracy, a similar behaviour can be observed in both products. Thus, the year that presented the highest accuracy metrics in one product also it did in the other product.

3.2.2 Temporal reporting accuracy assessment

About 4.7 million of active fires were used to estimate the temporal reporting accuracy of the three-year period of both products (Figure 7). On average (2017-2019) the results of the C3SBA10 showed that for deserts and xeric shrubland, both tropical and temperate savanna and Mediterranean biomes more than 90% of burned pixels were detected within the first 10 days after the fire (Table A1). This proportion was slightly lower in temperate and tropical forest where 80% and 86% of the pixels were labelled within 10 days of detection, respectively. The less accurate biomes were the boreal ones, where only the 70% (forest) and 69% (tundra) of the burned pixels were detected within the 10-day period. Although this period could be reasonable for some applications (e.g. dynamic vegetation modelling) there are other applications where it is important to know the precise date of burn. The analysis showed that, on average, only in both savannas and desert and xeric shrubland biomes C3SBA10 was able to detect within 0- or 1-day difference more than 20% of the cases. This accuracy significantly increased if the difference was set to a maximum of 3-days difference, where all biomes double the number of cases that were detected within this accuracy.

The comparison between products showed that the temporal reporting accuracy is quite similar, having little or no impact the availability of more S3 images in 2019. In fact, for all forest biomes and tundra C3SBA10 provides a slightly more accurate reporting of the date of burn (Figure 7).

3.3 Consistency assessment

The relative differences on the amount of BA detected by each product are presented in Table 3. Table 5 shows the consistency metrics (correlation coefficient (r), slope and RMSE) at global and annual scale for the three-years period and the four comparison grid sizes. Figure 8 and Figure 9 show the spatial agreements between FireCCI51 and C3SBA10 for the year 2019.

The global BA detected by the predecessor version FireCCI51 was 0.65 Mkm², 0.56 Mkm², 0.28 Mkm² higher than that one detected by the C3SBA10 product from 2017 to 2019, respectively. Almost all the disparities in terms of total BA can be explained considering the differences shown by both products in tropical savanna (see slope in Figure 8b), where C3SBA10 detected 0.49 Mkm², 0.46 Mkm², 0.26 Mkm² less BA, respectively in the 3 years of the series. The availability of both S3 and, hence, a higher temporal resolution, had a significant impact in the amount of BA detected by the C3SBA10 in those



415 biomes where fire activity is frequent. Thus, the difference between both products was reduced more than a half in tropical forest, temperate savanna, desert and xeric shrubland, and temperate forest and almost a half in tropical savanna (Table 3). In 2019, if tropical savanna was not considered, only 30,000 km² of BA were underestimated by the C3SBA10 versus those of the FireCCI51. This underestimation was partially compensated by the detection of more BA (5,305 km²) in boreal forest and tundra.

420 Even though the annual BA detected by the global BA products differs through the study period, the correlation coefficient (r) showed a significant agreement between them at all comparison grid sizes (Table 5). For instance, the r value at the finest comparison resolution (0.05°) was found to be 0.937 in the year with less agreement (2017) and 0.952 in the year with the highest agreement (2019). This agreement increases if the spatial resolution of the comparison is decreased reaching a 0.99 r value for all the years at 0.5° grid size. However, the slope shows a clear bias towards FireCCI51, pointing an
425 underestimation of the C3SBA10 product at global and annual scale, confirming the differences in the accuracy metrics shown above.

There is a clear increase on the agreement between products in 2019 as well. The correlation metric is the highest in that year (from 0.95 at 0.05° to 0.99 at 0.50° grid resolutions), as well as the slope (from 0.88 at 0.05° to 0.91 at 0.50° grid resolutions), which means an increase in the amount of BA detected by the C3SBA10. The RMSE is also the lowest for that year,
430 indicating that the availability of more images from both Sentinels makes the differences smaller. That is why a more detailed analysis of the results of 2019 was conducted to better understand the spatial and temporal trends of the discrepancies between the products.

Through the eight biomes the lowest correlations at annual scale were those of deserts and xeric shrublands with an $r=0.86$ at 0.05° grid resolution (Figure 8e). Tropical savanna showed the highest annual agreement with correlations from 0.96 (at
435 0.05°) to 0.99 (at 0.50°). However, it had the highest RMSE (2.28 km² at 0.05°) due to the large amount of BA per grid cell. Boreal forest and tundra show a very high agreement as well, with correlation coefficients between 0.94 and 0.98 for all grid sizes and the lowest RMSE, i.e. 0.27 km² for boreal forest and 0.14 km² for tundra at 0.05°. The rest of the biomes presented r values between 0.87-0.89 at 0.05° spatial resolution and above 0.90 from 0.10° upwards.

The two biomes that showed to be unbiased or with a little bias towards C3SBA10 were those located at highest latitudes
440 (boreal and tundra) with slope values near to 1. The rest of the biomes showed an underestimation of the C3SBA10, although in different proportion. Temperate forest and savannas, for instance, had a slope around 0.88 at 0.05° grid size, while at 0.50° it increased up to 0.97 for forest and 0.94 for savanna. The most noticeable underestimation was linked to the Mediterranean (Figure 8f) biome (maximum slope of 0.80 at 0.50° grid resolution). The rest of biomes presented slopes ranging from 0.8 to 0.9.

445 At monthly scale (Figure 8) the agreement between products changes depending on the month. Undoubtedly, the more stable biome through the year is tropical savanna (Figure 8b), since it is the most frequently burned biome and where most of the global BA is located. The lowest r is found in those months that contribute less to annual BA, i.e. February, March, April and May, with correlations from 0.86 to 0.89 at the finest comparison resolution, respectively. The rest of the months showed



450 correlations above 0.92 at 0.05° grid size, being higher in the rest of the comparison resolutions. In fact, at 0.50° the lowest correlation, which was found in April, showed a value higher than 0.95. A systematic underestimation of the C3SBA10 below 25% of the annual BA was noticeable through all the grid cells that belong to tropical savanna (Figure 9), which was traduced in slopes between 0.83 (April) and 0.90 (October) at 0.05° spatial resolution. The influence of tropical savanna is clear in the global metrics, which can in some cases obscure subjacent trends in other biomes.

In most cases, the correlation increases with the proportional contribution to the annual BA. This was the case of boreal 455 forest (Figure 8g) and tundra (Figure 8h) whose peak months clearly matches the summer months, i.e. July and August, which are responsible of the 70% and 77% annual BA in those biomes, respectively, and, at the same time, showed the highest agreement ($r > 0.91$ and slope > 0.92). However, not all biomes have a marked unimodal fire season since they encompass different continents and hemispheres and, therefore, fire activity is affected by diverse climatic and human conditions. In the case of deserts and xeric shrubland biome (Figure 8e) two separate peaks can be found, one in January 460 (20% of annual BA, $r=0.83$, slope=0.92 at 0.05° grid size) and another in July (19% of annual BA, $r=0.92$, slope = 0.88 at 0.05°). Something similar occurs in tropical forests (Figure 8a) where January (13% of annual BA, $r=0.87$, slope=0.77 at 0.05°) and August (14% of annual BA, $r=0.92$, slope=0.88) have the maximum monthly BA values. The tropical savanna shows two clear fire seasons as well (Figure 8b), following the dry seasons of each hemisphere.

The agreement shown between C3SBA10 and FireCCI51 is not as stable in the case of temperate and Mediterranean biomes. 465 In the former case there is a clear discrepancy in the monthly contribution of April to the annual BA. In the case of temperate forest (Figure 8c), the monthly contribution corresponds to 21.3% for C3SBA10 while FireCCI51 estimated it at 16.8%. The same occurs in the temperate savanna (Figure 8d) where the contribution increases from 20.4% (FireCCI51) to 26.8% (C3SBA10). However, this discrepancy affected mainly the RMSE, which is much higher for this month than for the rest of the months in both biomes (RMSE=0.44 km² in savanna and RMSE=0.27 km² in forest at 0.05° grid size). In the 470 Mediterranean biome (Figure 8f), from July to November there is a significant increase in the bias towards FireCCI51, losing the stability shown in the first half of the year, and with slope values as low as 0.51, e.g. in the case of October.

4 Discussion

After adapting the original FireCCI51 algorithm to the characteristics of the C3SBA10, we found that the spatial and temporal trends in global BA of the two products showed consistent temporal and spatial trends. Most differences were 475 observed in the first two years of the time series (2017 and 2018) when only one S3 satellite was operating, and therefore the temporal resolution of OLCI (2-3 days) was much lower than the MODIS sensor (1 day). BA estimations between the two products were very similar in 2019, although C3SBA10 had 0.28 Mkm² less BA (global BA: 3.63 Mkm² for C3SBA10 and 3.91 Mkm² for FireCCI51). Most of this difference (92%) was found in tropical savanna, where a systematic underestimation of the C3SBA10 product was noticeable (Figure 9). This tendency towards underestimation was observed in other global BA 480 products (generally based on coarse spatial resolution sensors) when comparing with regional products based on medium



resolution sensors. For instance, a continental BA product at 20 m derived from Sentinel-2 (S2) for the year 2016 and Sub-Saharan Africa found that global BA products significantly underestimate total BA, as they included 80% less area than the S2 BA product (Roteta et al., 2019). This was mainly caused by the poor detection of small fires (< 250 ha) in global products. C3SBA10 is likely missing more small fires than FireCCI51 due to its coarser spatial resolution (300 m vs 250 m).

485 This tendency towards underestimation was confirmed by the validation exercise carried out through the overlapping three-year period (2017-2019), which showed that both global BA products presented higher omission than commission errors. This trend agrees with the validation exercises of other global BA products that can be found in the literature (Boschetti et al., 2019; Chuvieco et al., 2018; Padilla et al., 2015) as well as with the abovementioned inability of moderate resolution BA products to detect small fire patches. The omission errors presented in this paper were lower than those found by previous

490 authors, e.g. 72.6% for the NASA's standard product in Boschetti et al. (2019) or 81% and 71% for two consecutive versions of FireCCI products (FireCCI41 and FireCCI50) in Chuvieco et al. (2018). However, it must be highlighted that these numbers are not fully comparable to those presented in this paper since all these studies used the so called short units (reference data generated from a maximum period of 16 days) instead of long units (reference data covering several months, as done in this study) to estimate the spatial accuracy of the products. The short units' approach is more affected by the

495 temporal reporting accuracy (Section 3.2.2) of the global BA products than when using long units. A preliminary assessment of this effect was included in Lizundia-Loiola et al. (2020b) and in Franquesa et al. (2020b), showing a decrease of about 25% in commission and 15% in omission errors (in FireCCI51) for a validation sample of Sub-Saharan Africa 2016.

In any case, both the validation exercise and the consistency assessment showed a high agreement between the predecessor global BA product FireCCI51 and the operational C3SBA10. Although this consistency is still high in the years where only

500 S3A is available (2017-2018) it significantly increased in 2019 due to the availability of the second S3B satellite. In fact, in 2019 the C3SBA10 product showed a similar accuracy ($DC=0.59\pm 0.04$) to that estimated for the FireCCI51 ($DC=0.61\pm 0.03$). The global and annual comparison metrics for that year were $r = 0.952$, slope = 0.88 and RMSE = 1.07 km² with a 0.05° spatial resolution, the finest spatial resolution used in the comparison (Table 5). The agreement between the products improved as the comparison resolution decreased (Table 5, and Figure 8).

505 A deeper analysis of the year 2019 showed different trends in the consistency between products among the biomes. It was found that the changes in the correlation between products were linked to the BA detected for each month and biome. Although this relationship did not seem to be linear, i.e. a specific increase in BA did not mean a parallel increase in the correlation, it was clear that the larger the BA detected the higher the correlation between products. The tropical savanna was an obvious example, where most of global BA was found. The boreal regions were another example where the correlation is

510 the highest in the months with the highest fire activity, while it significantly decreases for the rest of the months. The extreme wildfire event of Eastern Australia (Bowman et al., 2020) clearly showed this phenomenon since it significantly increased BA and agreement between products in December in the Mediterranean biome and in November and December in temperate forests.



515 There were some unusual cases as well, where higher BA did not necessarily mean higher correlation. This was the case of
the temperate areas where the correlation decreased in April. Temperate savanna and forest biomes encompass almost all the
Northern Hemisphere mid-latitude croplands located in the Russian Federation and Kazakhstan (Hall et al., 2016). Cropland
fires are characterised to be small and transient (McCarty et al., 2009; Randerson et al., 2012), which made moderate
resolution BA products (250-500 m) unable to properly characterise the extent of cropland BA patches (Hall et al., 2016; Zhu
et al., 2017). It is reasonable to assume that the same issue was affecting both the FireCCI51 and C3SBA10 products, which
520 clearly increased the uncertainty in those regions and, hence, decreased the agreement. In fact, it was very likely that the rest
of the biomes were affected by a similar issue (Vadrevu et al., 2019), showing a decrease in the agreement of products from
March to May (Figure 8) when agricultural activity is high globally (Randerson et al., 2012).

The impact of agricultural fire activity in the performance of moderate BA products is only an example of the importance of
the human factor in global fire regimes (Andela et al., 2017). Benali et al. (2017) assessed the existence of bimodal fire
525 seasons, i.e. areas with two separate peak months, that were directly related to human activities. They found that about 25%
of the areas with relevant fire activity showed this behaviour. In those cases, at least one of the two fire seasons was found to
take place in sub-optimal weather conditions. Commonly, those periods have less valid images available due to clouds and
cloud shadows, hindering BA detection, increasing uncertainty, and reducing agreement between products.

The sensor characteristics are also affecting the correlation between products as well. Different sensors imply different
530 observation geometries, different angular effects and different times of overpass, directly influencing what is observed by
each sensor. This led to the detection of different burned patches in one product compared to the other, e.g. in cropland areas,
or to the detection of different regions of the same burned patch as in some xeric shrublands, where fires can be large, but also
irregular and with high spread rates (Laurent et al., 2018). Conversely, the temporal reporting accuracy (Section 3.2.2) was
not significantly affected by this issue since a similar accuracy was shown by both products through the studied three-year
535 period (Figure 7). This was due to the use of monthly composites, which tend to select the most suitable observation from
the images available and did not require statistics of several pre- and post-fire images as in the case of multitemporal
approaches.

The results presented in this paper represent a first step in the transition from MODIS-based to S3-based BA products. In
that sense, OLCI sensor have shown a good capacity to detect BA globally, although these estimations still depend on
540 MODIS active fires. Future work should be done to assess the impact of changing the source of the active fires to other
sensors such as the Visible Infrared Imaging Radiometer Suite (VIIRS) on board Suomi-NPP and NOAA-20 or the Sea and
Land Surface Temperature Radiometer (SLSTR) on board S3. In the former case, VIIRS active fires can considerably
increase the detection of small fires due to its improved spatial resolution of 375m (Oliva and Schroeder, 2015; Schroeder et
al., 2014), although the impact on the identification of burned patches should be studied, due to the limitations of the
545 moderate resolution reflectance data to detect small burned areas. In the latter case, a first study of S3 SLSTR active fire
product has shown a good performance of this product over small fires (Xu et al., 2020), which provides a unique
opportunity to develop a hybrid BA algorithm entirely based on S3. Therefore, global BA estimations may be improved in



the future if the operational algorithm used to detect BA within the C3S is updated with new and/or improved versions provided through research projects, such as FireCCI.

550 **5 Data availability**

The C3SBA10 product presented in this paper includes two types of BA files: monthly full resolution continental tiles and monthly global grid files at an aggregated resolution of 0.25 degrees, as described in Section 2.4 and Table 2. Both datasets, pixel and grid data types, are freely available through the Copernicus Climate Data Store (CDS) repository (DOI: <https://doi.org/10.24381/cds.f333cf85>, Lizundia-Loiola et al. (2020a)), referenced as OLCI C3S Burned Area v1.0. Users
555 should note that FireCCI51 is also available through the same DOI, referenced as MODIS FireCCI v5.1.1cds. Additionally, the reference files generated to spatially validate the data presented in this publication has been released and are available at <https://doi.org/10.21950/BBQQU7> (Franquesa et al., 2020a).

6 Conclusions

This paper presents a new global BA product, called C3SBA10, that was developed and produced within the Copernicus
560 Climate Change Service. The last version of a series of global BA algorithms developed by the ESA's FireCCI project, known as FireCCI51 (Lizundia-Loiola et al., 2020b), was adapted to detect burned area for this Service. The algorithm follows a two-phase hybrid approach, which is based on S3 OLCI surface reflectance and MODIS active fires. Initially, the algorithm detects pixels with a high probability of being burned based on active fire information and NIR reflectance. Then, thresholds adapted to each fire and fuel conditions are computed using spatio-temporal active-fire clusters and their
565 surroundings. Finally, a contextual growing is applied to fully detect the burned patch. Thus, the first step aims to reduce commission errors while the second reduces omission ones. This algorithm is being operationally applied to produce the C3SBA10 product, which is the official BA product of the C3S service. At the time of writing this paper more than three years (January 2017 – present) of global BA are publicly available at the Climate Data Store in pixel (300 m) and grid (0.25°) format. The C3SBA10 may be understood as a continuation of the FireCCI51 global BA product (Lizundia-Loiola et
570 al., 2020b), which is available from 2001 to 2019. The spatial and temporal consistency assessment and the temporal reporting accuracy analysis showed a very high correlation between both products, mainly in 2019 when both S3A and B were available, ensuring an uninterrupted provision of global BA information from 2001 to present.



Appendix A: Supplementary tables

7 Author contributions

575 JL, MLP and EC have written the first draft of the manuscript. JL has coordinated the manuscript production and developed the algorithm, prepared the figures, and made the intercomparison analysis. MF generated the reference data and validated the products. GK and MB were responsible of the pre-processing of the OLCI data and the production of the global burned area product. All the authors have contributed to the writing and reviewing of the manuscript and agreed on the final version.

8 Competing interests

580 The authors declare that they have no conflict of interest.

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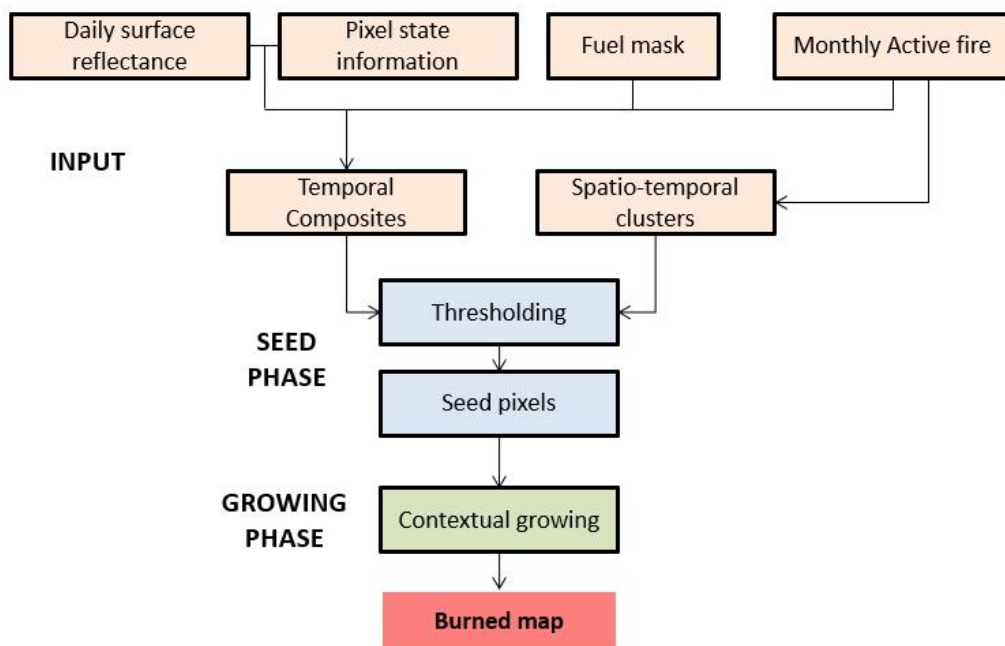
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Figures



830 **Figure 1.** Simplified version of the main scheme of the global BA algorithm. Adapted from Lizundia-Loiola et al. (2020b).

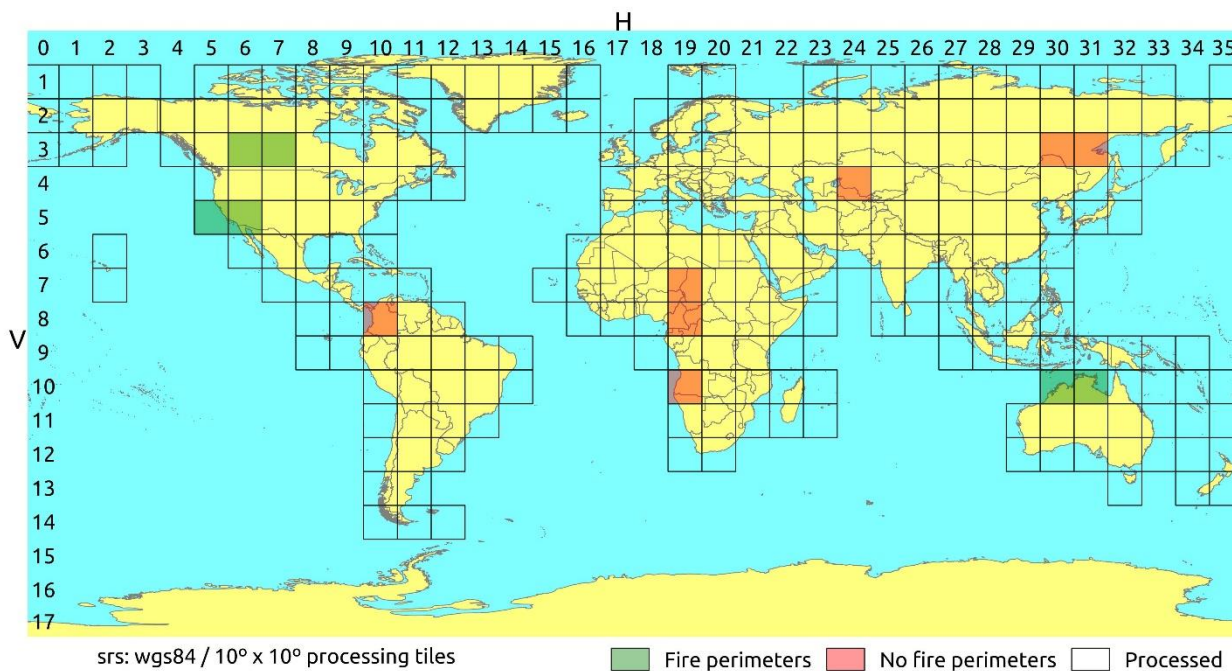
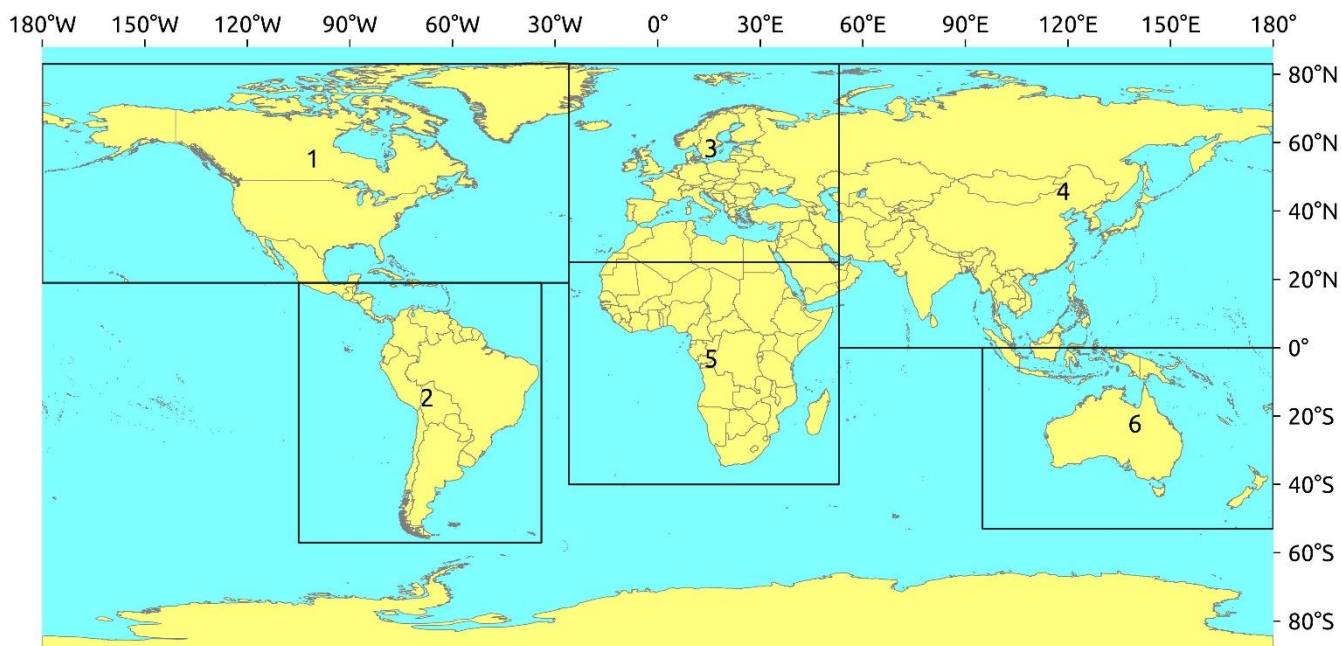


Figure 2. The 10x10 degree tiles used for the processing of the operational OLCI global BA product. A total of 273 tiles were processed, of which 13 were used to adapt the algorithm (red and green tiles).



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Figure 3. Continental tiles of the global BA pixel product.

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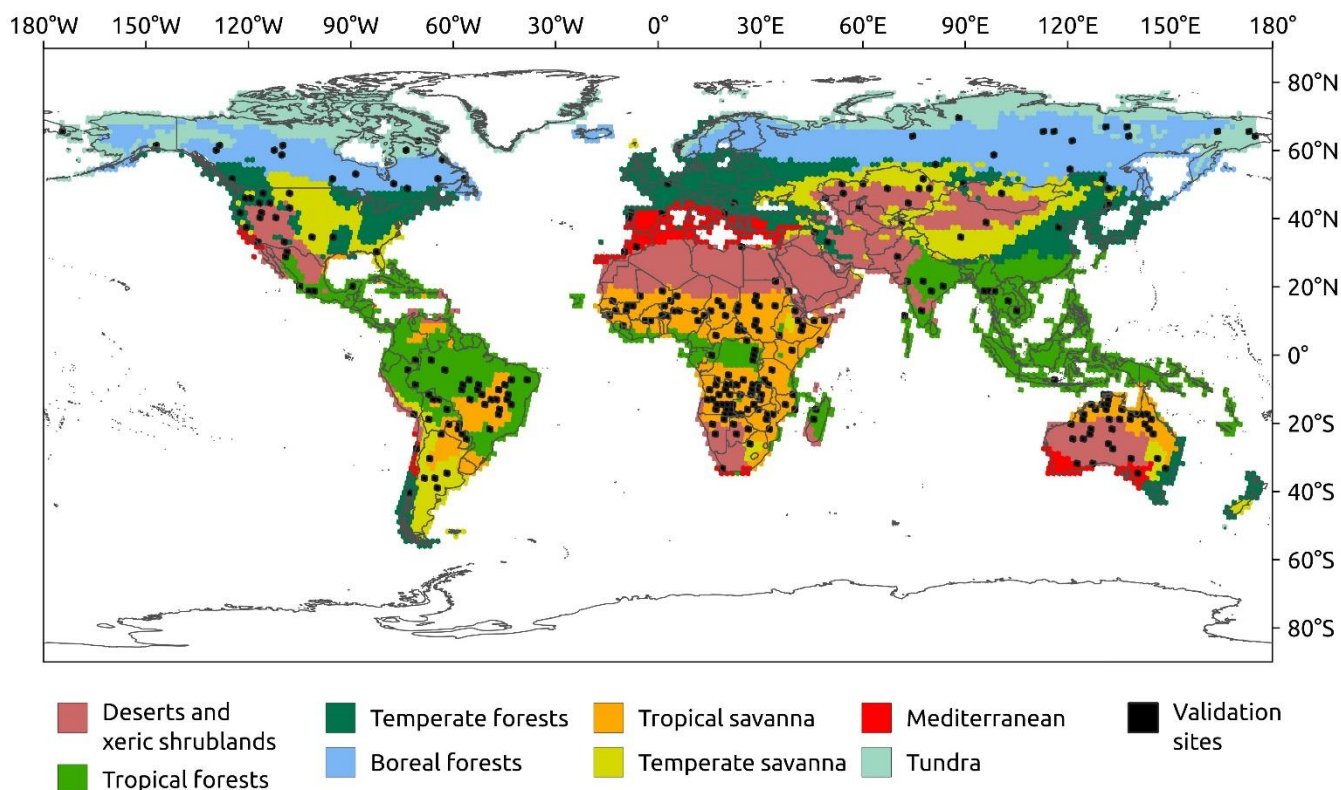


Figure 4. Biomes representing the first stratum of the validation sampling. The black locations show the 300 validation sites that were selected for 2017, 2018, and 2019 (100 for each year).

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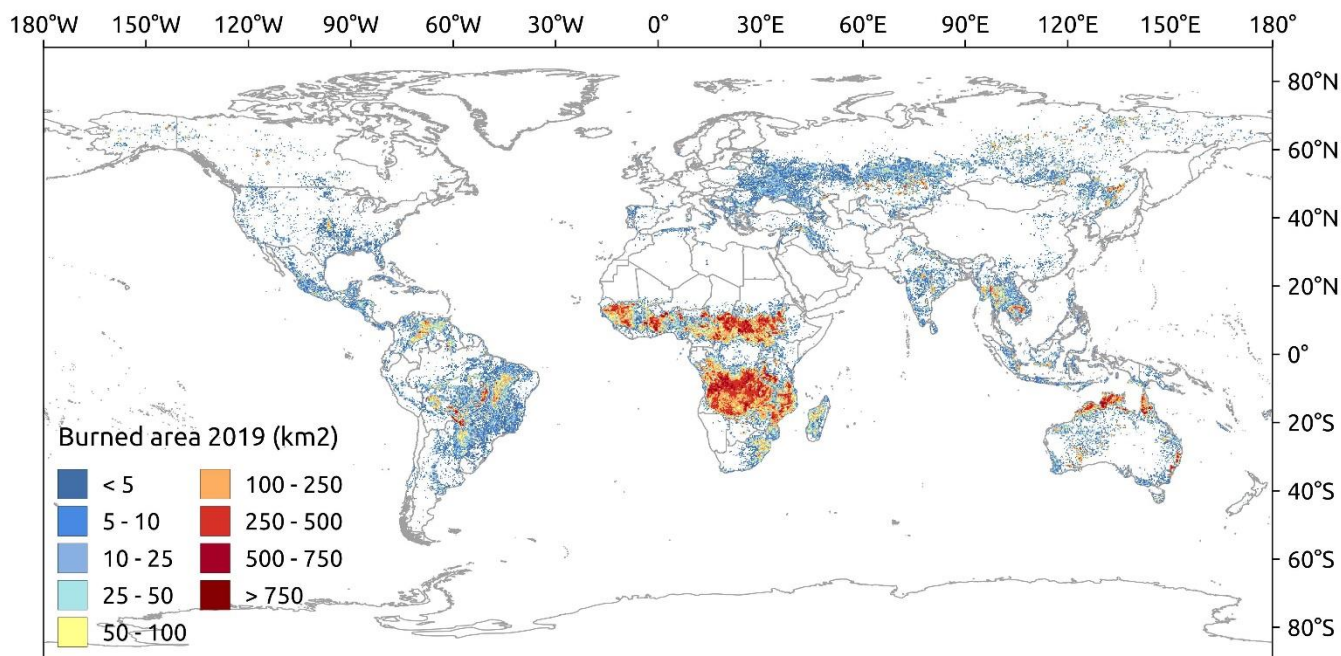


Figure 5. C3SBA10 annual accumulated BA for the year 2019 at 0.25° spatial resolution.

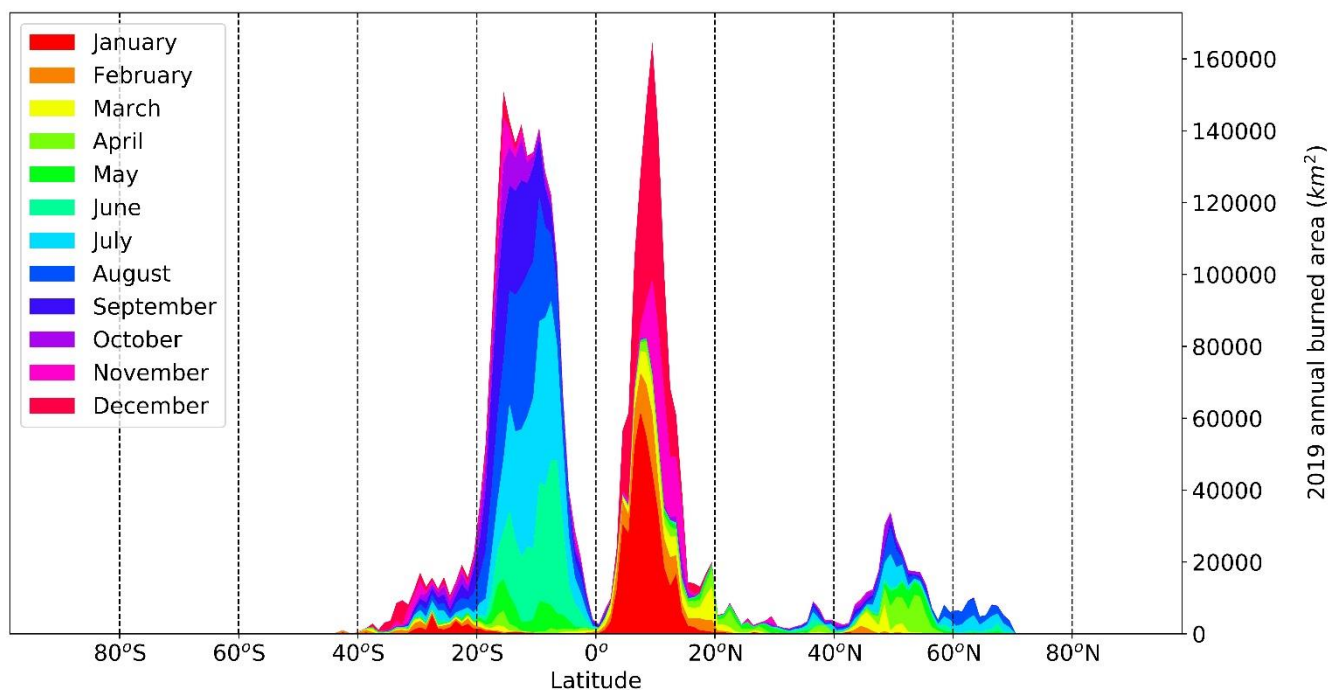


Figure 6. Monthly burned area for 2019 distributed along latitudes determined by C3SBA10.

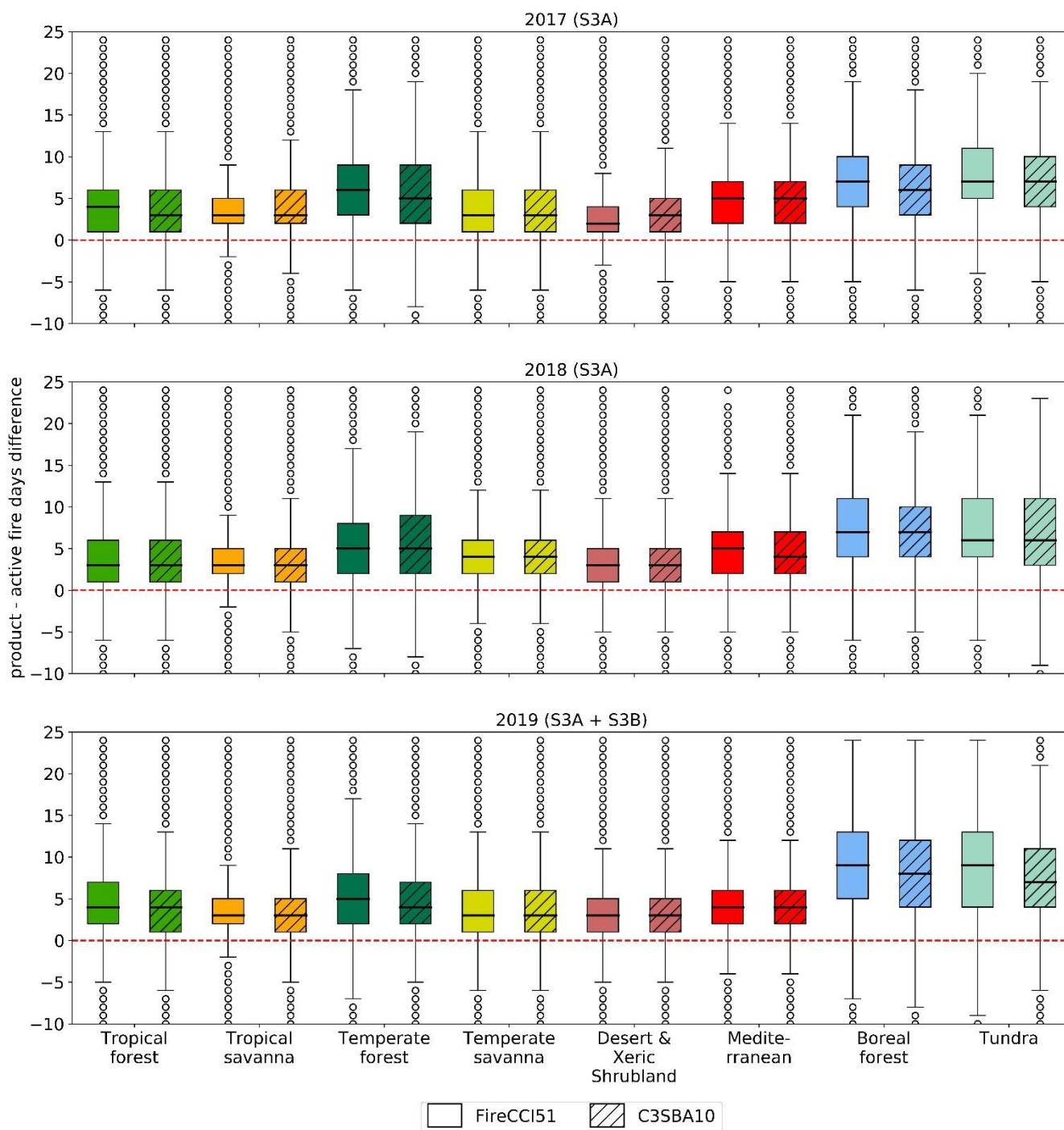


Figure 7. Temporal reporting accuracy assessment of both FireCCI51 and C3SBA10 products.

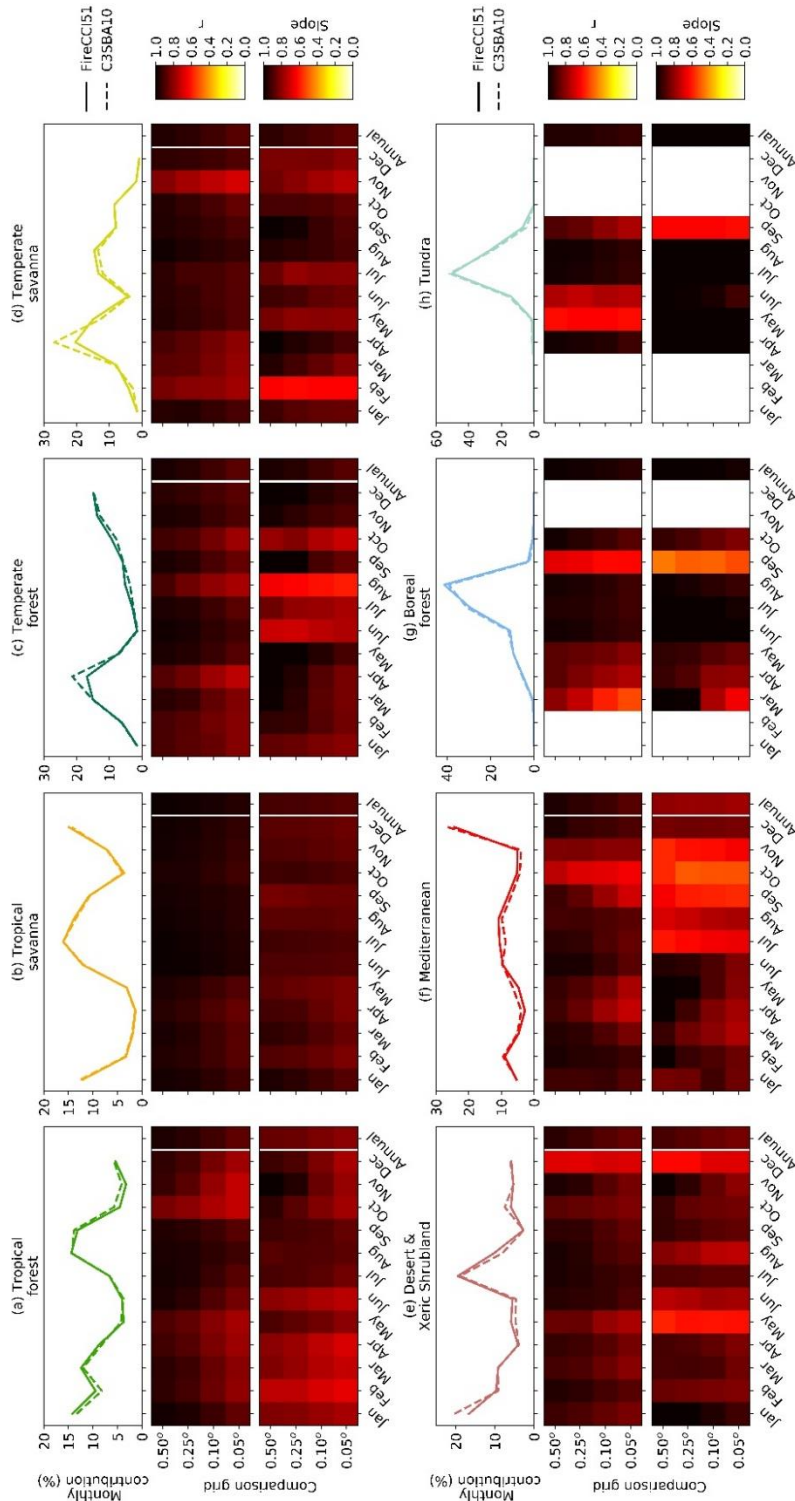
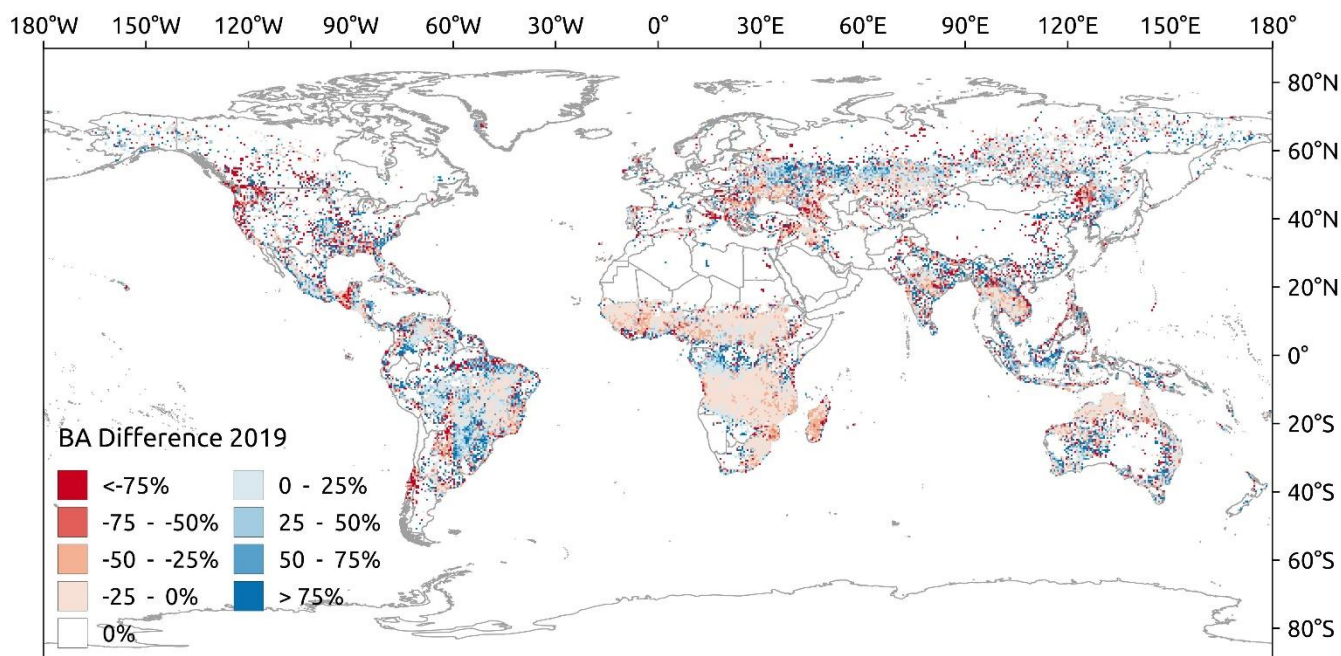


Figure 8. Monthly correlation between the FireCCI51 and C3SBA10 products per biome and grid size for the year 2019.



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Figure 9. Proportional differences of burned area detection between the FireCCI51 and C3SBA10 at 0.5° spatial resolution. The negative values (red) means that the FireCCI51 detected more burned area while the positive ones (blue) that the C3SBA10 detected more.

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Tables

Table 1. OLCI bands characteristics. The band indicated in grey was the one used as input for the algorithm.

Band	Band centre (nm)	Bandwidth (nm)	MERIS heritage
Oa01	400	15	No
Oa02	412.5	10	Yes
Oa03	442.5	10	Yes
Oa04	490	10	Yes
Oa05	510	10	Yes
Oa06	560	10	Yes
Oa07	620	10	Yes
Oa08	665	10	Yes
Oa09	673.75	7.5	No
Oa10	681.25	7.5	Yes
Oa11	708.75	10	Yes
Oa12	753.75	7.5	Yes
Oa13	761.25	2.5	Yes
Oa14	764.375	3.75	No
Oa15	767.5	2.5	No
Oa16	778.75	15	Yes
Oa17	865	20	Yes
Oa18	885	10	Yes
Oa19	900	10	Yes
Oa20	940	20	No
Oa21	1020	40	No

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Table 2. Pixel and Grid products specifications.

Product	Spatial resolution	Layers	Description
Pixel	300m	JD	Julian date or day of the year when the burned pixel was detected
		CL	Confidence level of the classified pixel (both burned and unburned)
		LC	Land cover class that was burned
Grid	0.25°	Burned area	Sum of the burned area within the grid cell
		Standard error	Estimation of the standard error of the burned area
		Fraction of burnable area	Fraction of the grid cell that could burn (vegetated land covers)
		Fraction of observed area	Fraction of the burnable area that was observed during the month
		Burned area of each land cover class	Sum of the burned area within the grid cell per land cover class

915 **Table 3. Burned area per biome derived from the C3SBA10 product, and relative differences against the predecessor global BA product FireCCI51.**

Biome	C3SBA10 Burned area (km ²)			Difference C3SBA10-FireCCI51		
	2017*	2018*	2019**	2017*	2018*	2019**
Tropical savanna	2782564	2801290	2701210	-17.7%	-16.4%	-9.5%
Tropical forest	335032	313597	411926	-22.3%	-17.6%	-5.2%
Temperate savanna	210382	135370	164995	-9.8%	-3.9%	-0.4%
Desert and xeric shrubland	258622	148073	113707	-20.7%	-19.4%	-2.8%
Temperate forest	89537	111662	110884	-8.9%	-11.5%	-1.0%
Boreal forest	61173	74014	90503	+4.5%	+2.4%	+4.4%
Mediterranean	27421	8884	24975	-18.9%	-32.8%	-16.8%
Tundra	5734	1697	13044	+3.8%	+4.1%	+13.1%
Global	3770465	3594588	3631243	-17.3%	-15.7%	-7.8%

* Only S3A satellite available

** Both S3A and B satellites available



Table 4. Accuracy metrics for the years 2017, 2018 and 2019. Standard errors are shown between parentheses.

Year	2017*		2018*		2019**	
	FireCCI51	C3SBA10	FireCCI51	C3SBA10	FireCCI51	C3SBA10
Dice coefficient	0.64(0.03)	0.59(0.03)	0.68(0.03)	0.64(0.03)	0.61(0.03)	0.59(0.04)
Commission error	0.23(0.03)	0.23(0.03)	0.17(0.02)	0.14(0.02)	0.22(0.02)	0.20(0.02)
Omission error	0.44(0.03)	0.52(0.04)	0.42(0.04)	0.50(0.04)	0.49(0.04)	0.53(0.04)
Relative bias	-0.27(0.05)	-0.38(0.05)	-0.31(0.04)	-0.41(0.04)	-0.35(0.04)	-0.42(0.04)

920 * Only S3A satellite available ** Both S3A and B satellites available

Table 5. Global correlation coefficient between FireCCI51 and C3SBA10, slope and RMSE for the years 2017, 2018 and 2019 at 0.05°, 0.10°, 0.25° and 0.50° grid size.

		Grid Size			
		0.05°	0.1°	0.25°	0.5°
r	2017*	0.937	0.962	0.981	0.988
	2018*	0.943	0.966	0.983	0.989
	2019**	0.952	0.972	0.986	0.992
slope	2017*	0.823	0.836	0.844	0.847
	2018*	0.838	0.850	0.857	0.859
	2019**	0.882	0.894	0.904	0.907
RMSE (km ²)	2017*	1.347	4.029	17.810	57.404
	2018*	1.279	3.807	16.800	54.338
	2019**	1.073	3.060	12.577	38.092

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Table A1. Average (2017-2019) temporal reporting accuracies of the C3SBA10.

Reporting accuracy	Tropical forest	Tropical savanna	Temperate forest	Temperate savanna	Desert & xeric shrubland	Mediterranean	Boreal forest	Tundra	Global
0-1 days	18.4%	20.1%	14.3%	22.6%	26.8%	14.8%	7.0%	8.4%	18.8%
0-3 days	42.3%	49.5%	33.0%	49.4%	57.2%	37.9%	18.5%	20.4%	45.4%
0-5 days	61.4%	69.5%	51.1%	68.2%	75.4%	59.5%	33.0%	34.7%	64.6%
0-10 days	85.5%	90.8%	80.3%	91.9%	94.1%	91.0%	70.0%	69.3%	88.1%